

Mechanisms maintaining high species diversity in tropical rainforests



Photo G. Weiblen



Quiz question: where is the global maximum of vascular plant diversity?

Global diversity maxima for vascular plants

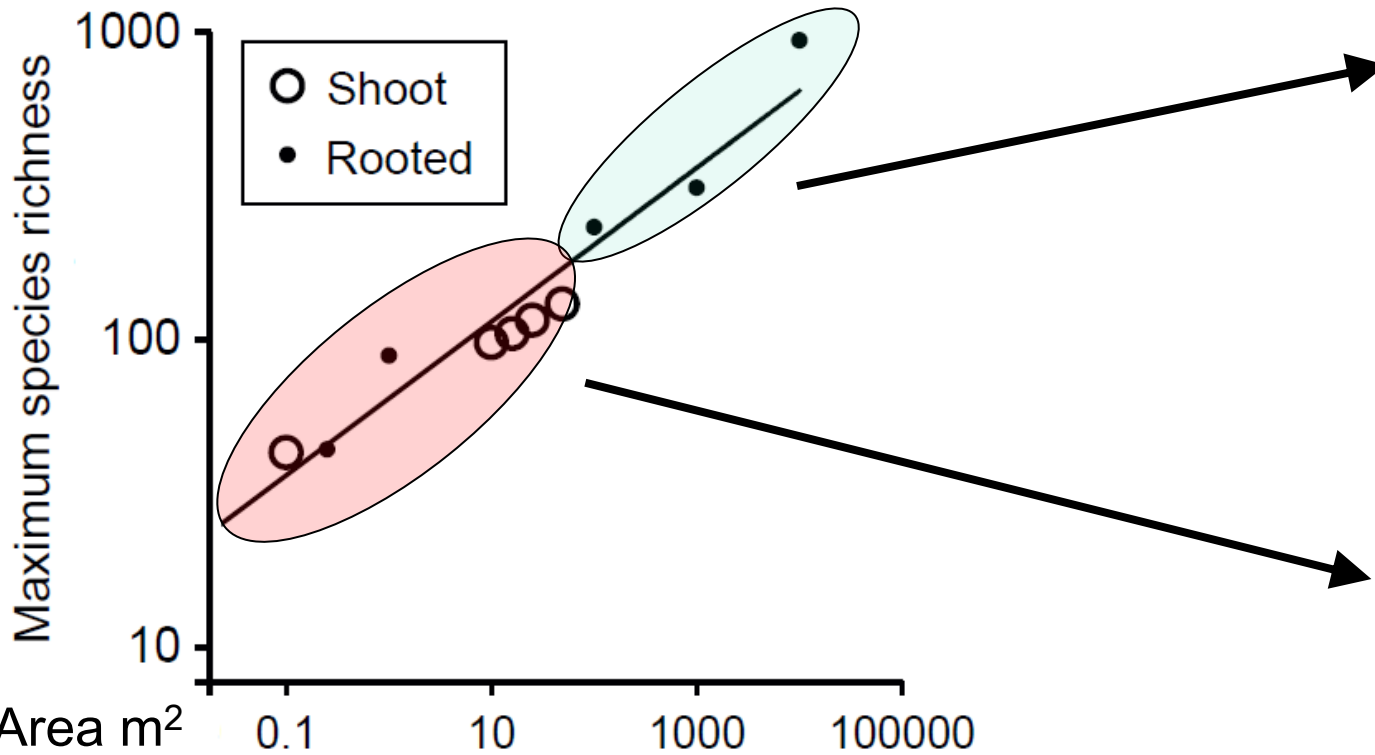


Table 1. The communities used as the richest in vascular plant species at a range of spatial grains.

| Area (m ²) | Richness | Method | Community | Region | References |
|------------------------|----------|--------|--------------------------------|----------------|--------------------------|
| 0.25 | 44 | Rooted | Semi-dry basiphilous grassland | Czech Republic | Klimeš et al. (2001) |
| 1 | 89 | Rooted | Mountain grassland | Argentina | Cantero et al. (1999) |
| 10 | 98 | Shoot | Semi-dry basiphilous grassland | Romania | Dengler et al. (unpubl.; |
| 16 | 105 | Shoot | Semi-dry basiphilous grassland | Czech Republic | Z. Otýpková (unpubl.) |
| 25 | 116 | Shoot | Semi-dry basiphilous grassland | Czech Republic | Z. Otýpková (unpubl.) |
| 49 | 131 | Shoot | Semi-dry basiphilous grassland | Czech Republic | Z. Otýpková (unpubl.) |
| 100 | 233 | Rooted | Tropical lowland rain forest | Costa Rica | Whitmore et al. (1985) |
| 1000 | 313 | Rooted | Tropical lowland rain forest | Colombia | Duivenvoorden (1994) |
| 10 000 | 942 | Rooted | Tropical rain forest | Ecuador | Balslev et al. (1998) |



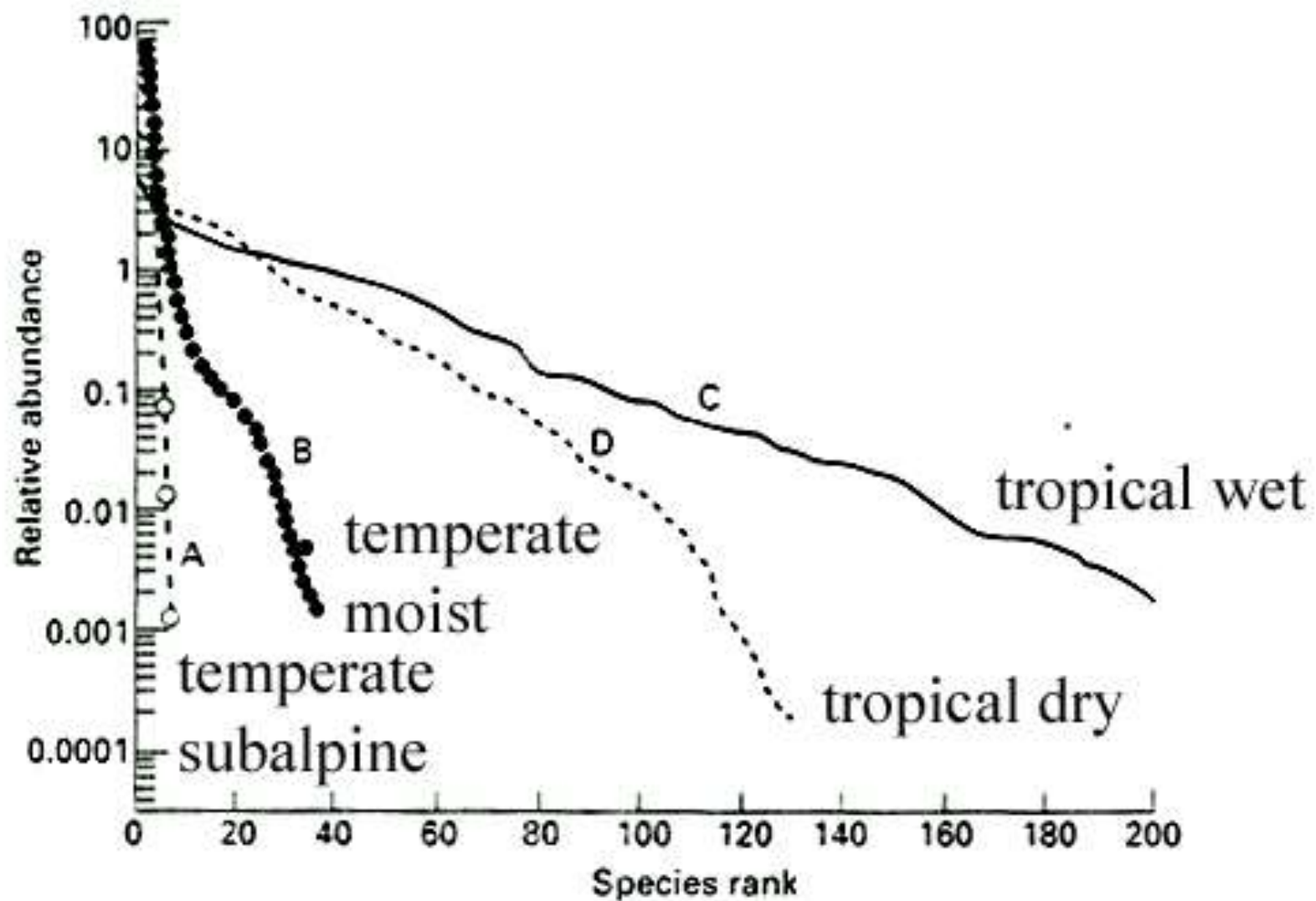
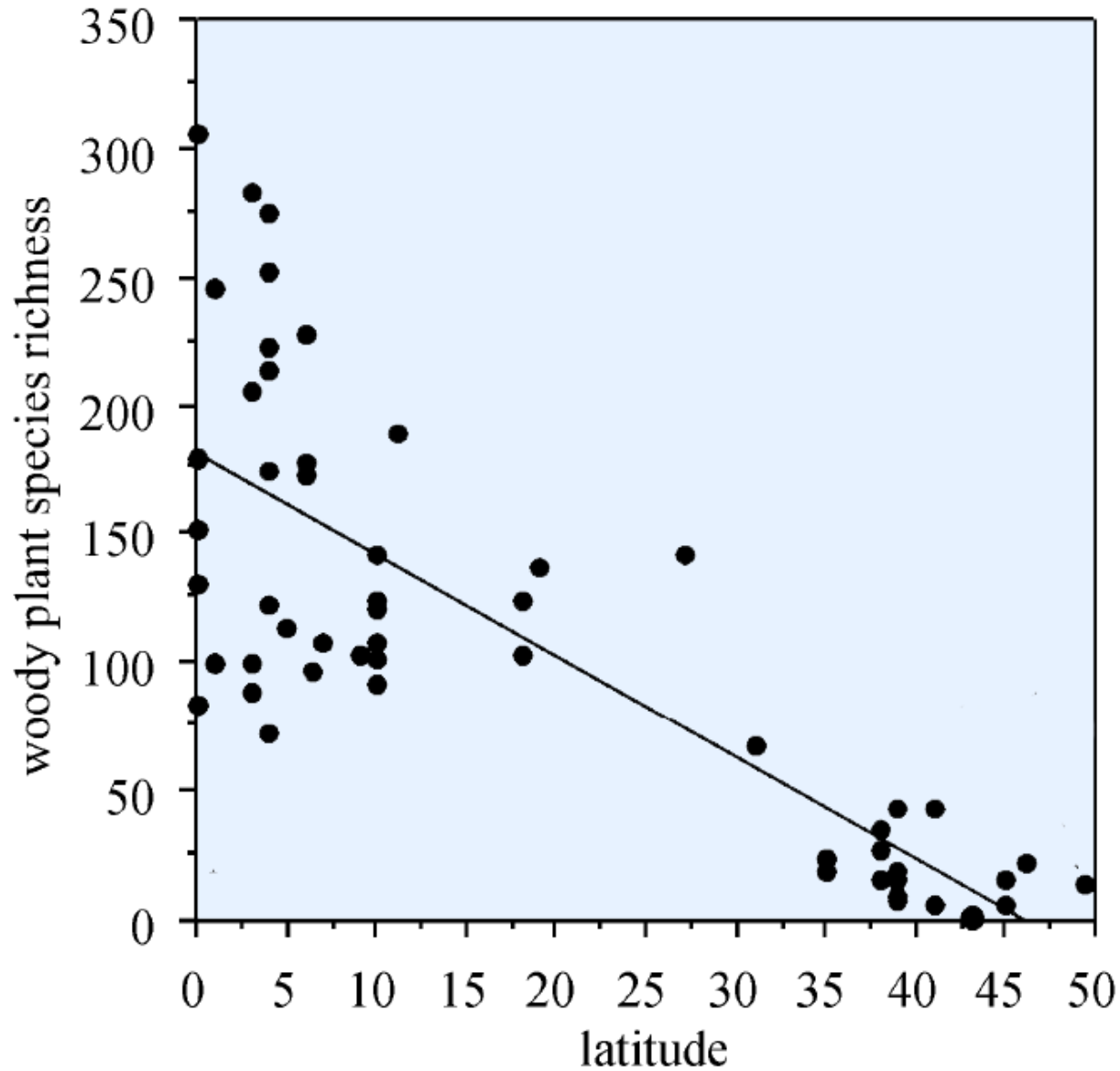


Fig. 7.16 Rank-abundance patterns of tropical and temperate forests: A, temperate montane (subalpine) forest (Smoky Mountains, Tennessee); B, temperate moist forest (Smoky Mountains, Tennessee); C, tropical wet forest (Manaus, Brazil); D, tropical dry forest (Guanacaste, Costa Rica). (After Hubbell 1979.)

Latitudinal gradient in species diversity

woody plants with DBH>10cm in 1 ha forest plots



1 ha of lowland rainforest = 150-300 tree species with DBH>10cm

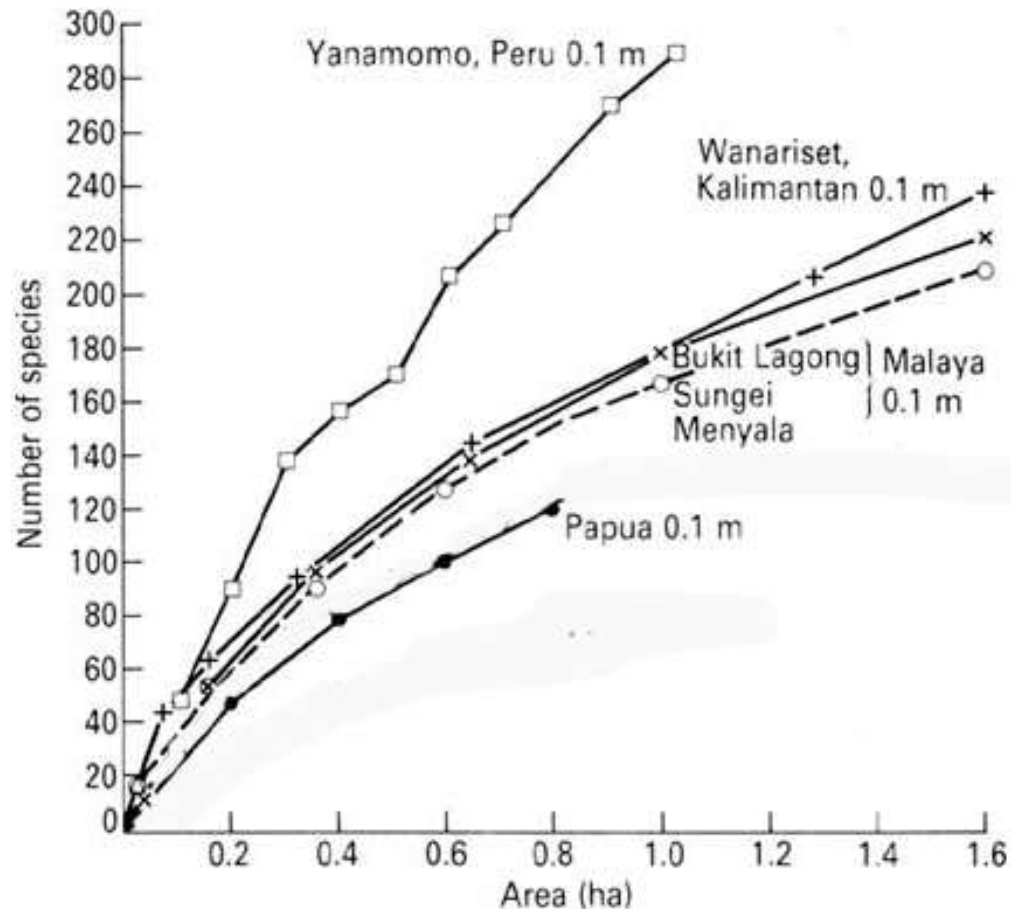
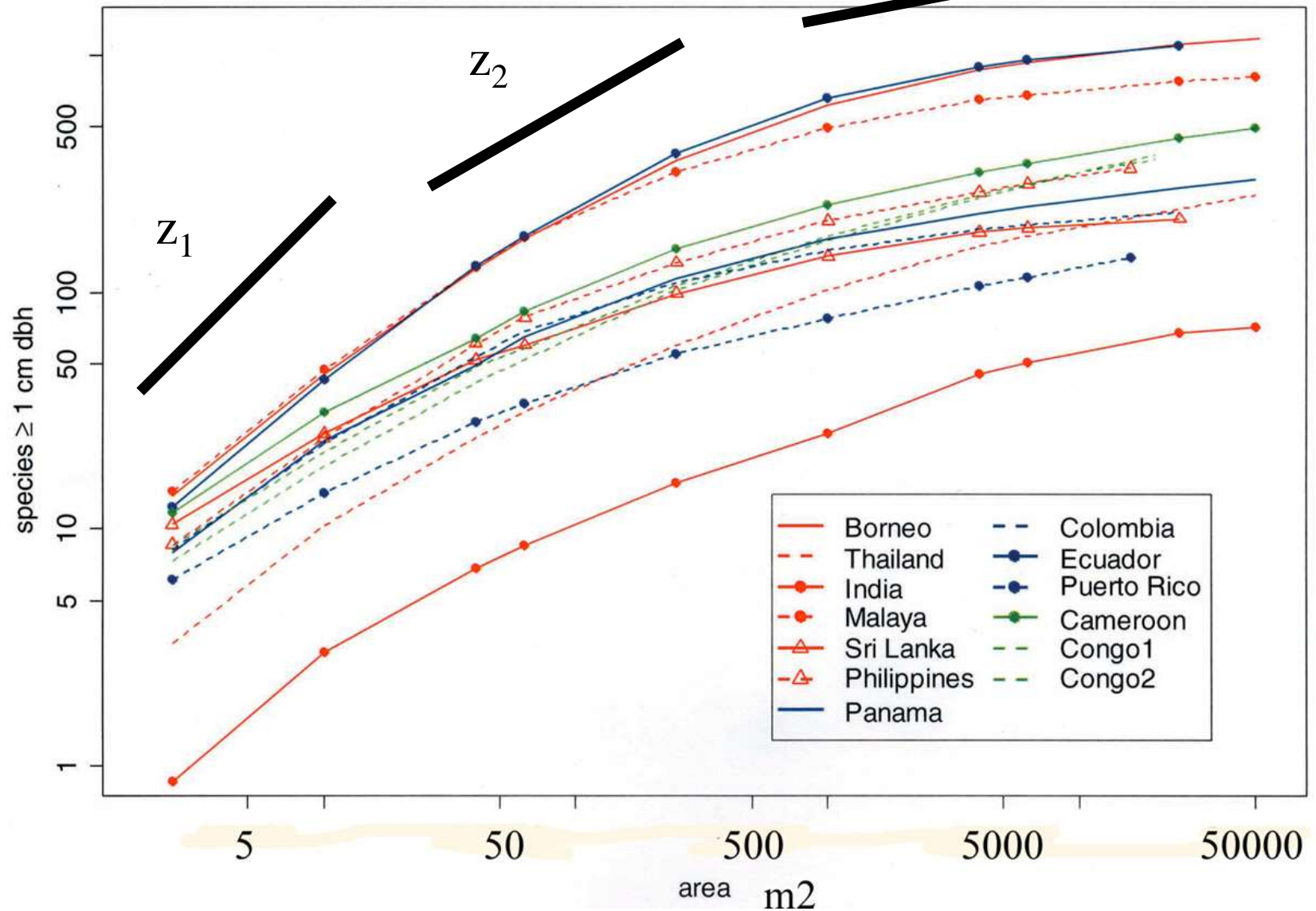
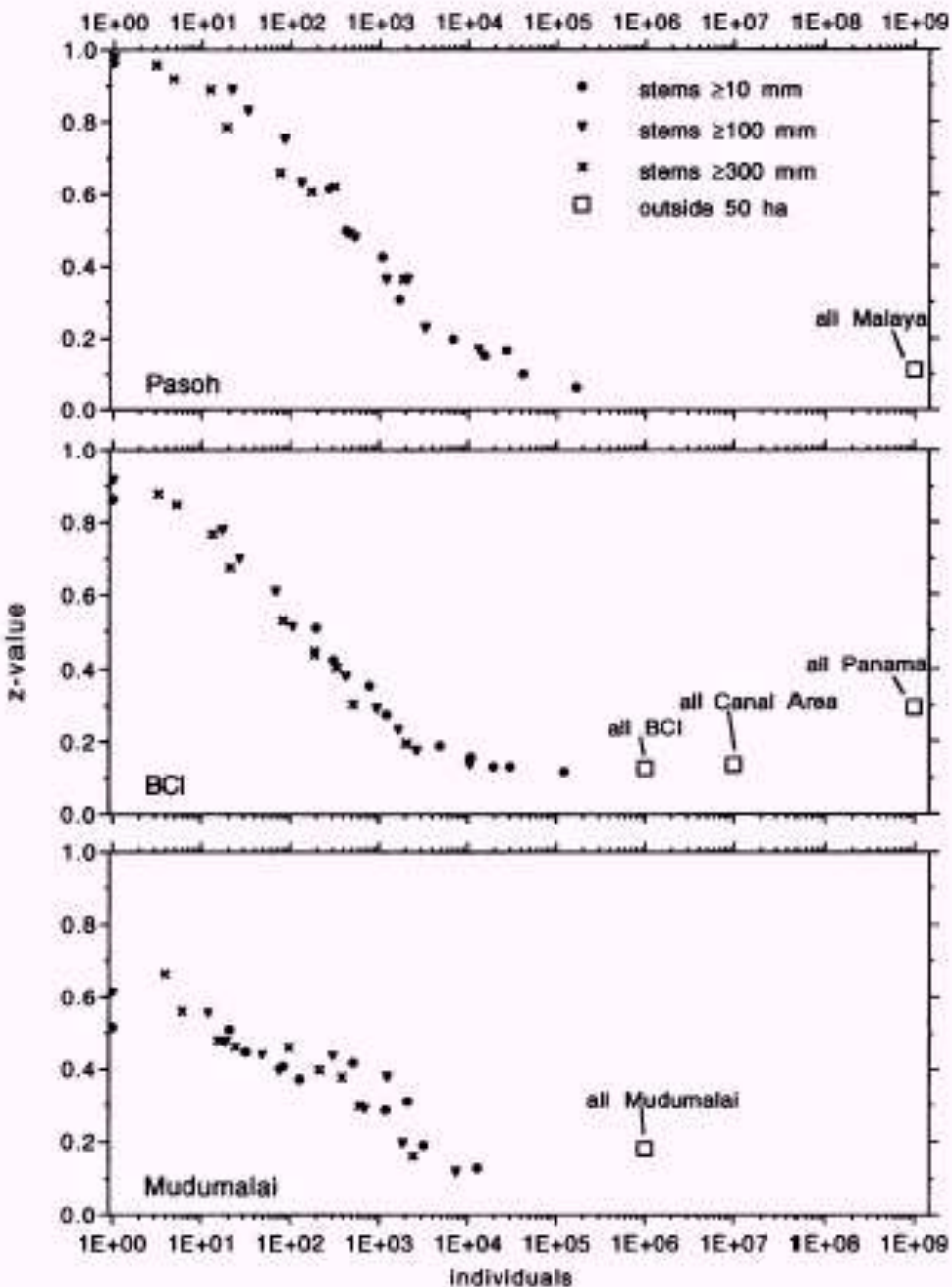


Fig. 2.28. Species area curves for tropical lowland evergreen rain forests. (After Whitmore 1984a and Genry 1988b).

$$S = c A^z \quad \log S = \log c + z * \log A$$

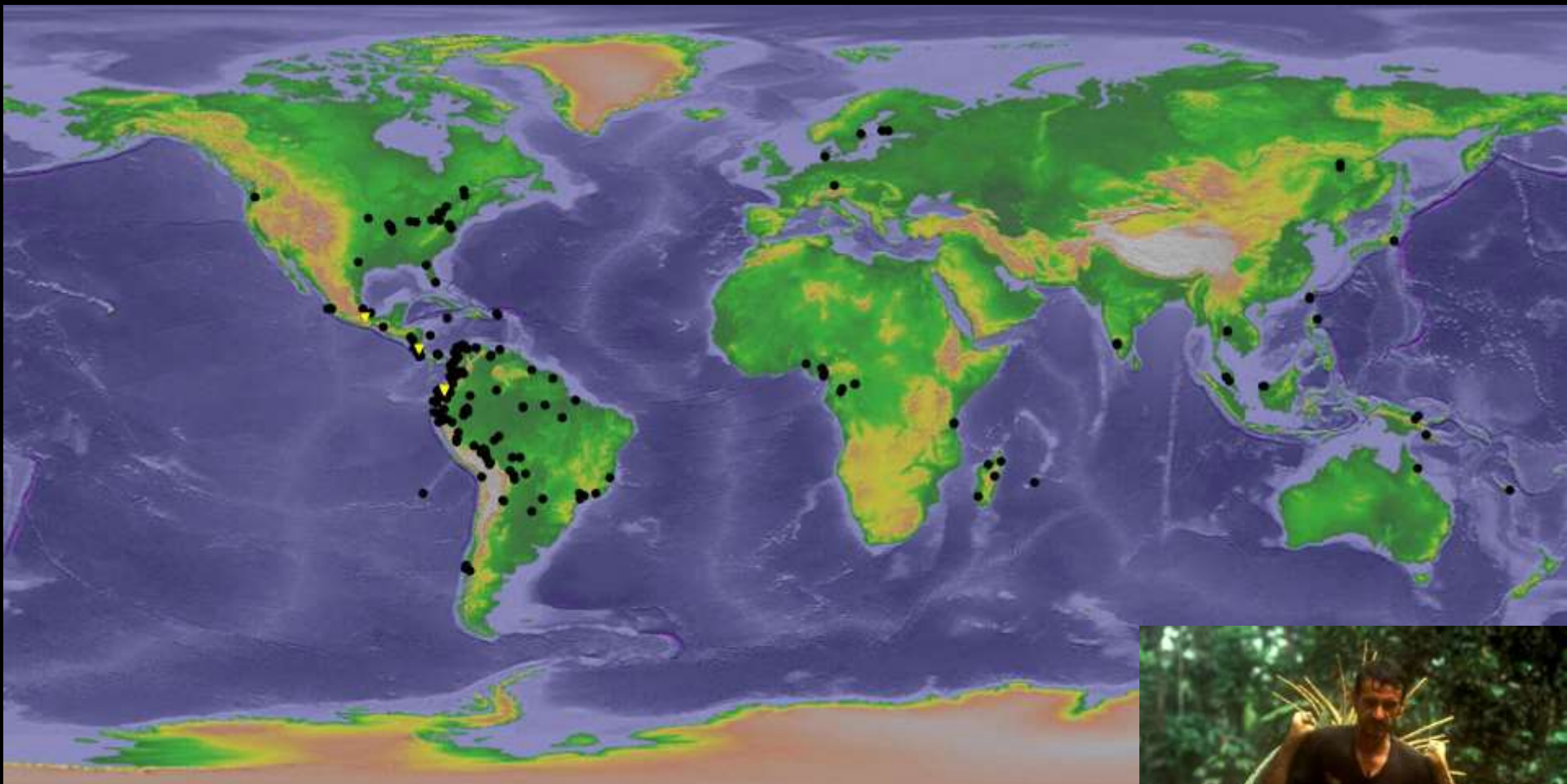
 z_3




Slope of log-log species accumulation curves depends on spatial scale

Fig. 4 Slope of log-log species-individual curves (z-values) for three 50-ha plots and from 50-ha plots to larger regions.

Alwyn Gentry's plots: 2 x 50 m = 0.1 ha DBH>2.5cm, 226 plots

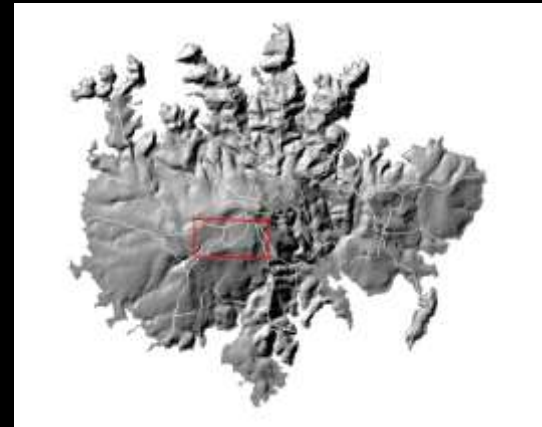


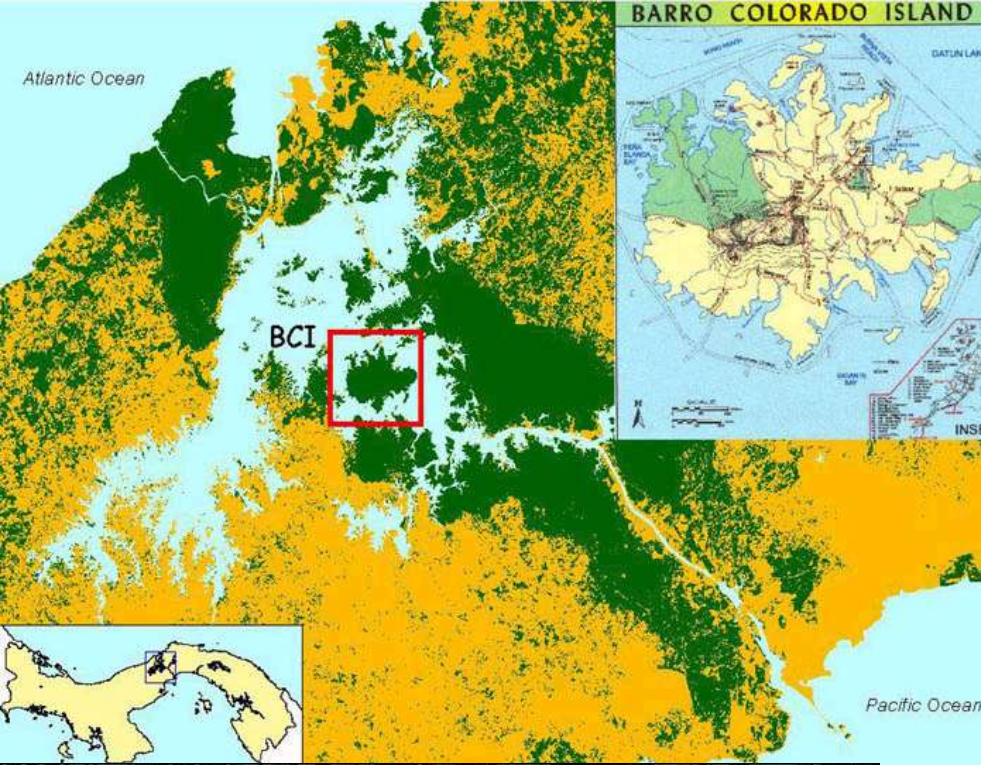
Start of the most successful ecological research in tropical botany:
in 1980, Stephen Hubbell and Robin Foster had a good inventory

1 ha stems with DBH>5cm
~1,000 stems



50 ha stems with DBH>1cm
~300,000 stems





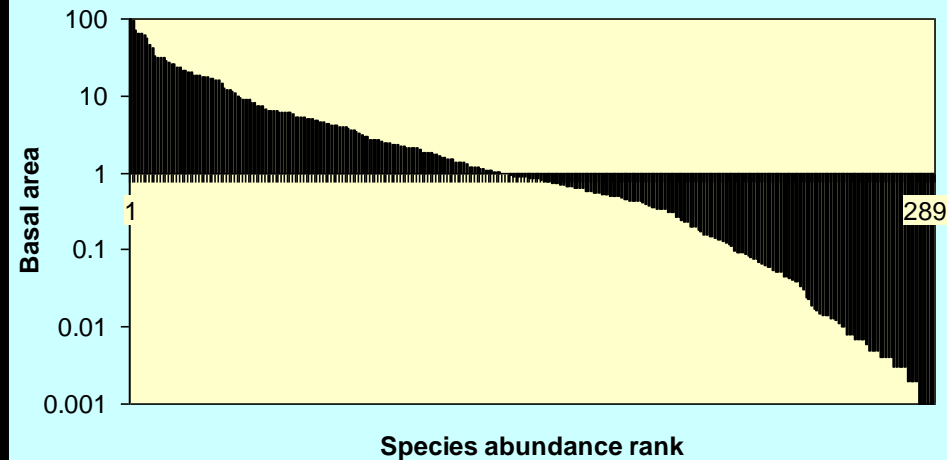
Mecca of tropical ecology: the Barro Colorado Island



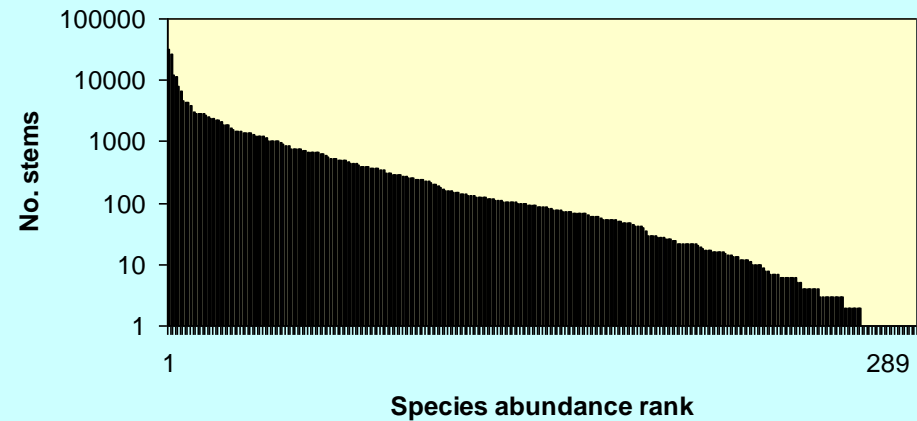
Barro Colorado Island: the first 50-ha plot



BCI 50 ha plot



BCI 50 ha plants DBH>1cm



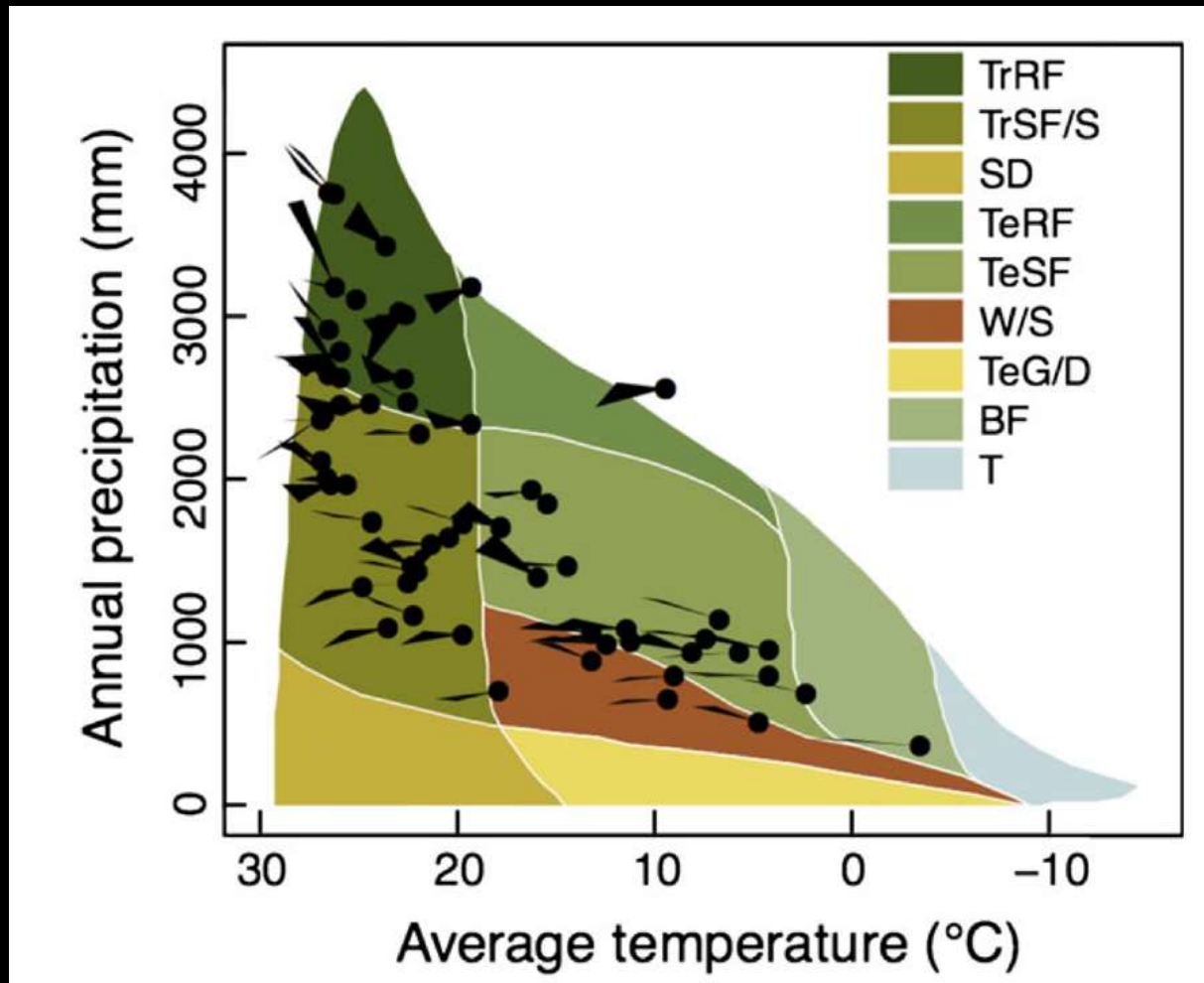
Forest Dynamics Plots

- Large scale (16-52 ha)
- All stems ≥ 1 cm mapped, measured, tagged, and identified to species
- Entire plot recensused every 5 years
- Standardized method used by all CTFS research sites

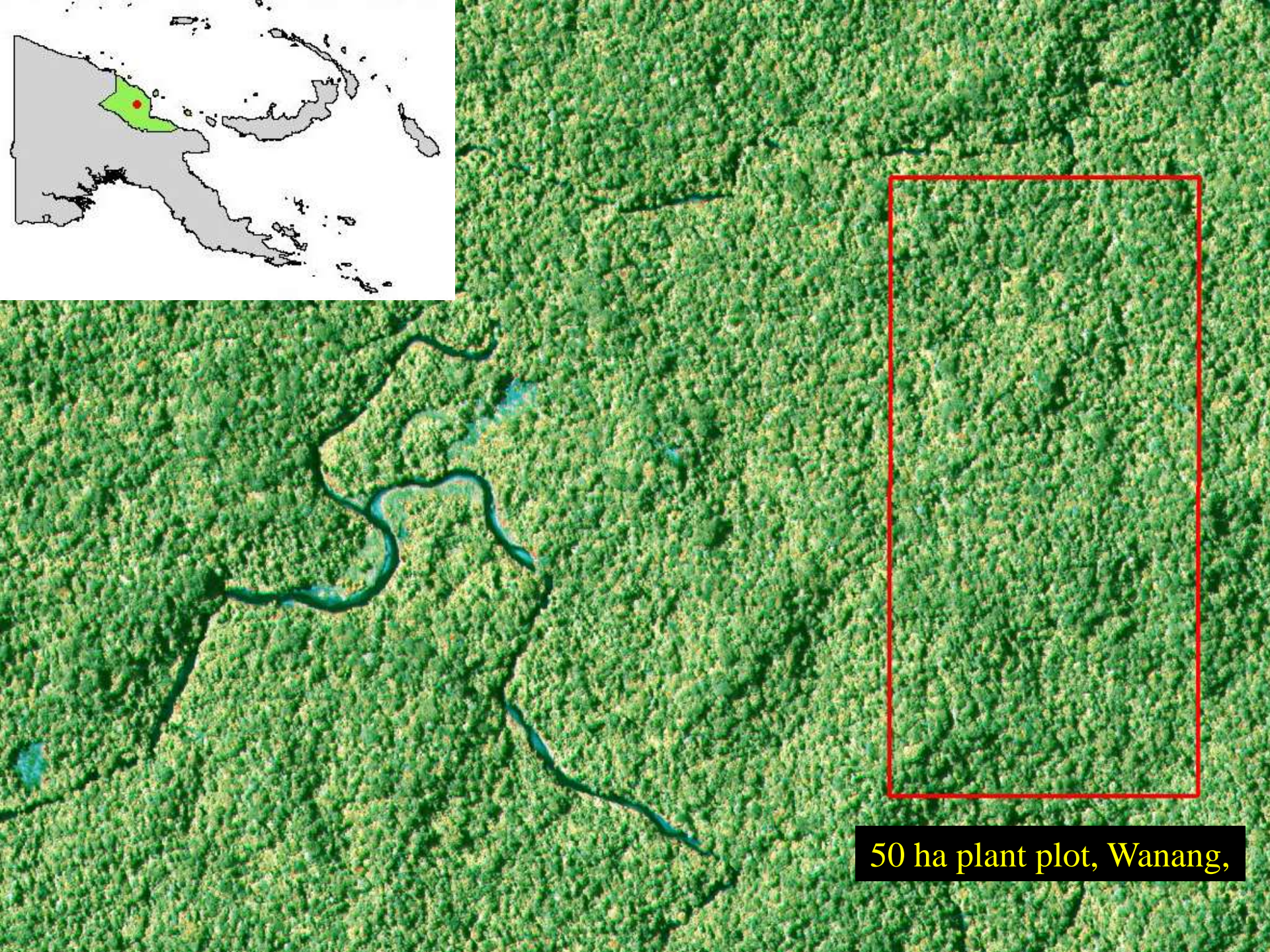
The Center for Tropical Forest Science network



Current and projected future (2050) mean annual temperature and precipitation of CTFS plots



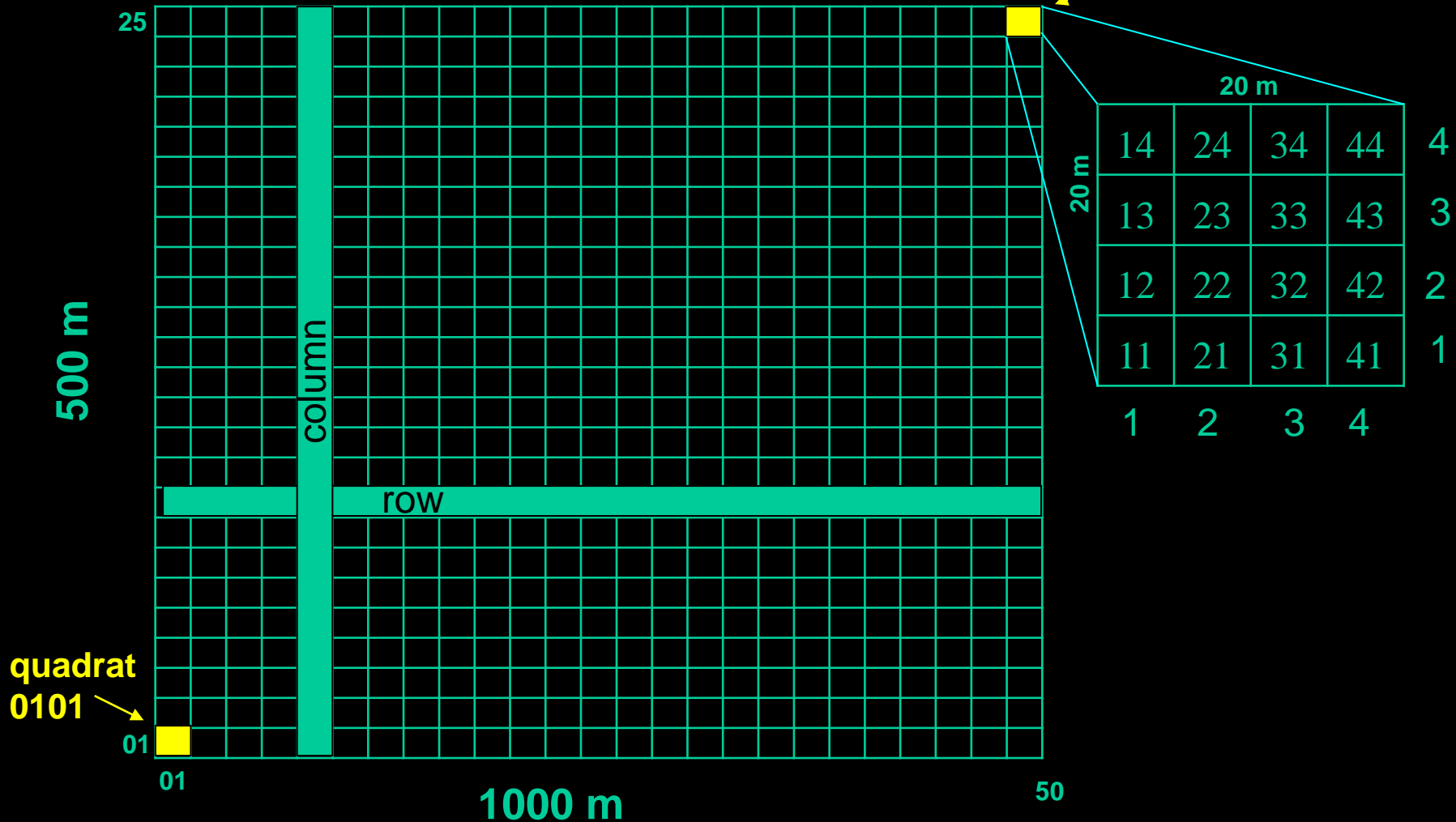
TrRF, tropical rain forest; TrSF/S, tropical seasonal forest/savanna; SD, subtropical desert; TeRF, temperate rain forest; TeSF, temperate seasonal forest; W/S, woodland/shrubland; TeG/D, temperate grassland/desert; BF, boreal forest; T, tundra

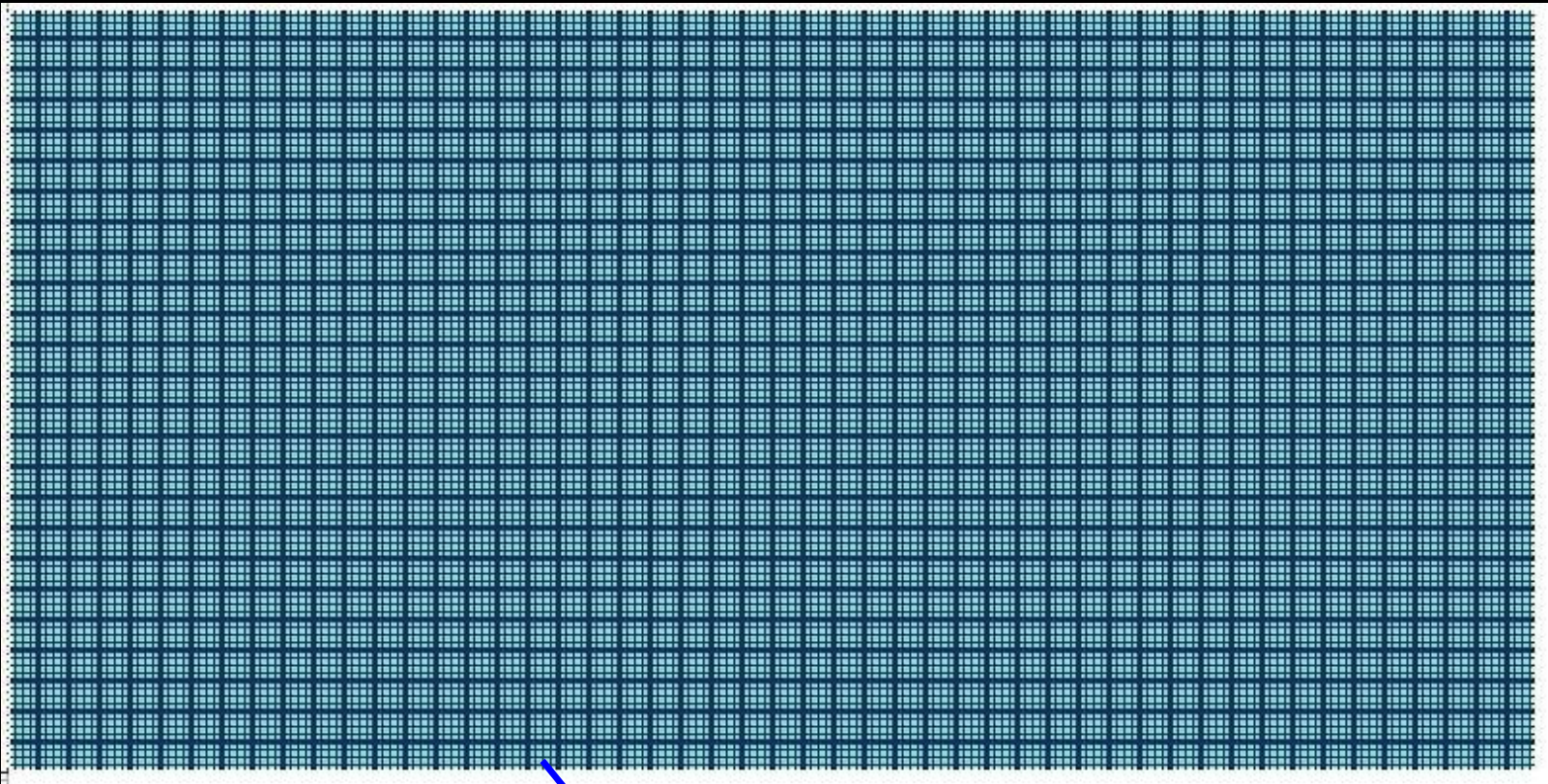


50 ha plant plot, Wanang,

Quadrats (20 m x 20 m) & Subquadrats (5 m x 5 m)

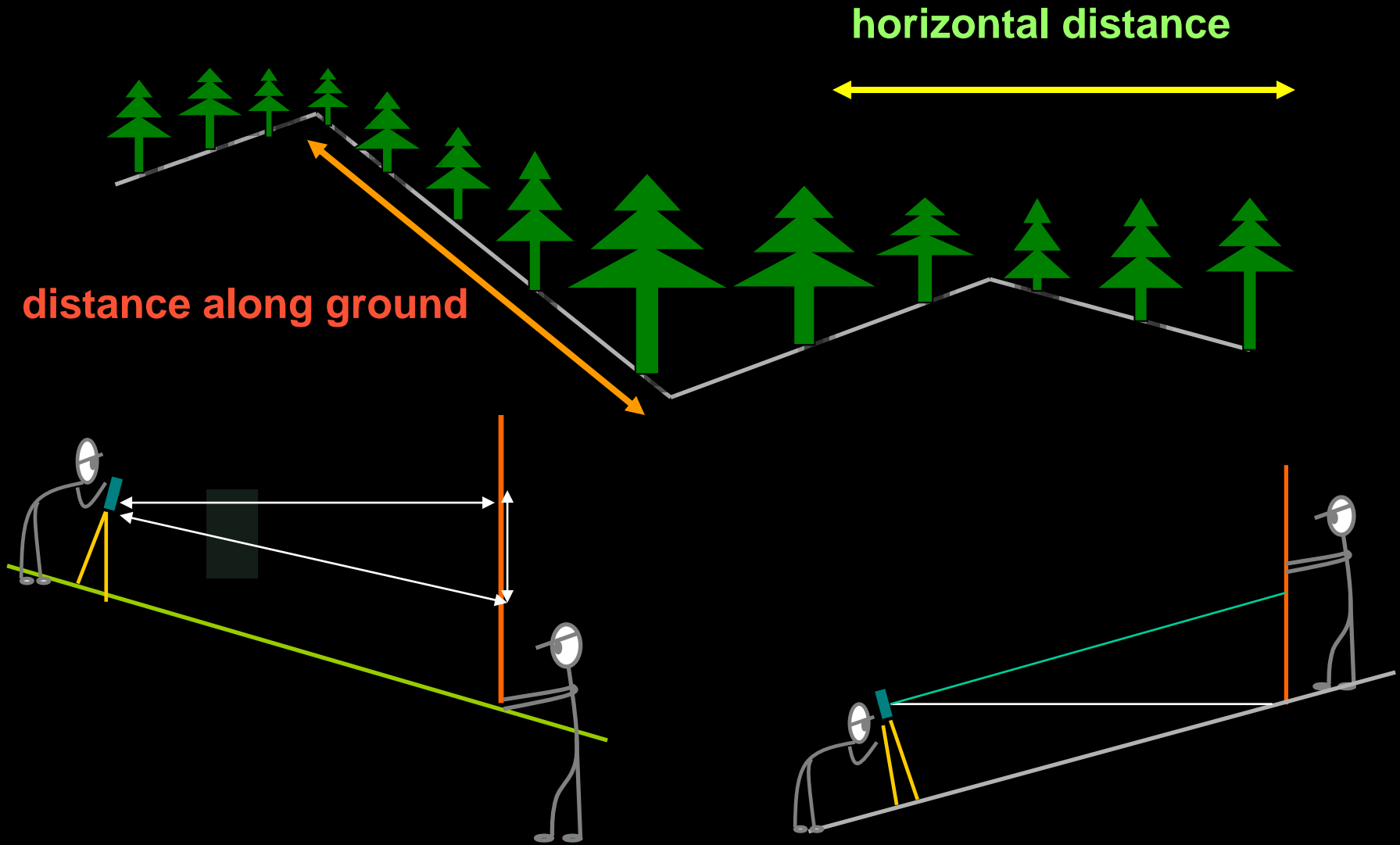
column & row coordinates from 0101 to 5025





1,250 quadrats &
20,000 sub-quadrats
in 50 hectares

20 x 20 survey principals







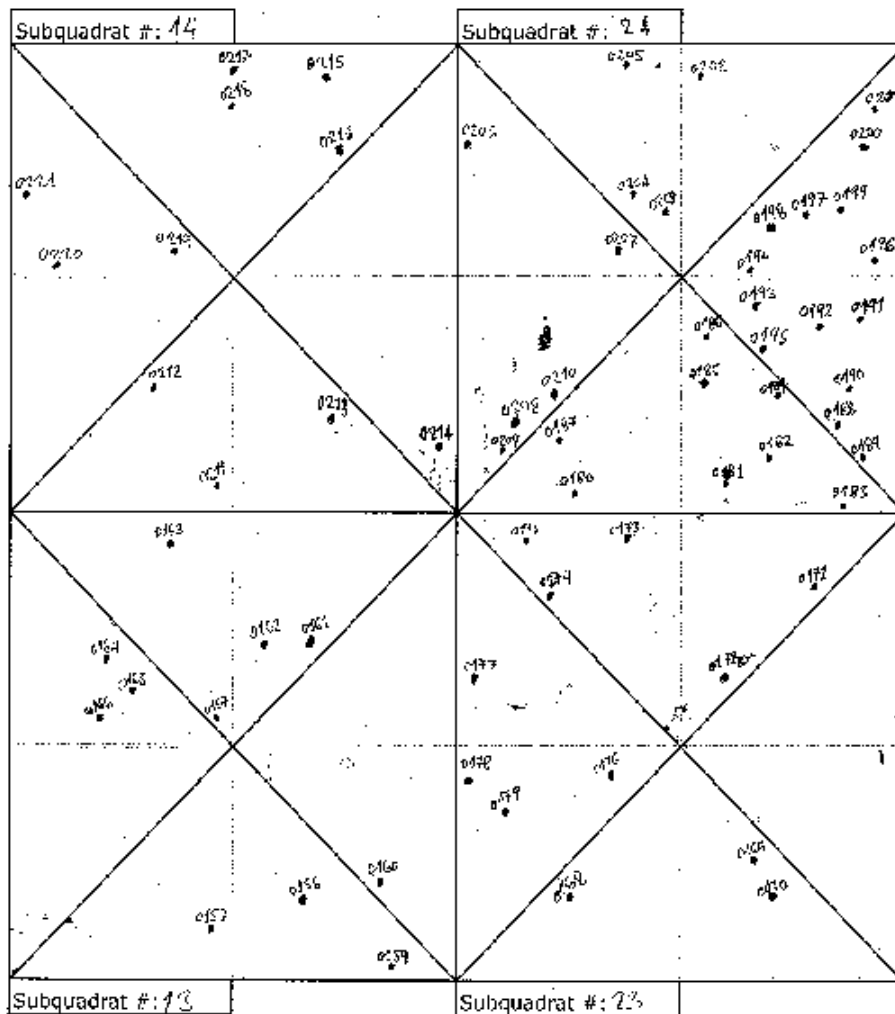


Quadrat Map

Name: RSLH-Lilip Quadrat: 0101

Date (day/mo/yr): 30/04/2009 Checked by: _____

Tag sequence: 0157 to 0224 Sheet: 4 of 4





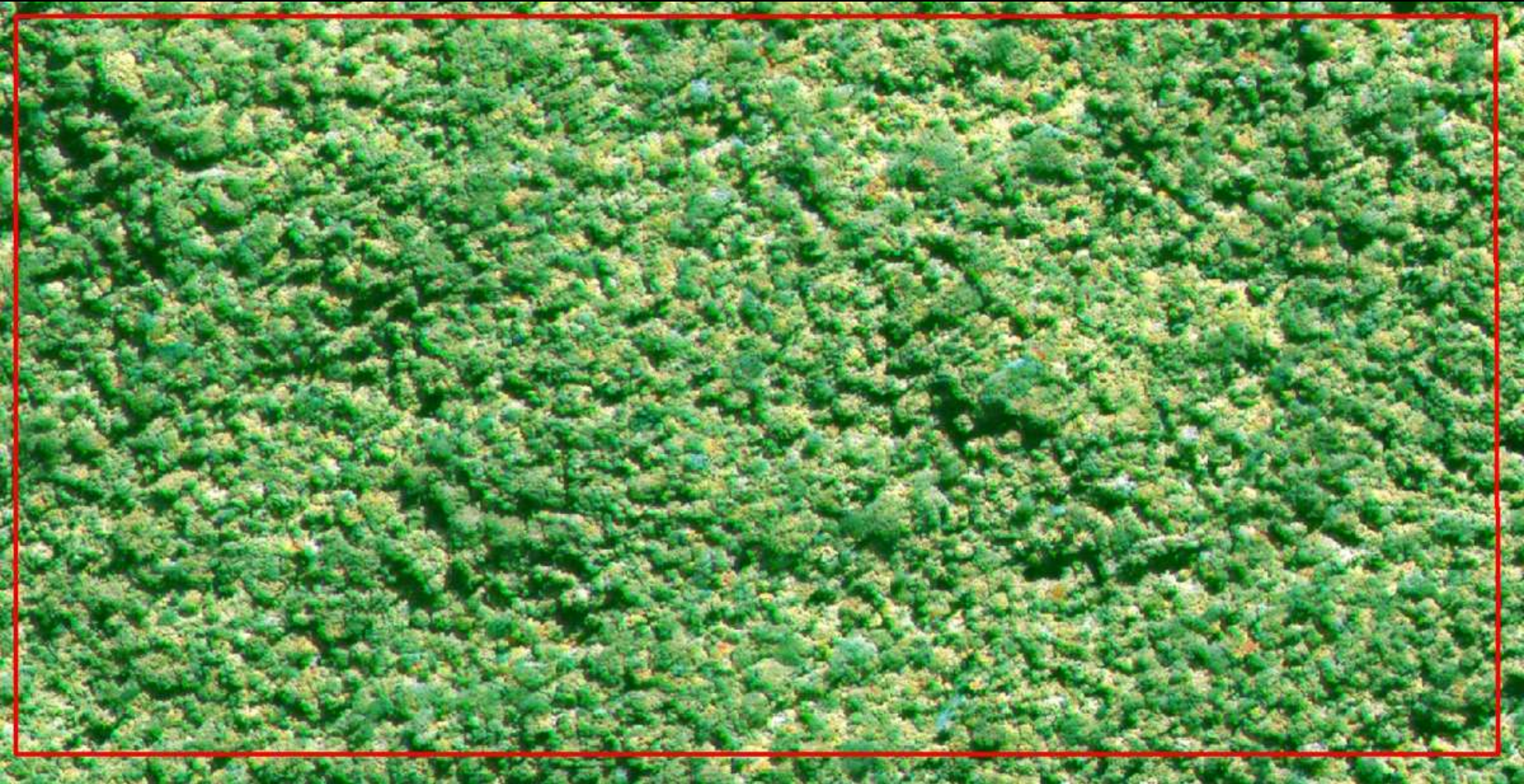


288,204 plant stems
with DBH > 1 cm,
536 species

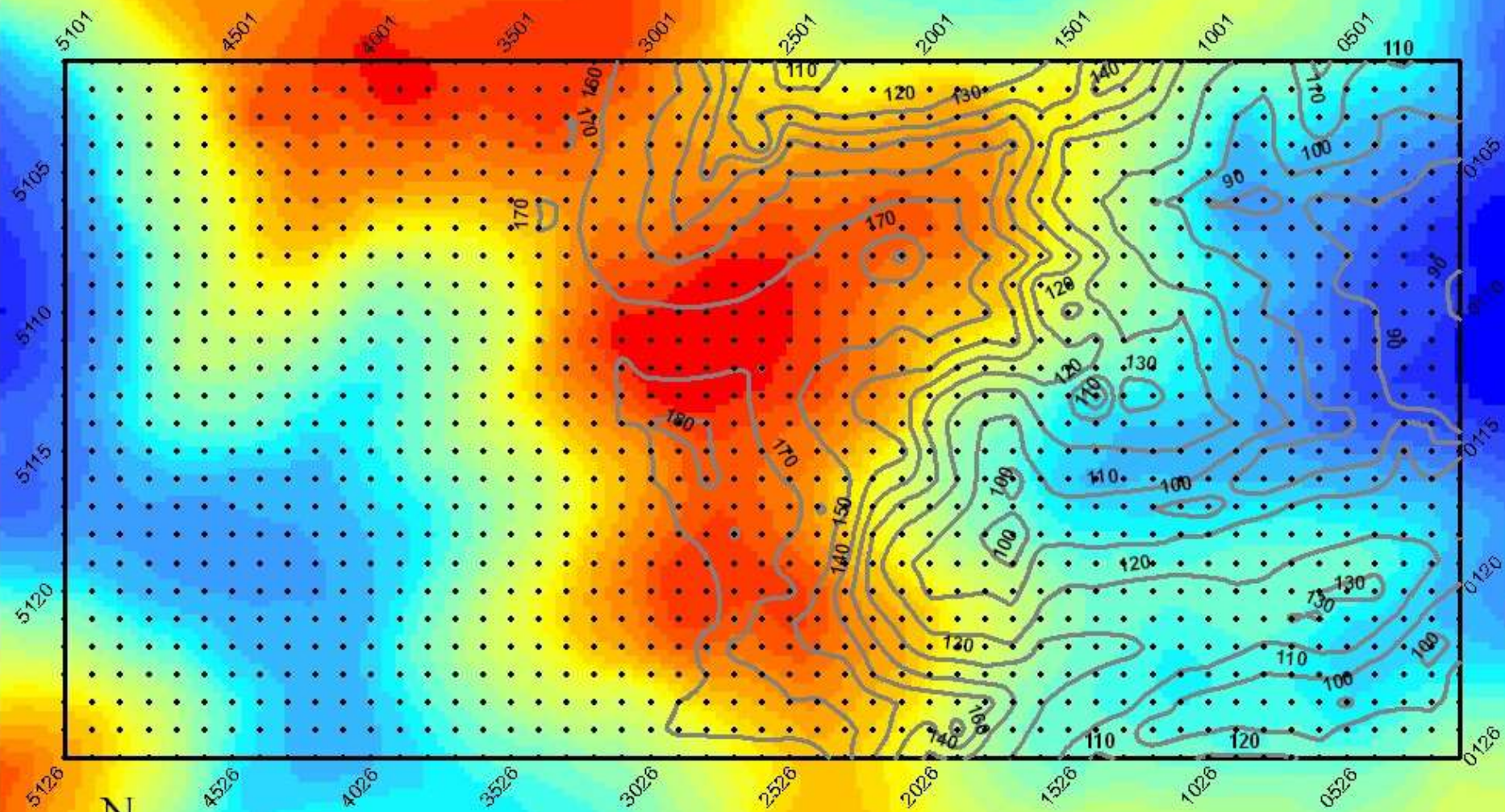


500 kg of aluminium tags,
50 km of copper wire

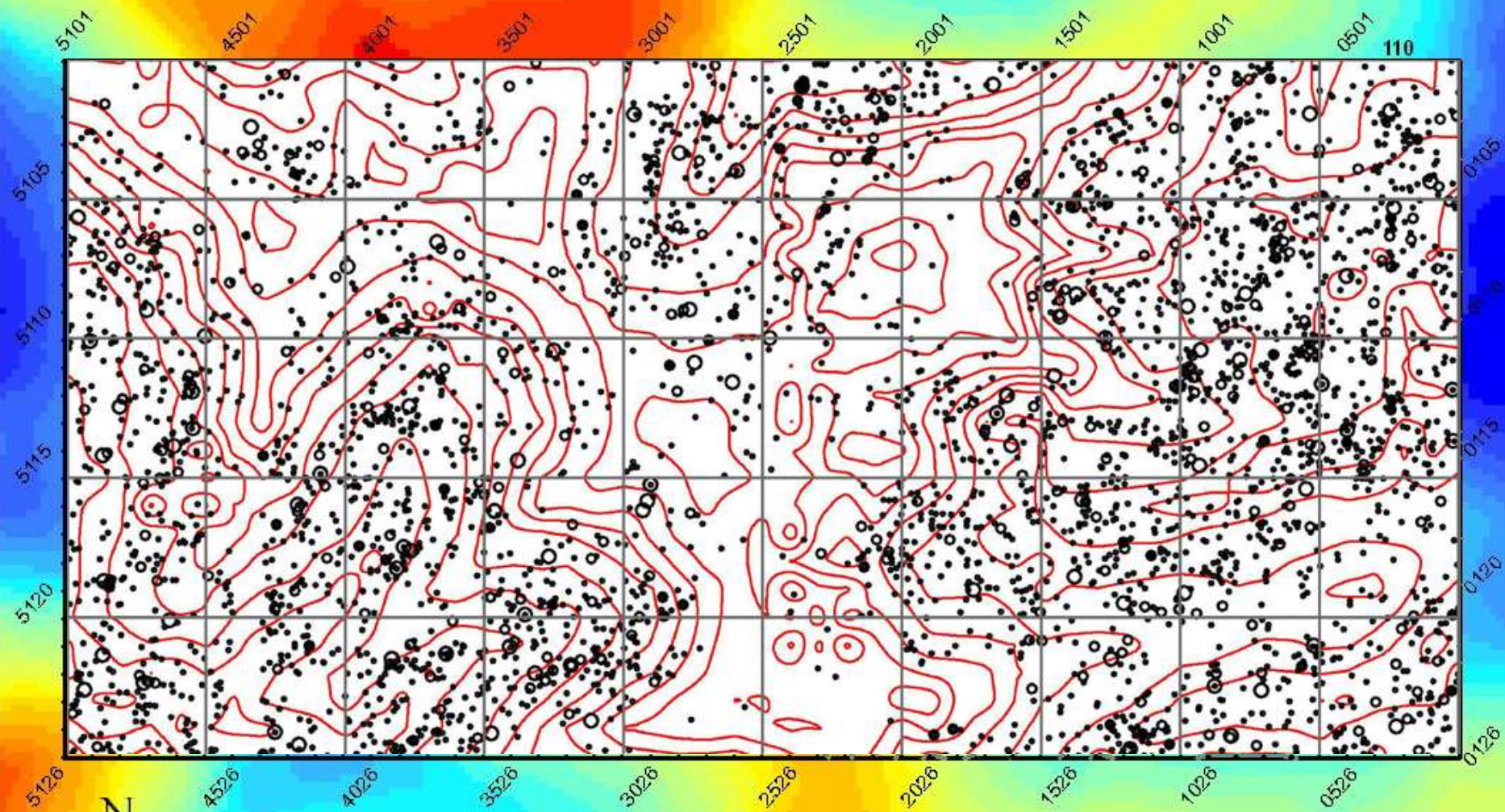
Landsat satellite image of 50 ha plot in Wanang
Papua New Guinea



Wanang 50ha Plot

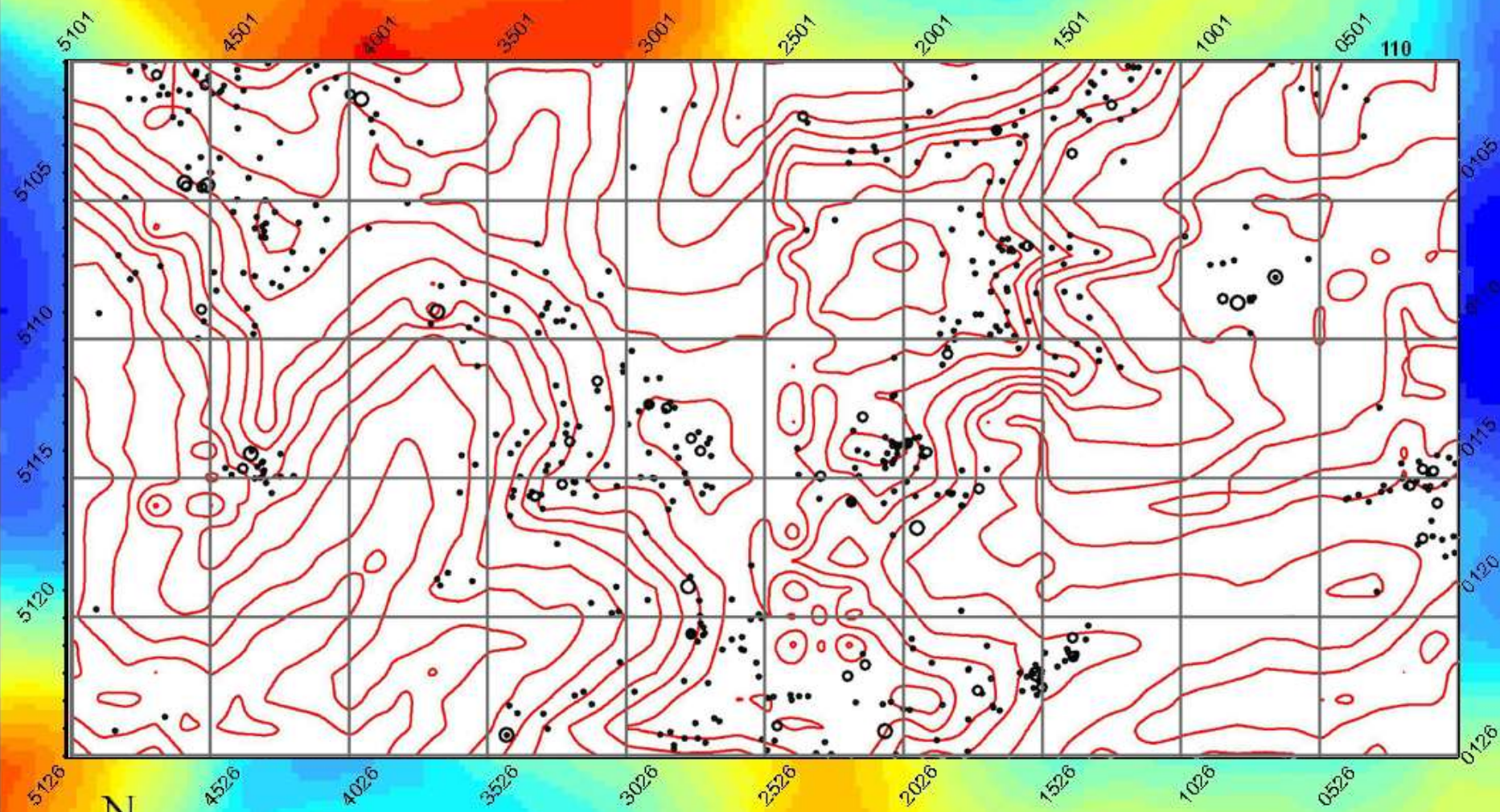


Wanang 50ha Plot



Pometia pinata

Wanang 50ha Plot

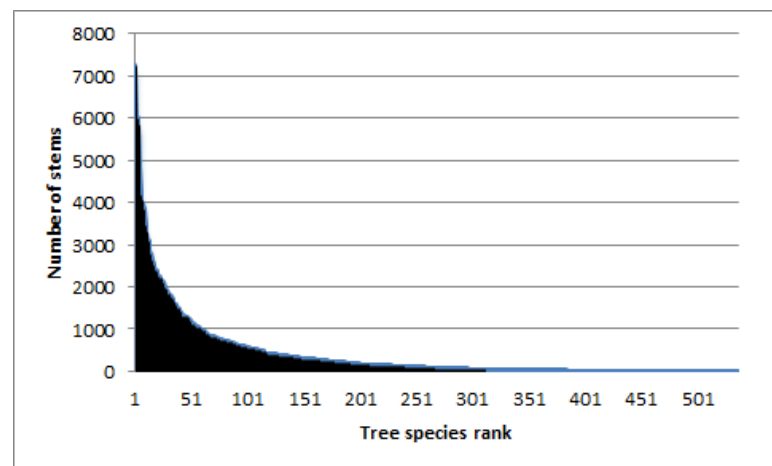
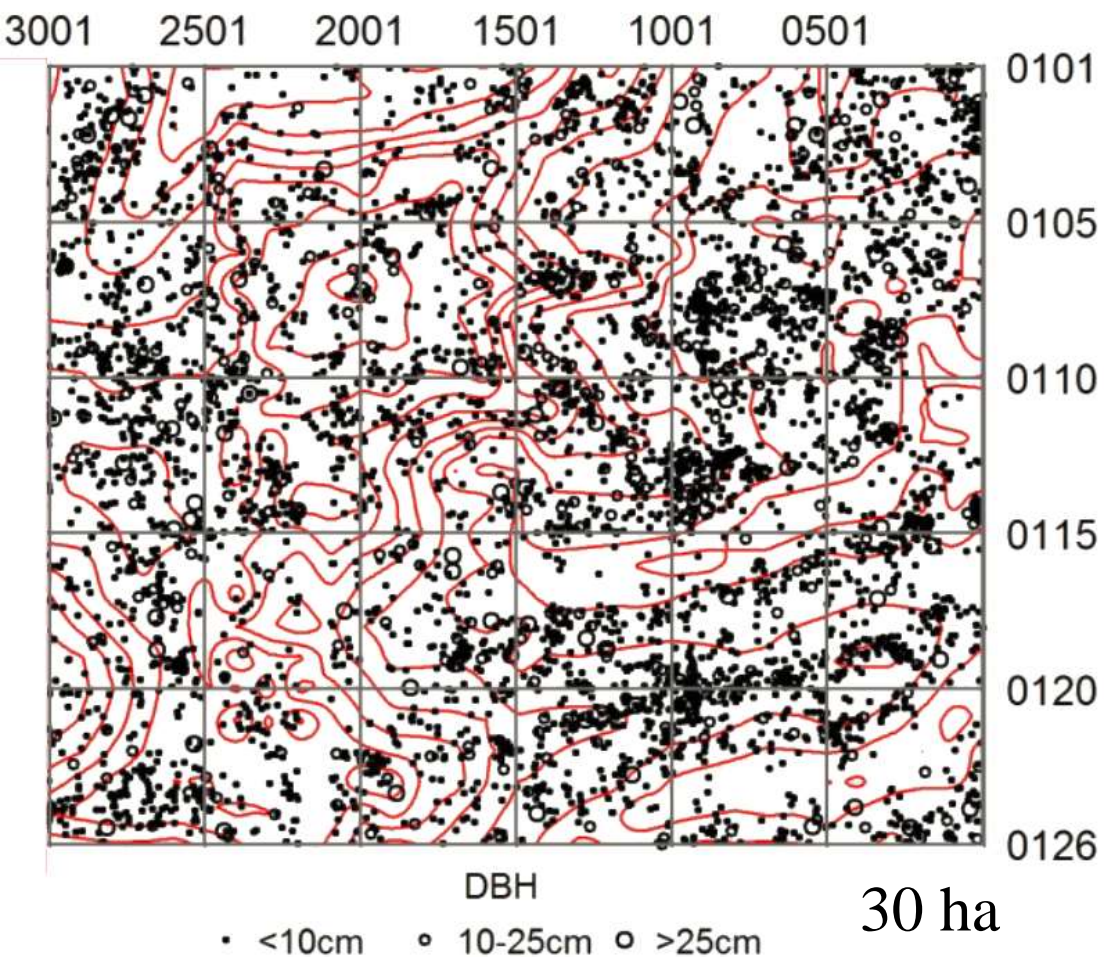


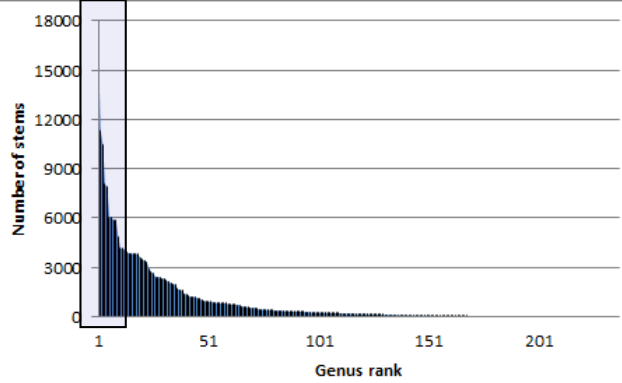
Intsia bijuga



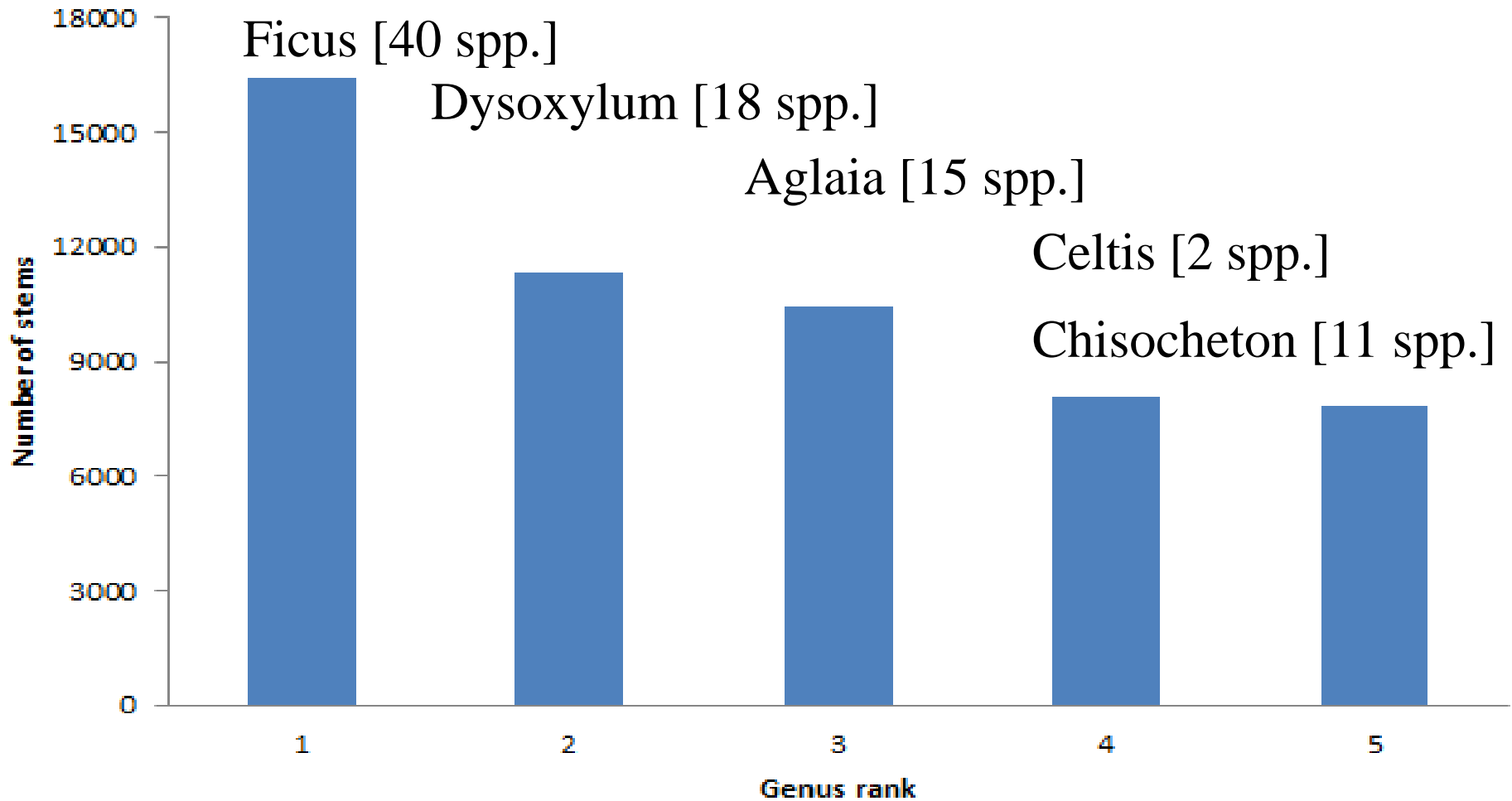
Celtis latifolia

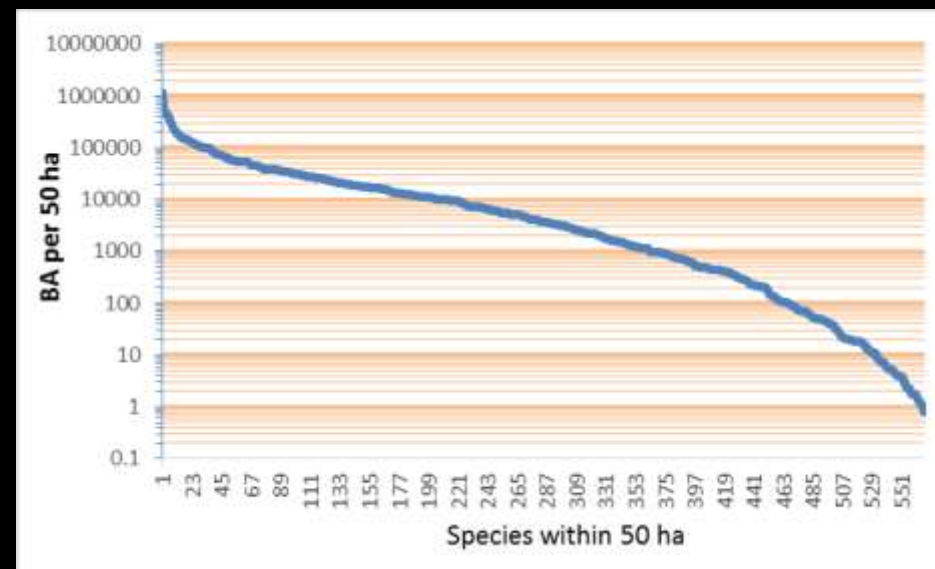
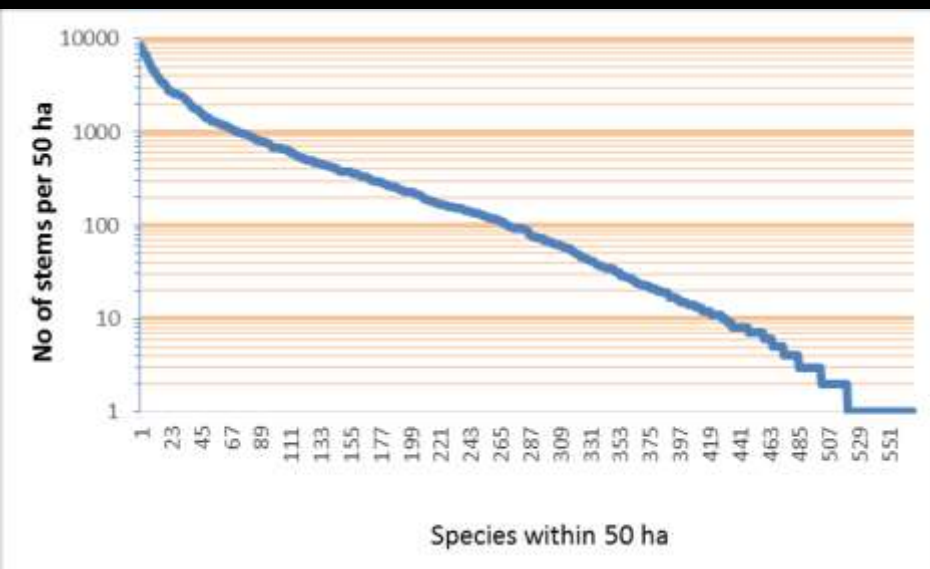
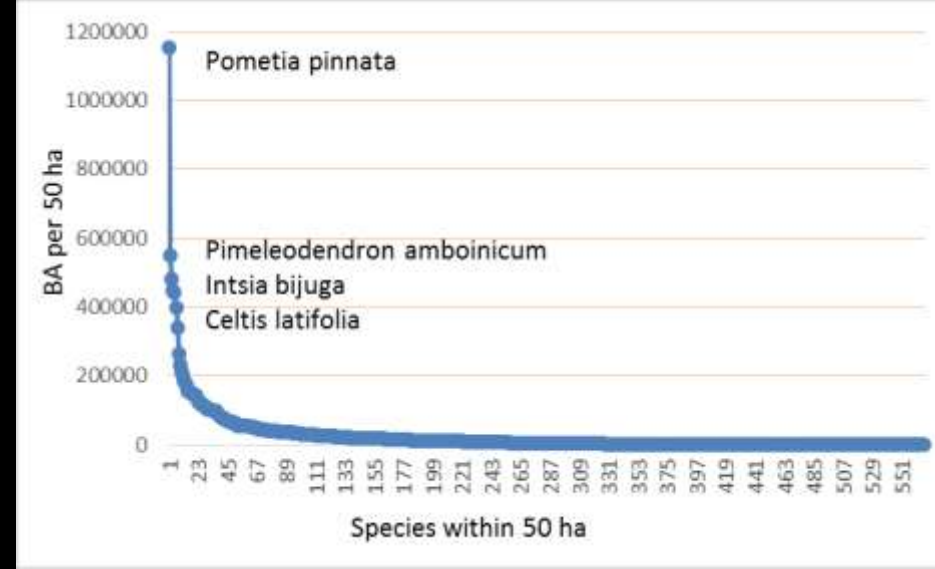
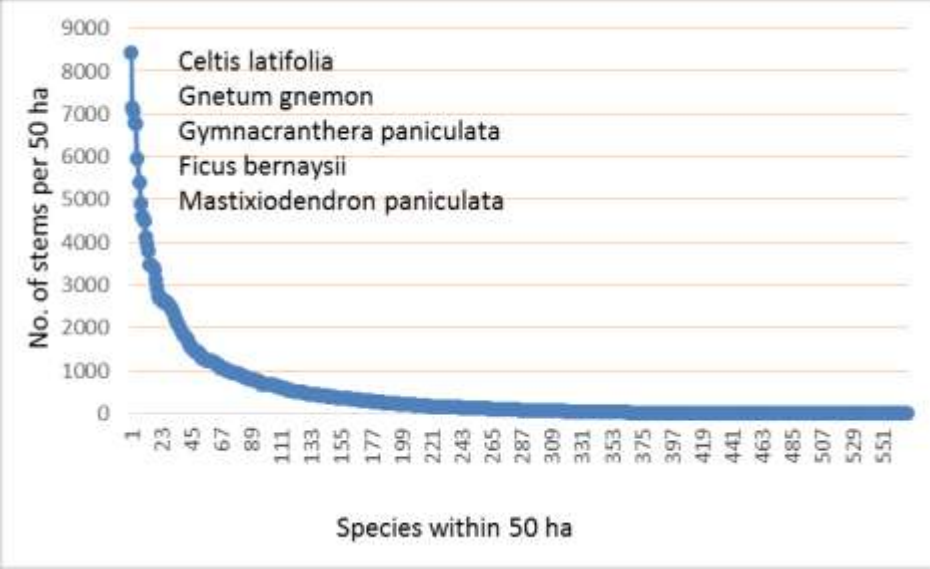
3.4% of all stems

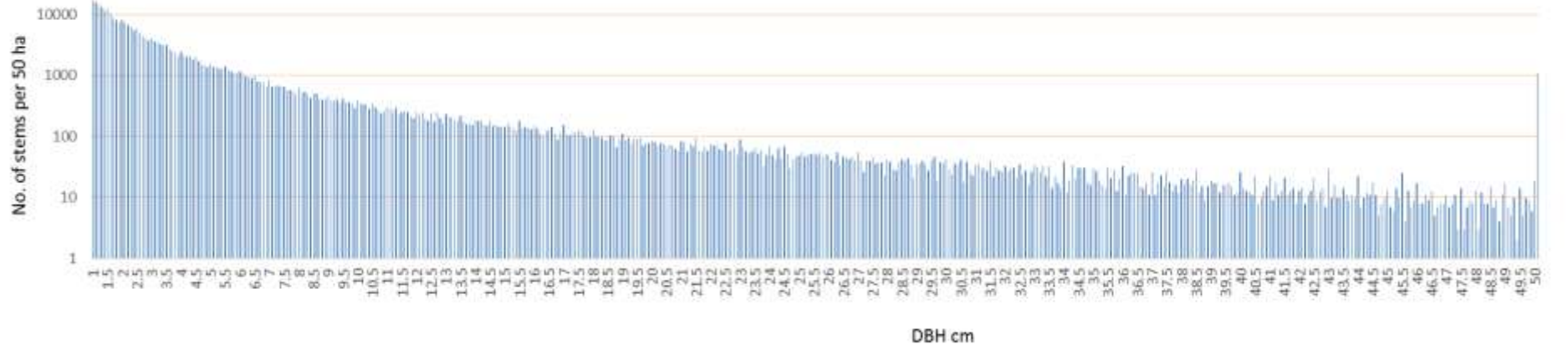
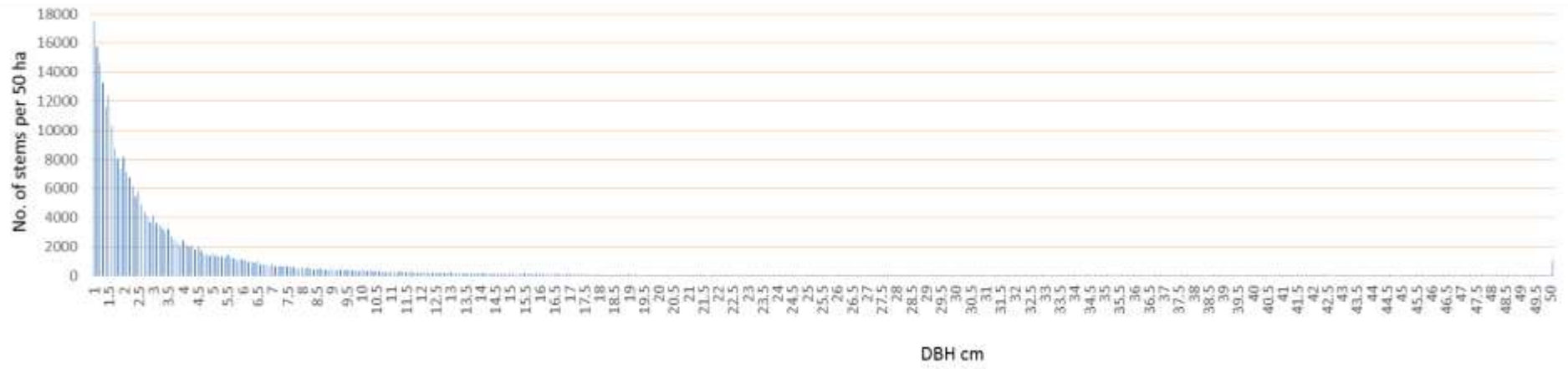


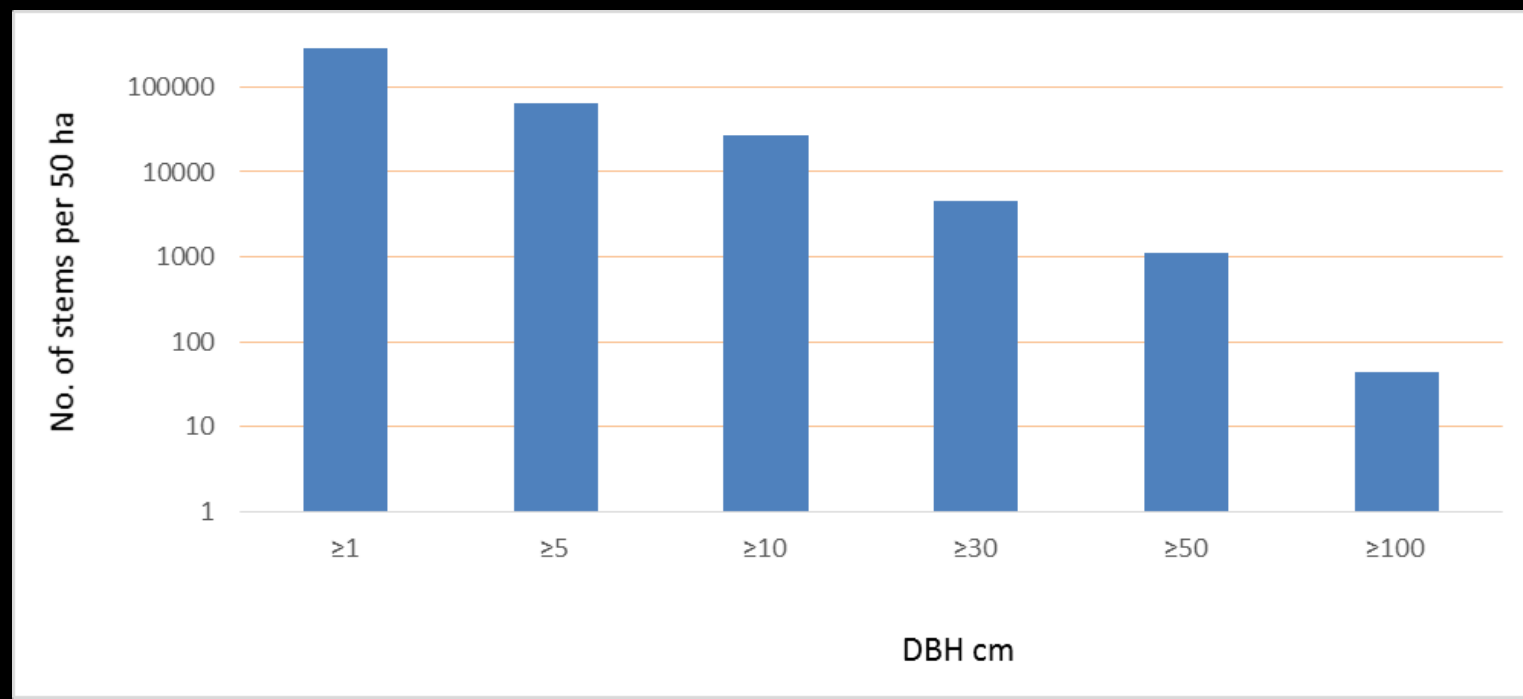
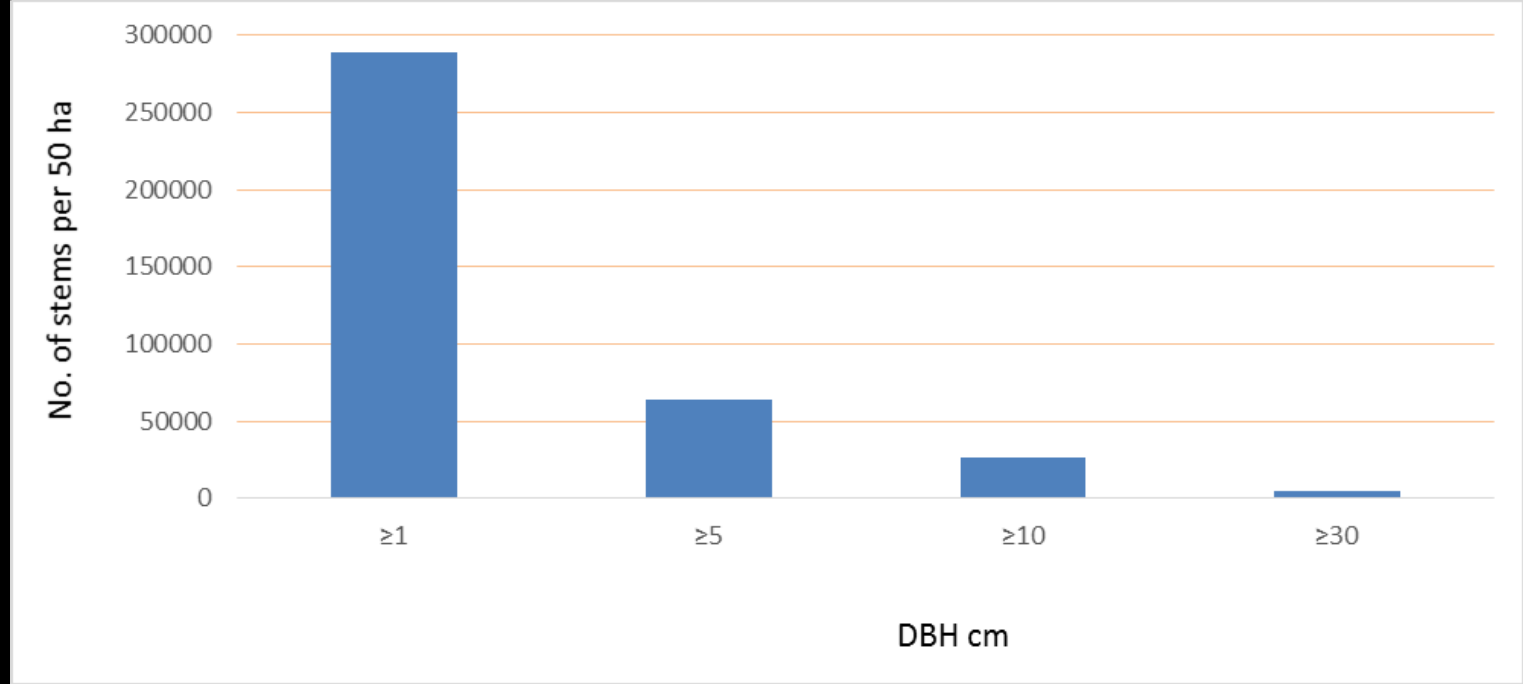


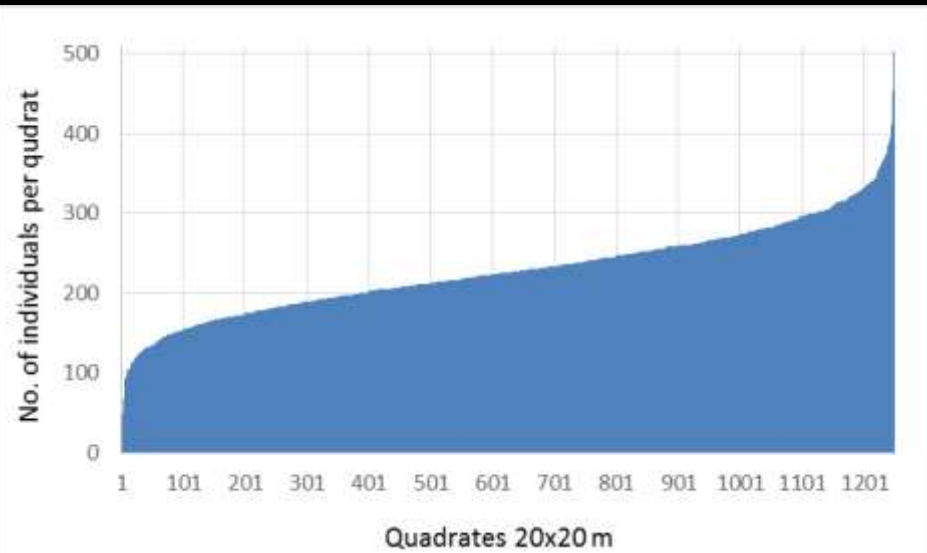
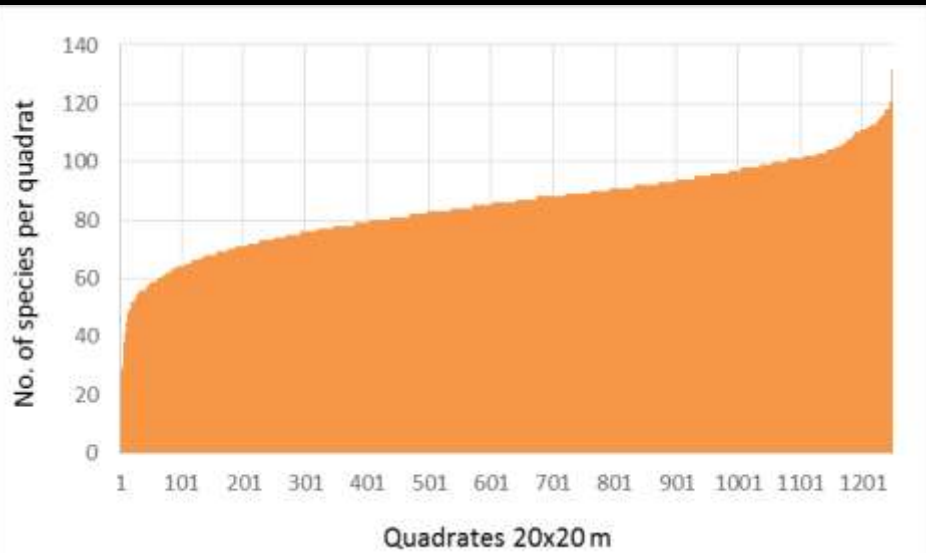
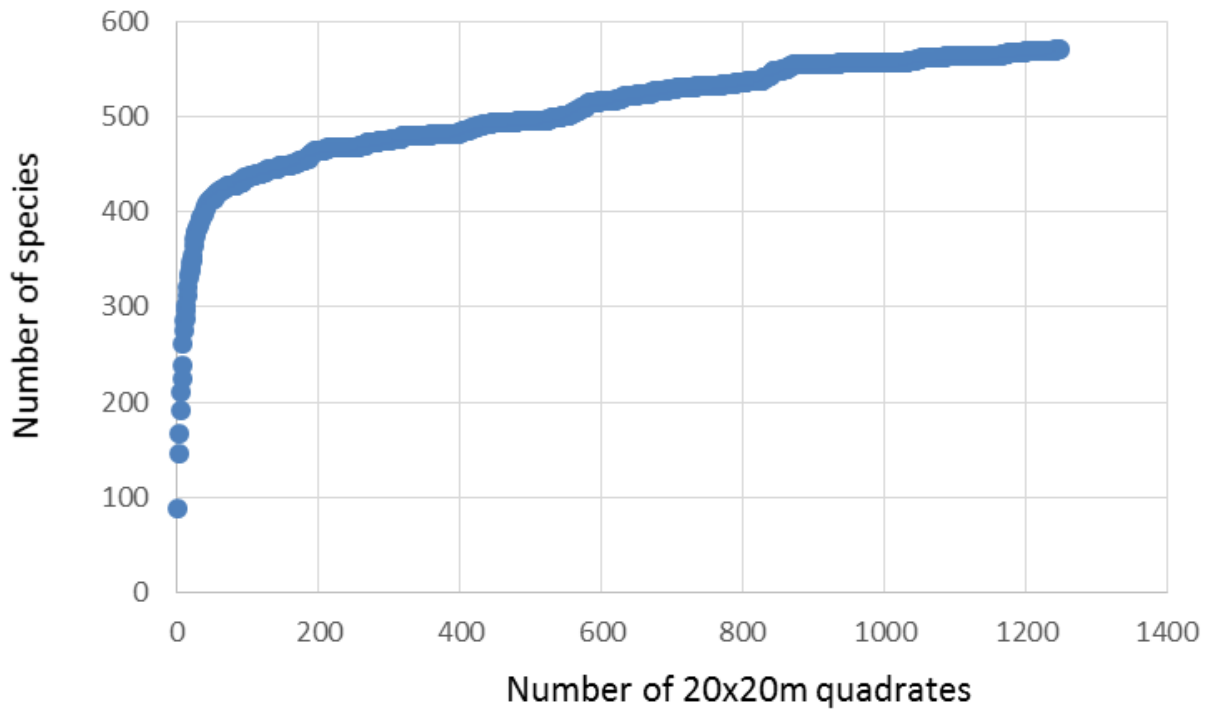
Ficus
7.7% of all stems



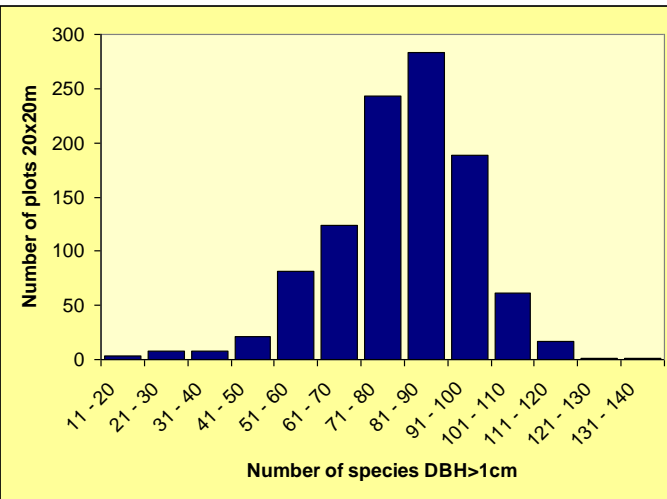
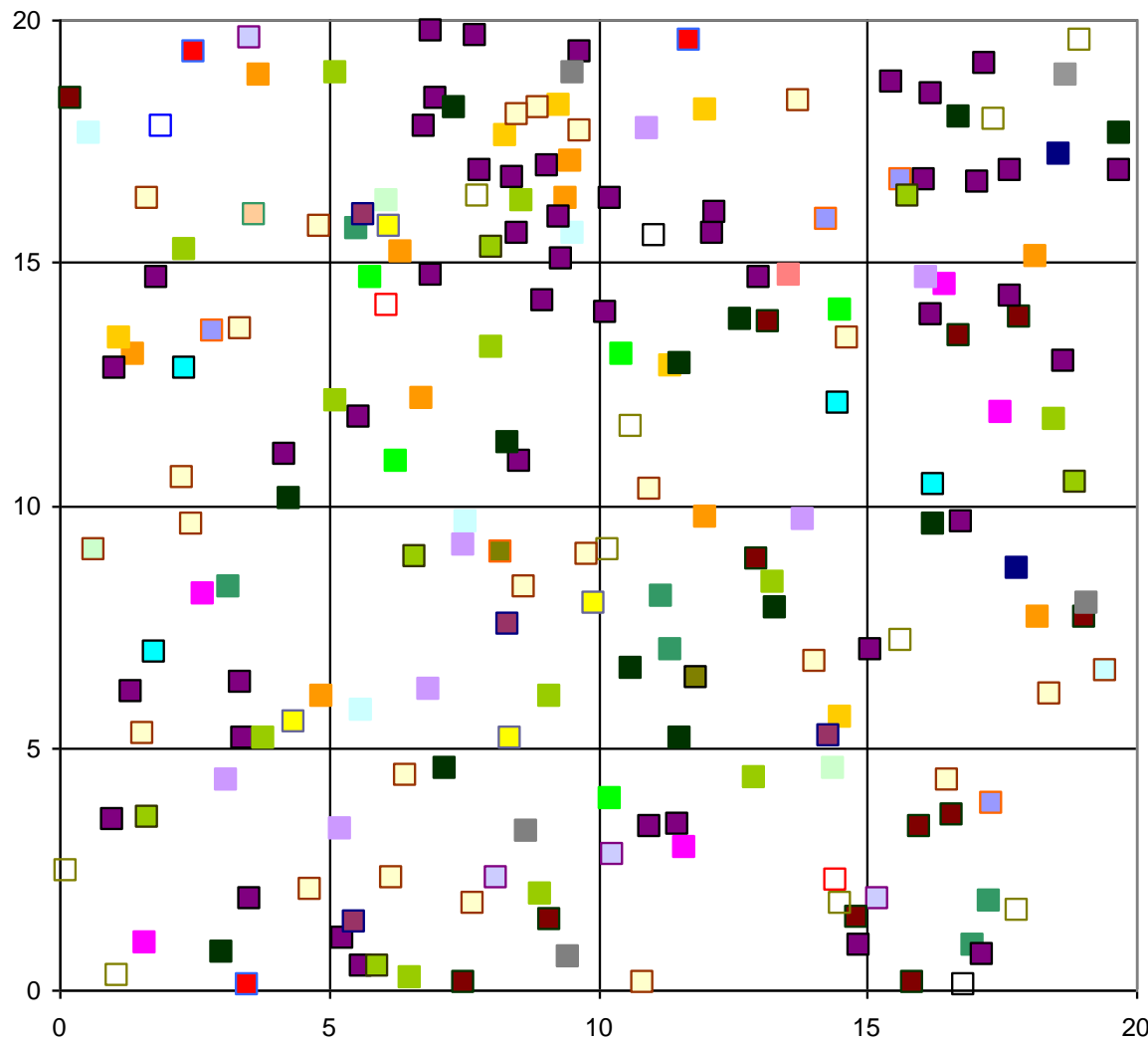








Wanang:
 20 x 20 m plot
 DBH > 1 cm
 34 plant families



modus
 81 – 90 species
 per 20x20 m plot
 DBH > 1 cm

- | | | | | |
|-----------------|----------------|--------------------|------------------|------------------|
| ■ Acanthaceae | ■ Annonaceae | ■ Apocynaceae | ■ Arecaeae | ■ Burseraceae |
| ■ Cannabaceae | ■ Clusiaceae | ■ Elaeocarpaceae | ■ Euphorbiaceae | ■ Fabaceae |
| ■ Gnetaceae | ■ Icacinaceae | ■ Lauraceae | ■ Loganiaceae | ■ Malvaceae |
| ■ Meliaceae | ■ Monimiaceae | ■ Moraceae | ■ Myricaceae | ■ Myrtaceae |
| ■ Nyctaginaceae | ■ Oleaceae | ■ Pentaphylacaceae | ■ Phyllanthaceae | ■ Pittosporaceae |
| ■ Podocarpaceae | ■ Polygalaceae | ■ Rosaceae | ■ Rubiaceae | ■ Salicaceae |
| ■ Sapindaceae | ■ Sapotaceae | ■ Simaroubaceae | ■ Stemonuraceae | |



ŽOFÍNSKÝ PRALES

ZALOŽEN: 28.8.1838

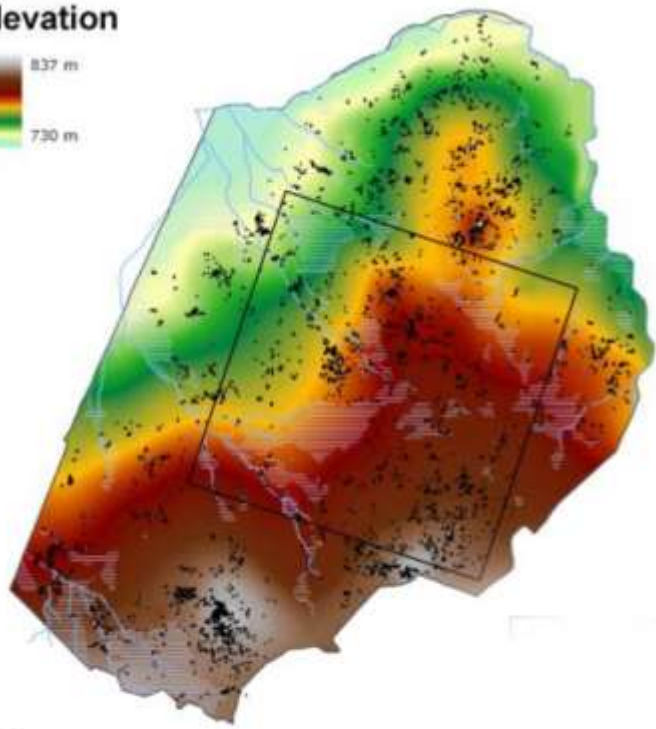
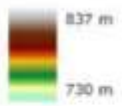
VÝMĚRA: 98,10 ha

ŽOFÍN
28.8.1838

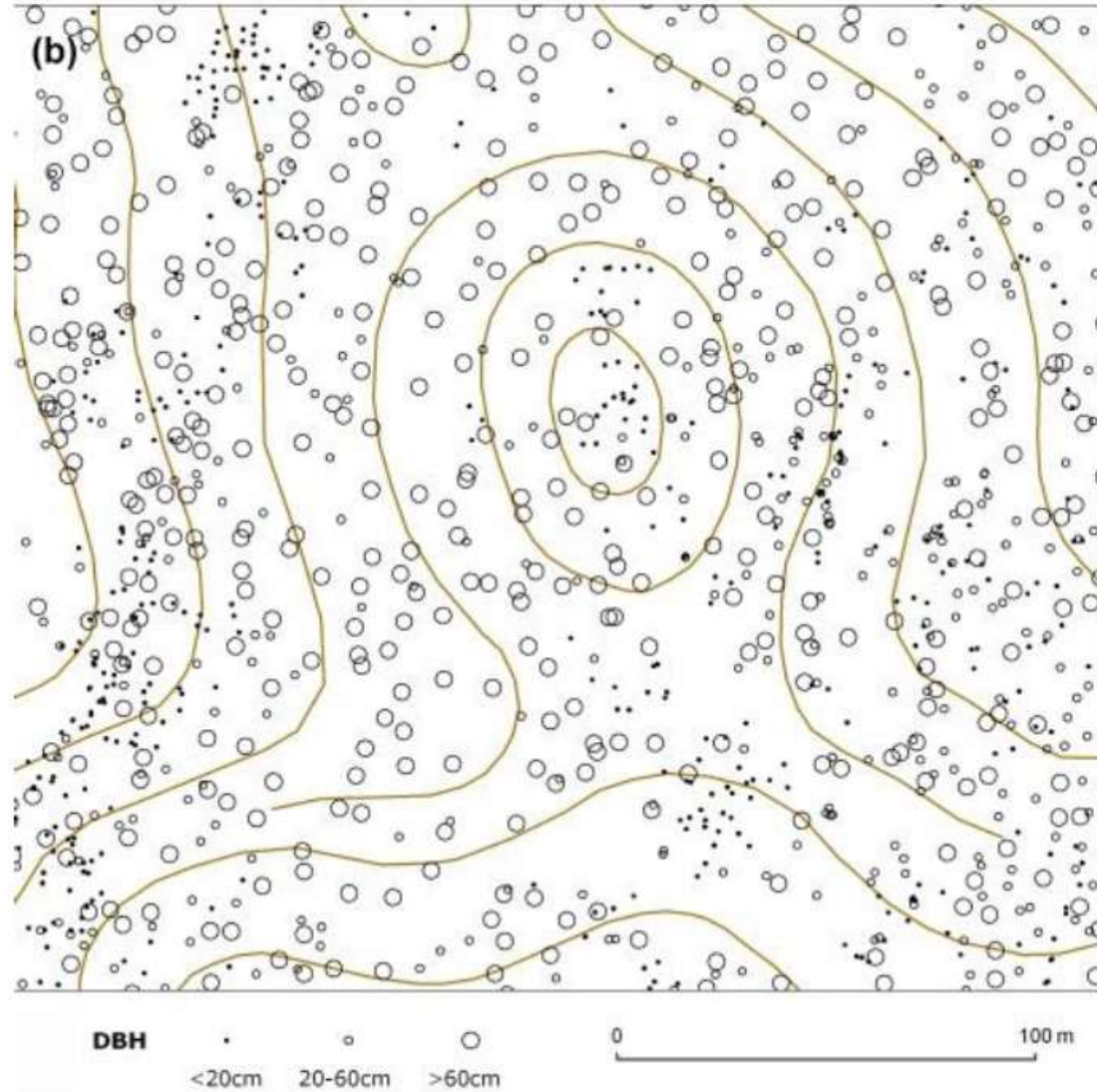
LEZ NOVÉ HRADY
ČSVTSLZ NOVÉ HRADY



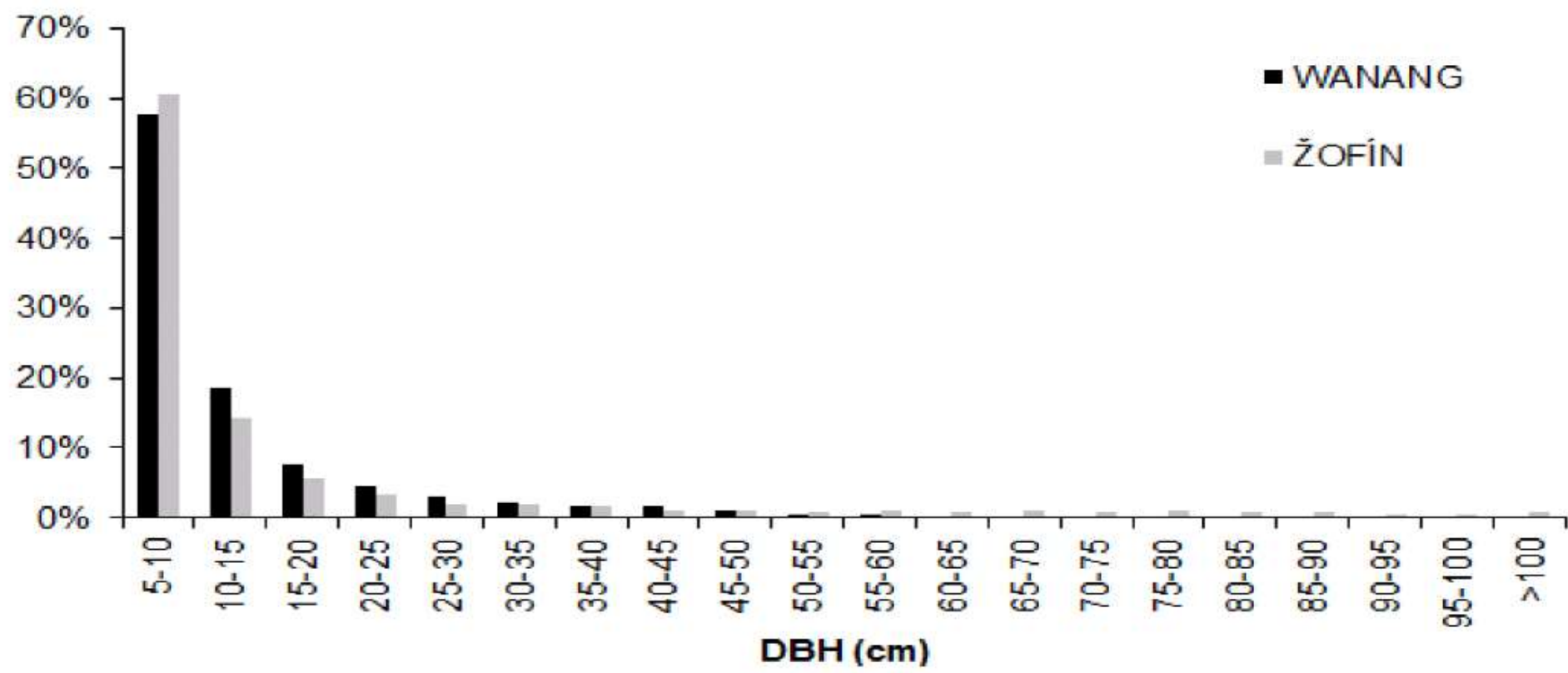
Elevation



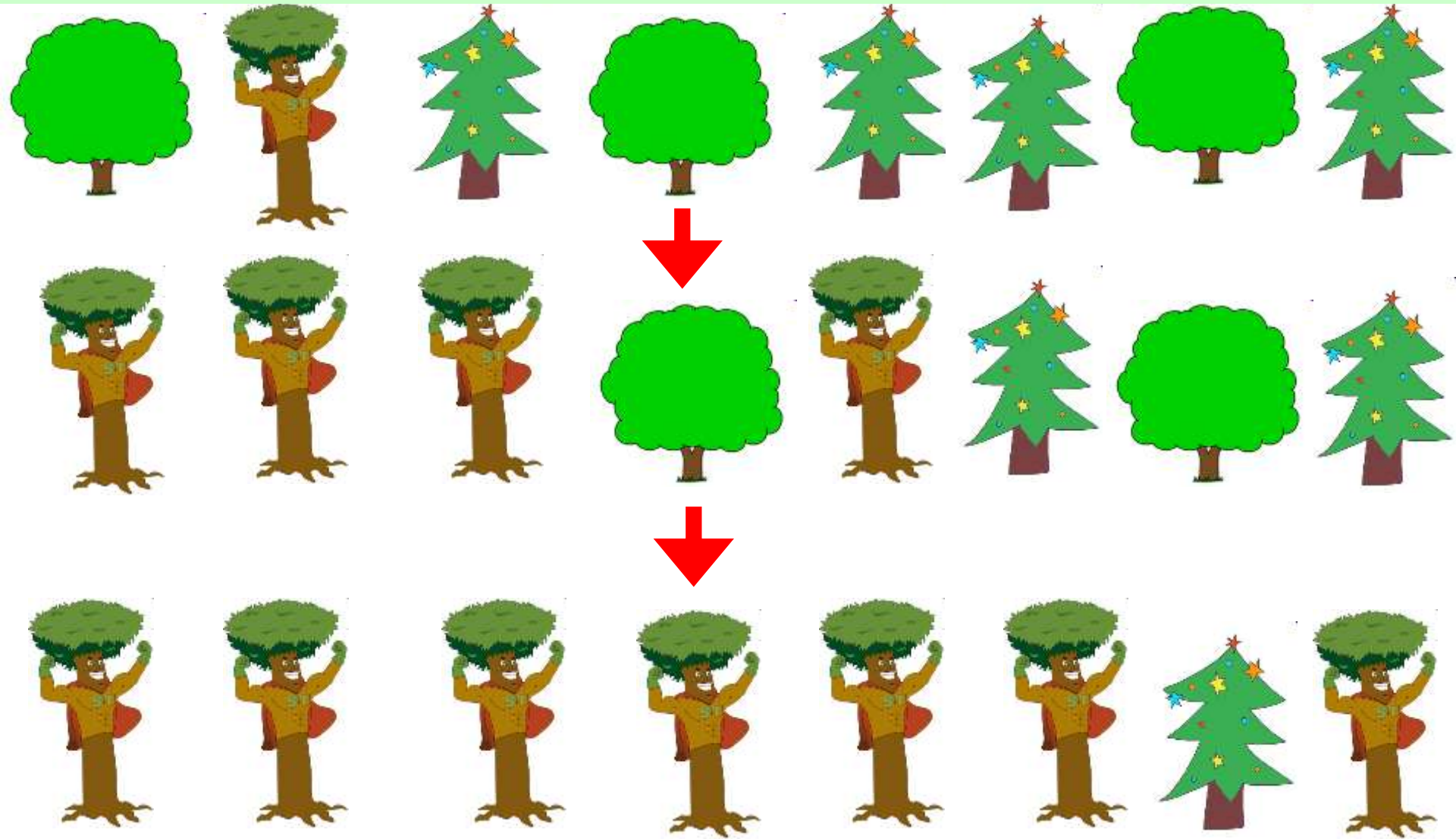
Fagus sylvatica distribution



| | Žofín Forest Dynamic Plot | Wanang, Papua New Guinea |
|----------------------------|------------------------------|-----------------------------|
| Number of tree species | 11 | 536 |
| Dominant tree species | <i>Fagus sylvatica</i> | <i>No dominant</i> |
| Area (ha) | 25 | 50 |
| Middle elevation (m) | 780 | 120 |
| Mean temp. (°C) | 6.2 | 25.8 |
| Annual prec. (mm) | 866 | 4000 |
| Years of census | 2012 | 2010-2 |
| Last tree count | 64345 | 288204 |
| DBH threshold (cm) | 1 | 1 |
| Tree density: stems per ha | 2,574 | 5,764 |



How can there be hundreds of tree species coexisting in tropical forests when they all compete for the same solar energy and nutrients?



What prevents a single "superman" tree species, best adapted to local conditions, from out-competing others and prevailing in the rainforest?

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Each species is limited by a different resource
[niche differentiation]

The environment is variable, changing direction of competition
[tree gap dynamics]

Immigration prevents competitive exclusion
[non-equilibrium community composition]

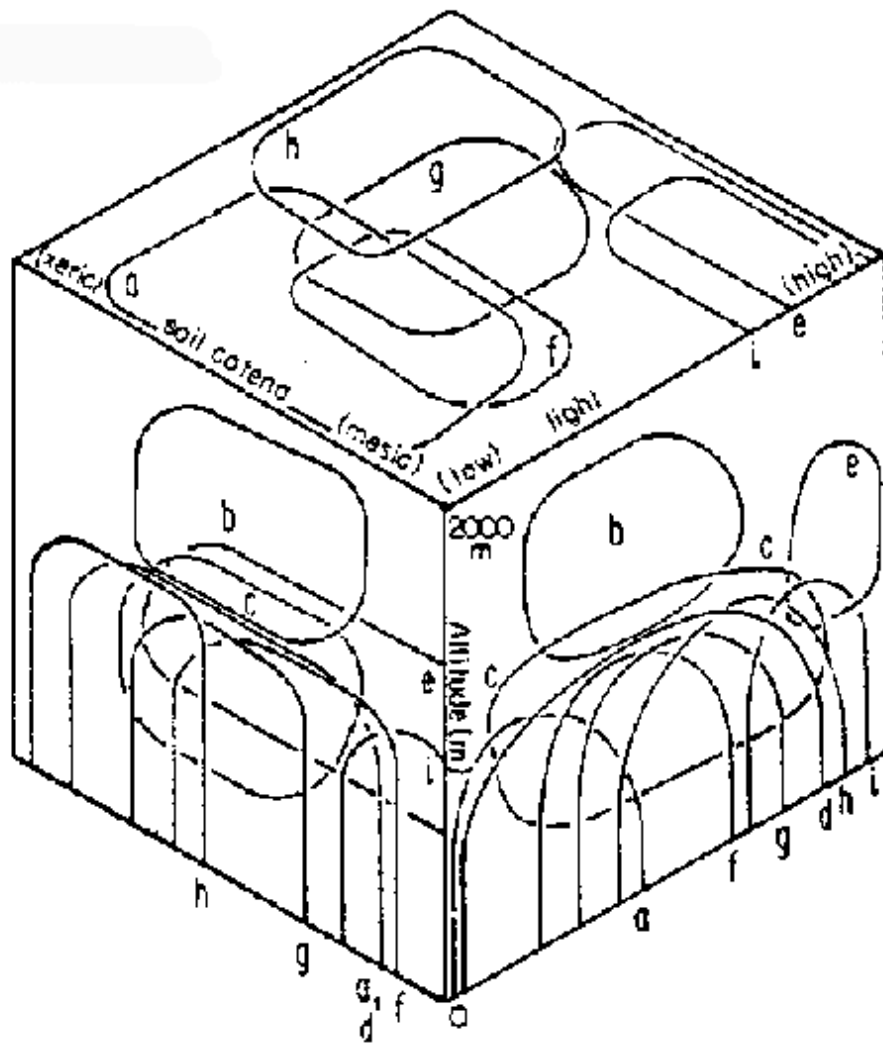
Species do not have opportunity to compete
[recruitment limitation]

Rare species are demographically favoured
[density-dependent mortality, predation and parasitism]

Recommended reviews:

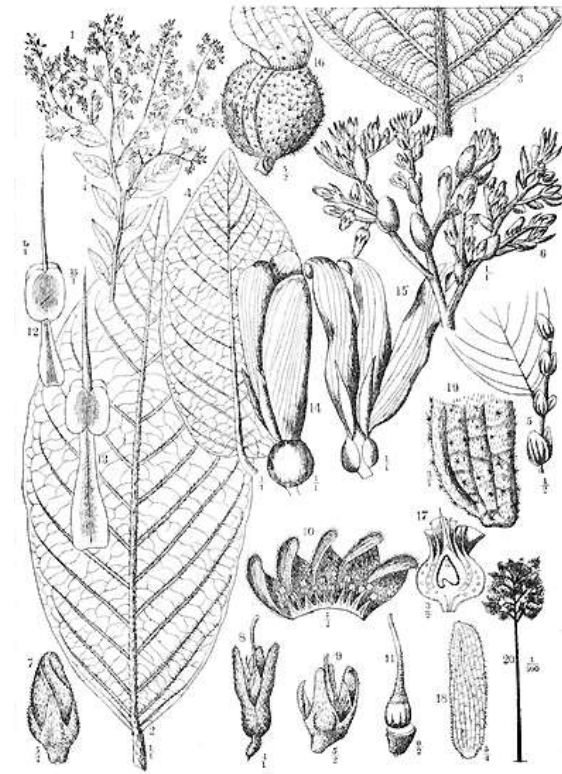
Leigh et al. 2004. Why Do Some Tropical Forests Have So Many Species of Trees? *Biotropica* 36: 447–473

Wright, S. J. 2002. Plant diversity in tropical forests: A review of mechanisms of species coexistence. *Oecologia* 130: 1–14.



Icones Biotropenses IV.

Tab. LXXX.



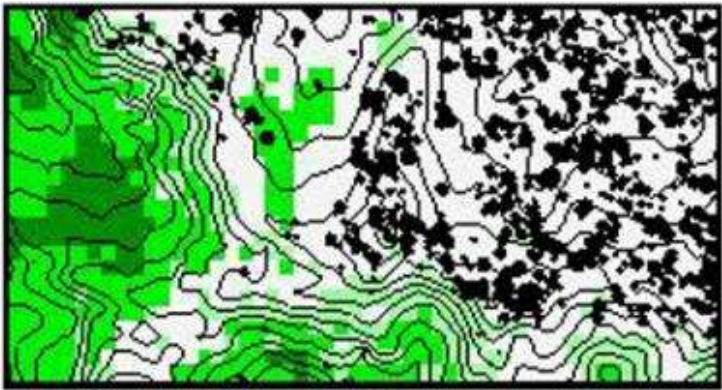
Shorea Koortersii Brandis.

Figure 2 Schematic ecological range, in the wet zone forests of Sri Lanka, of the species of Shorea, section Doona. Key: a. Shorea cordifolia b. S. gardneri c. S. zeylanica d. S. ... e. S. trapezifolia f. S. megistophylla g. S. disticha h. S. worthingtonii i. S. confiflora

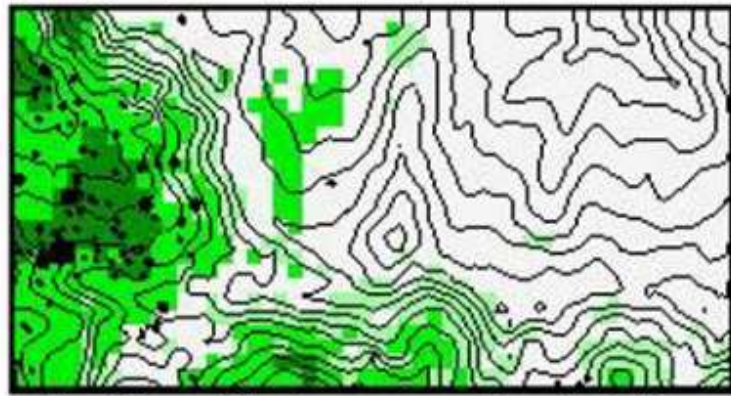
SHOREA

Niche differences and environmental heterogeneity

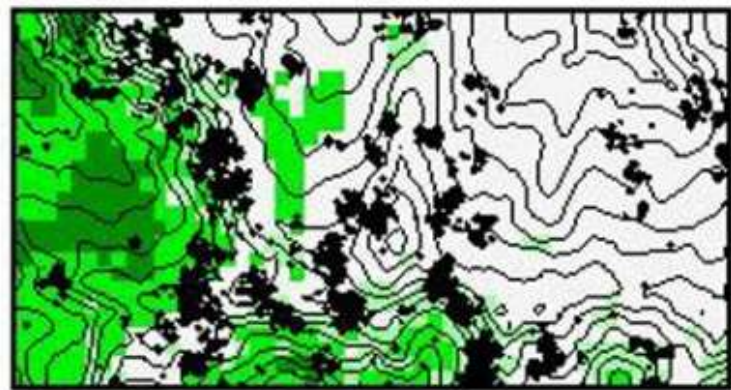




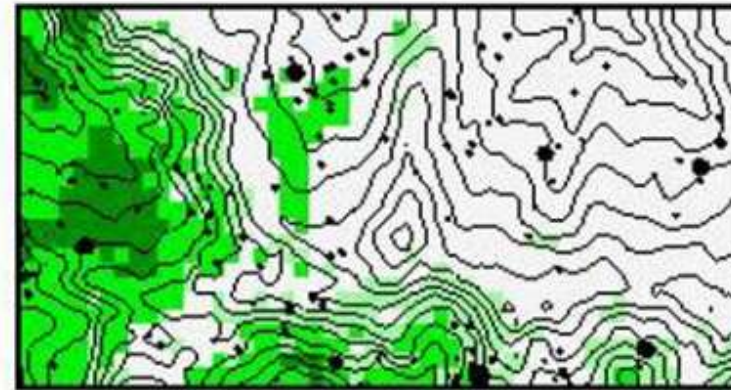
A. Sandy loam (*S. acuta*)



D. Clay (*S. inappendiculata*)

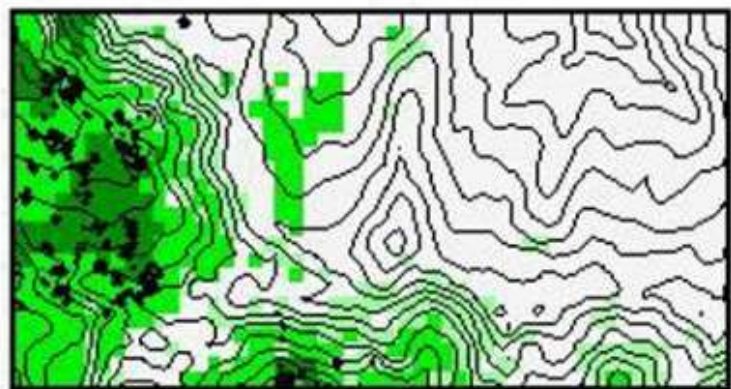


B. (*S. amplexicaulis*)



E. Generalist (*S. ochraceae*)

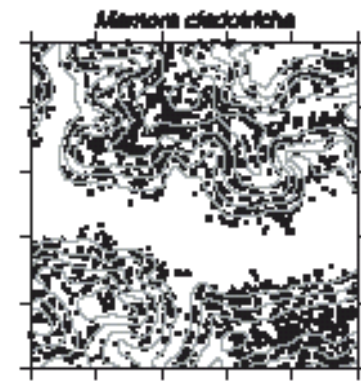
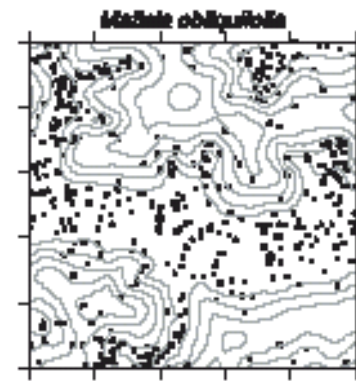
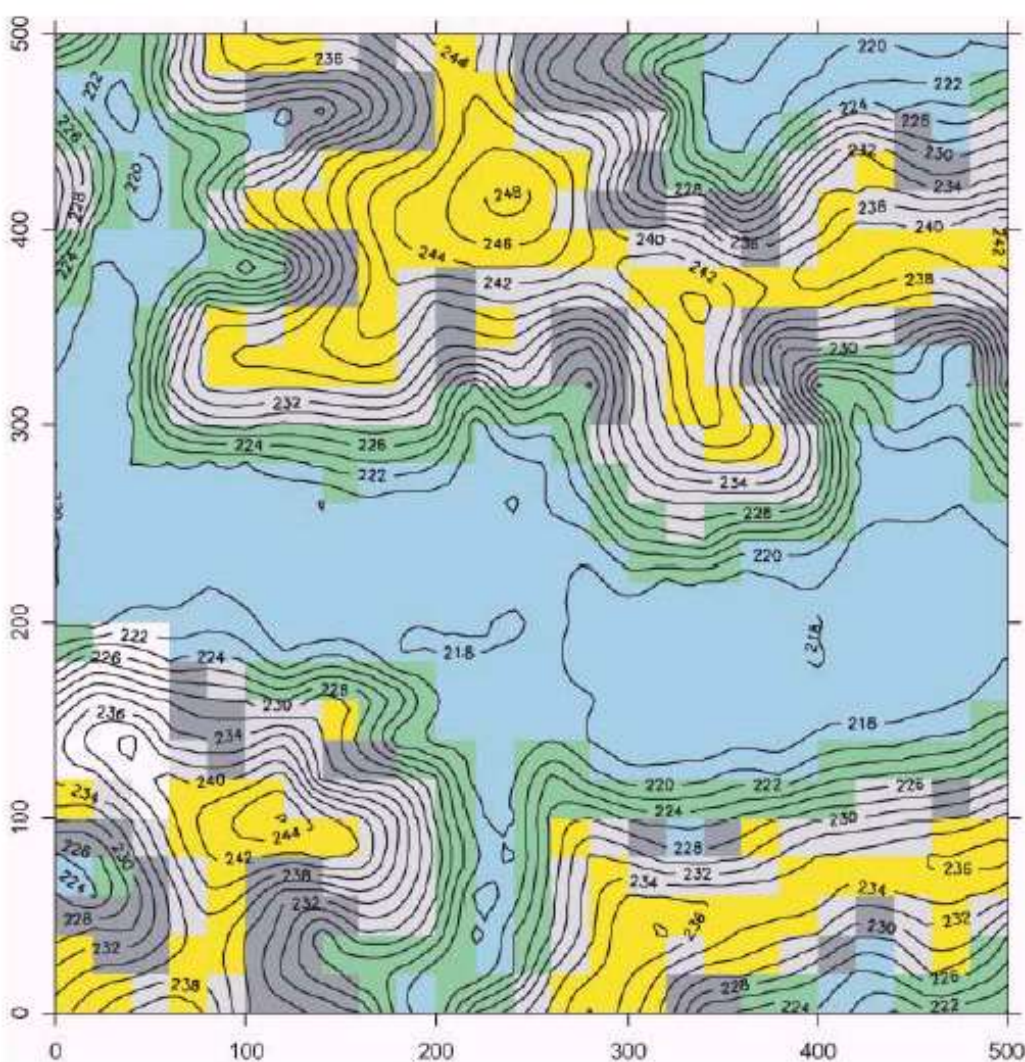
Shorea
in Lambir Hills



C. Fine loam (*S. xanthophylla*)

*increasing fertility and
moisture:*
white, sandy loam
light green, loam
green, fine loam
dark green, clay

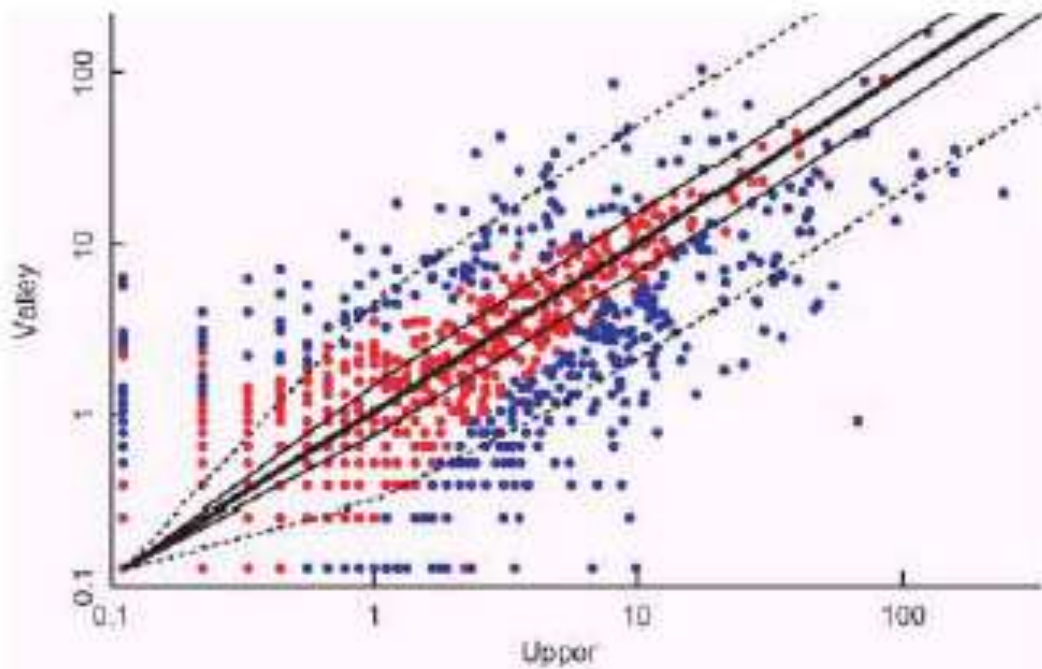




blue: valley
 green: low slope
 dark grey: high gully
 light grey: upper slope
 yellow: ridge top
 white: disturbance

Fig. 2 Topographic map of the 25-ha plot, with 2-m contour intervals. Numbers marking each line are metres above sea level. Six habitats are indicated: valley (blue), low-slope (green), high-gully (dark grey), upper-slope (light grey), ridge-top (yellow), and secondary forest (white). Axes are marked in metres; north is up.

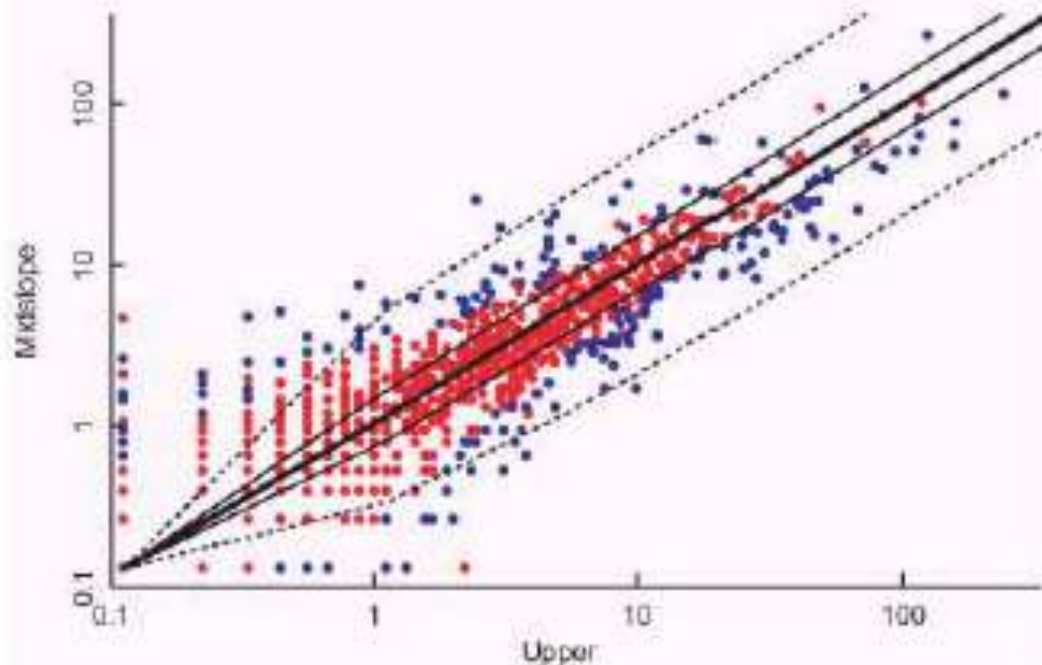
Distribution of plants with respect to topography and habitats Lowland rainforest, Yasuni National Park, Peru



Habitat preference
of tree species:
blue - significant
red - not significant

A: valley x upper slope

B: mid-slope x valley



it vs.
fats)
gure
lines
axes,
icate

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Recommended reviews:

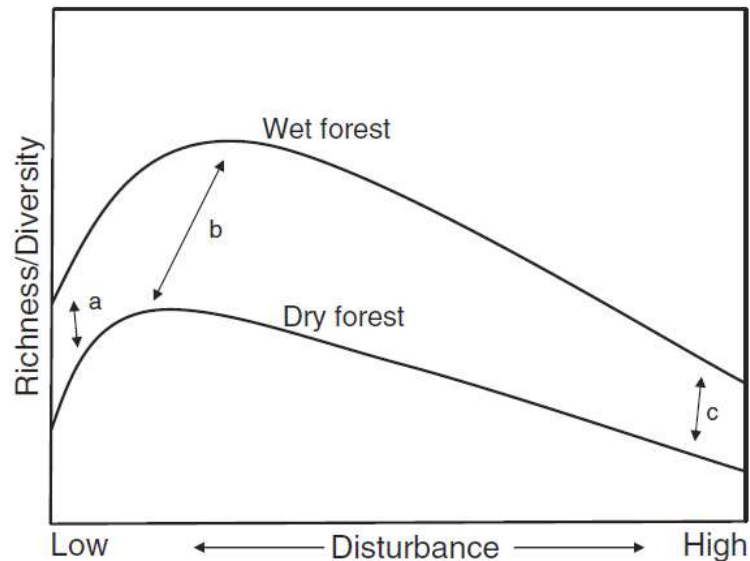
Leigh et al. 2004. Why Do Some Tropical Forests Have So Many Species of Trees? *Biotropica* 36: 447–473

Wright, S. J. 2002. Plant diversity in tropical forests: A review of mechanisms of species coexistence. *Oecologia* 130: 1–14.

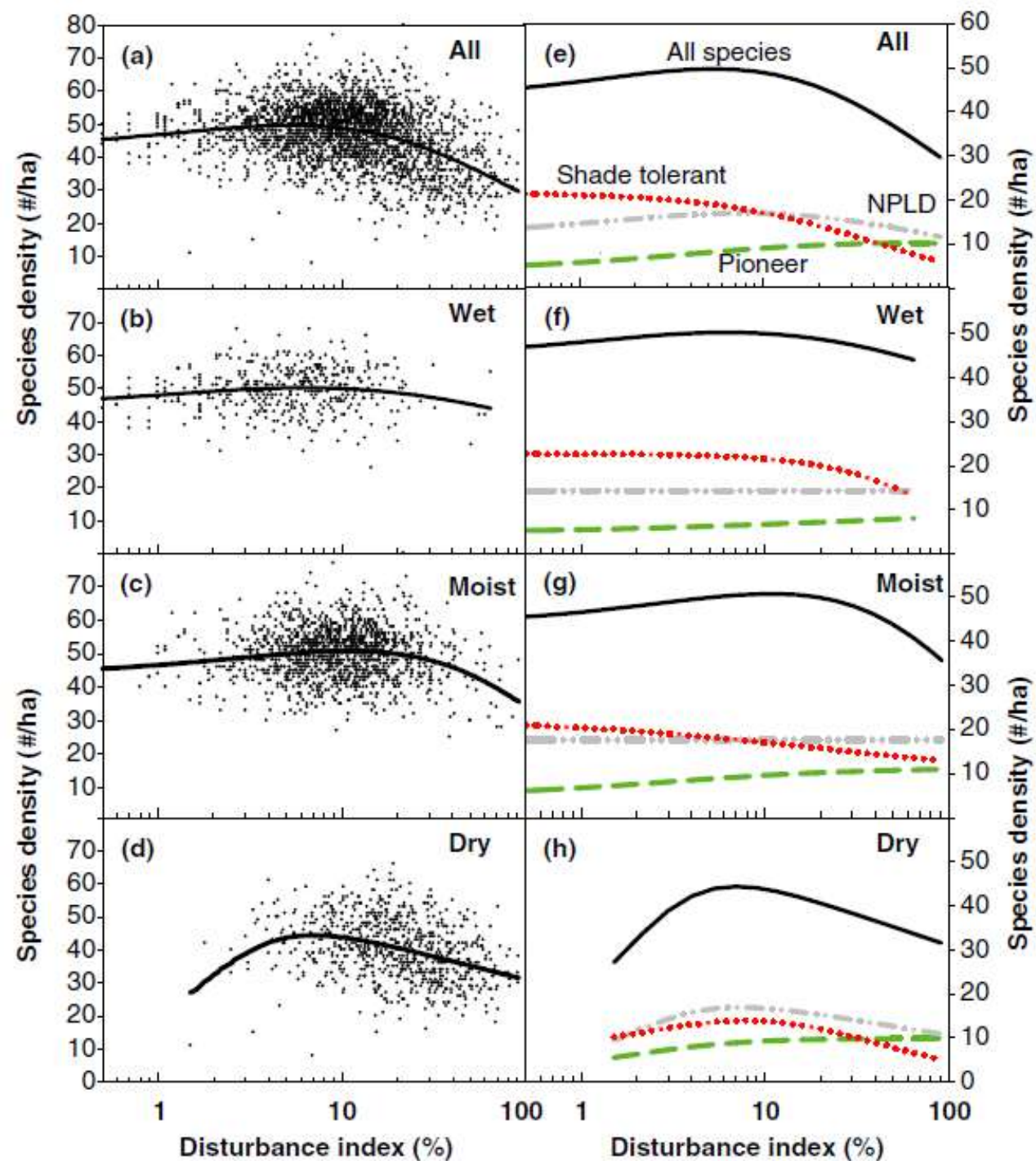
Intermediate Disturbance Hypothesis

Two possible mechanisms:
disturbance

- removes dominant species opening thus an opportunity for others
- creates new habitat [for pioneer spp.]



Bongers et al. 2009: The intermediate disturbance hypothesis applies to tropical forests, but disturbance contributes little to tree diversity. *Ecology Letters*, 12: 798–805



Diversity peaks at intermediate disturbance levels but little variation is explained outside dry forests

points = 1ha tropical forest plots [2504 plots],
 NPLD = non-pioneer light demanding spp.
 Disturbance index = % of trees that belong to the pioneer species

LONGEVITY

Shorter

Longer

MAXIMUM GROWTH RATE

Slower

Faster

| | |
|---------------------------------------|----------------------------------|
| I understory species | II subcanopy species |
| IV shade- intolerant species | III shade-tolerant species |

FIGURE 25.2. Diagrammatic classification of tree species according to maximum growth rate (slow v. fast) and projected lifespan (short v. long). *Group I*: understory species, *Group II*: slow-growing subcanopy species, *Group III*: fast-growing, shade-tolerant canopy and subcanopy species that respond opportunistically to increased light levels; *Group IV*: putative shade-intolerant canopy and subcanopy species.

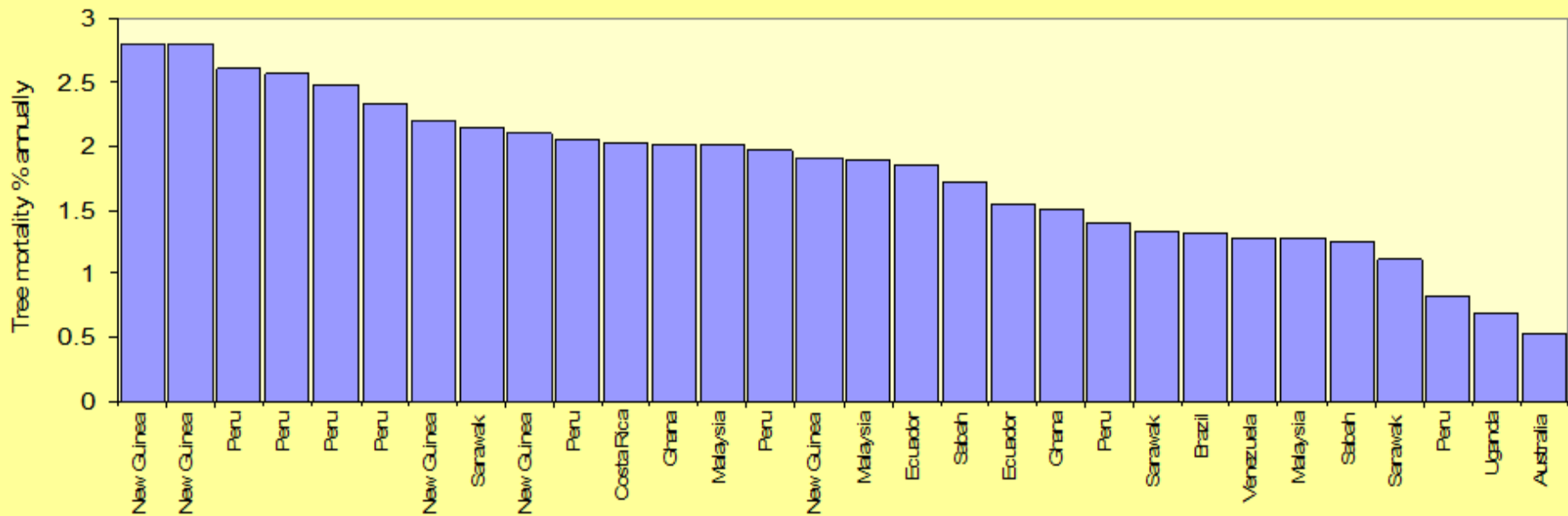
Problem:

pioneer trees always represent only small % of the total diversity



Rainforests are dynamic

Annual mortality of trees with DBH>10cm : 0.5 – 2.7%



Dynamic forests are more diverse: the mean mortality matters

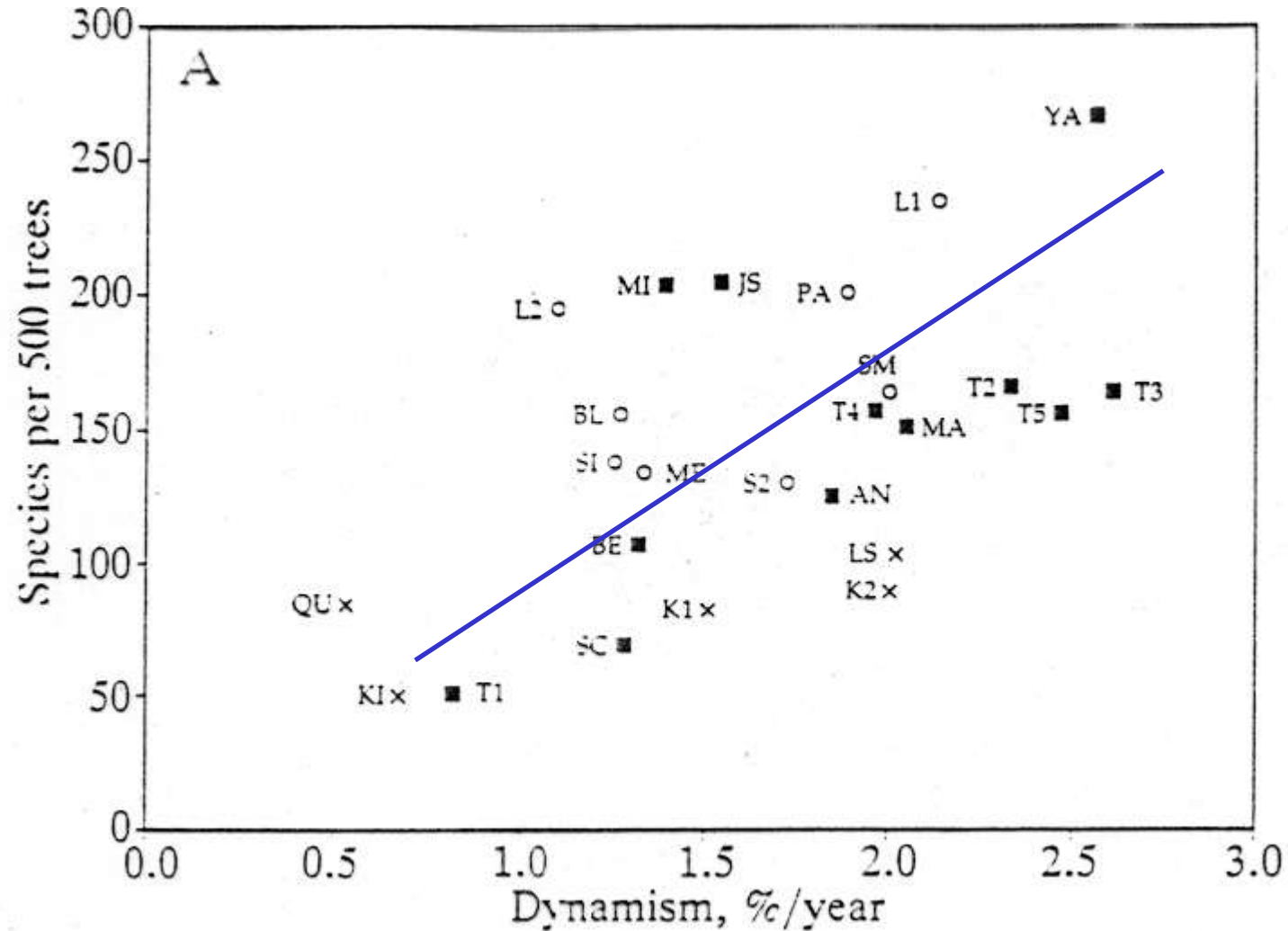
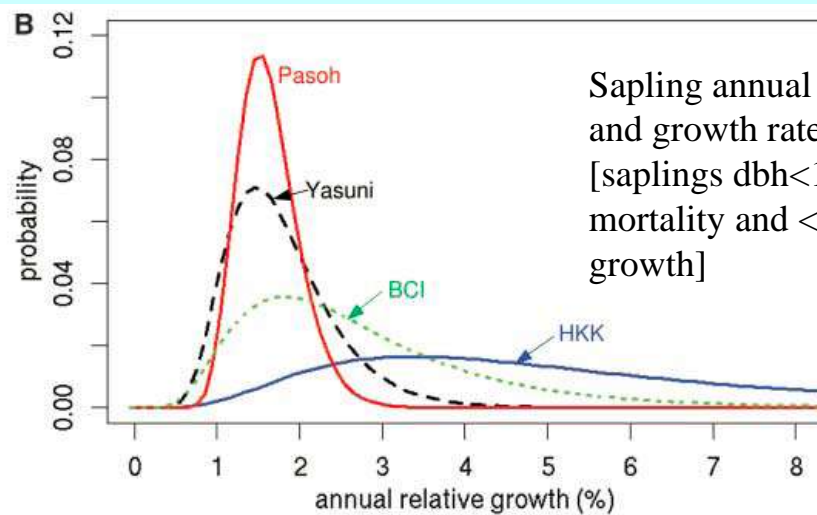
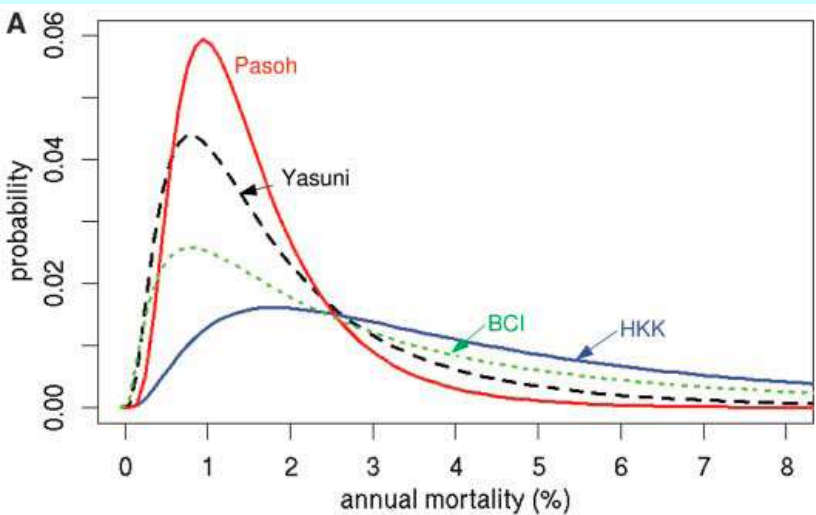
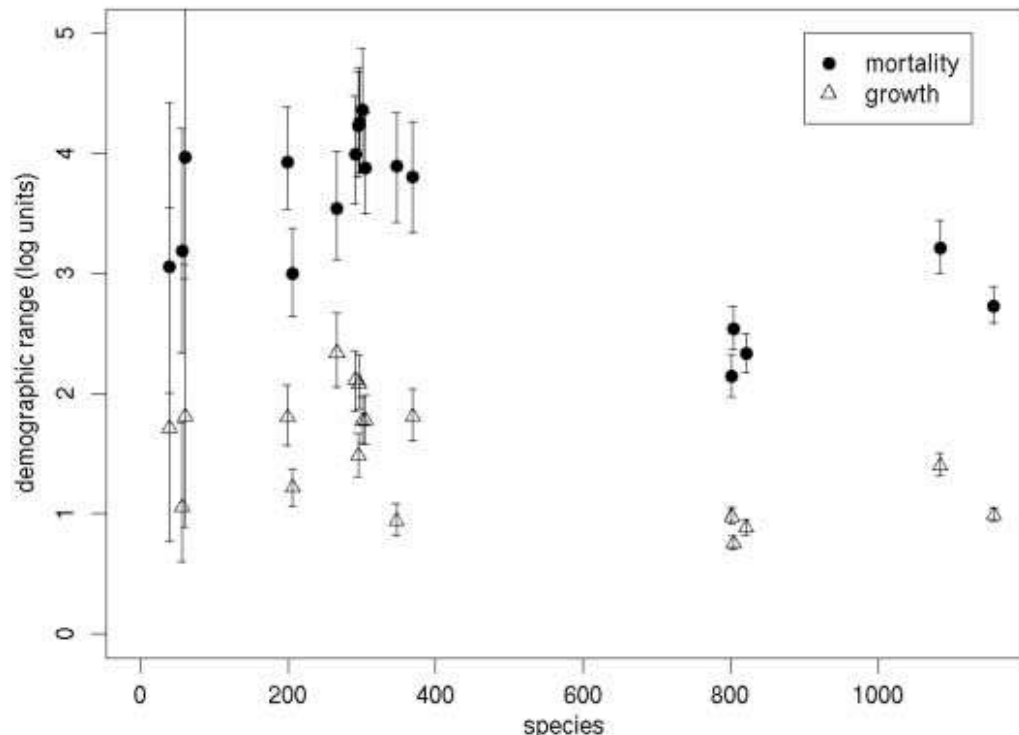


FIG. 1. (A) Tree species richness vs. dynamism (mean of annual mortality and recruitment). (B) Tree species richness predicted by multivariate regression model vs. observed tree species richness. \circ , Sites in South-East Asia; \blacksquare , sites in Amazonia; \times , sites in "Other" phytogeographical regions (Africa, Australia, and Central America). See Table 1 for complete site descriptions, Table 4 for multivariate model.

Demographic differences facilitate species richness: more species would coexist in plots with higher variability in growth and mortality among species – not confirmed in 50-ha plots



Sapling annual mortality and growth rate in 4 forests [saplings dbh<10cm for mortality and <5cm for growth]



Range of sapling demographic rates (mortality, growth) for tree species within a community versus the number of species at the site. The range is the logarithm of the ratio between the 97.5 and 2.5 percentiles of the fitted distributions

Variability in mortality & growth does not matter

What prevents a single "superman" tree species, best adapted to local conditions, from out-competing others and prevailing in the rainforest?

Each species is limited by a different resource
[niche differentiation]

The environment is variable, changing direction of competition
[tree gap dynamics]

Immigration prevents competitive exclusion
[non-equilibrium community composition]

Species do not have opportunity to compete
[recruitment limitation]

Rare species are demographically favoured
[density-dependent mortality, predation and parasitism]

Recommended reviews:

Leigh et al. 2004. Why Do Some Tropical Forests Have So Many Species of Trees? *Biotropica* 36: 447–473

Wright, S. J. 2002. Plant diversity in tropical forests: A review of mechanisms of species coexistence. *Oecologia* 130: 1–14.

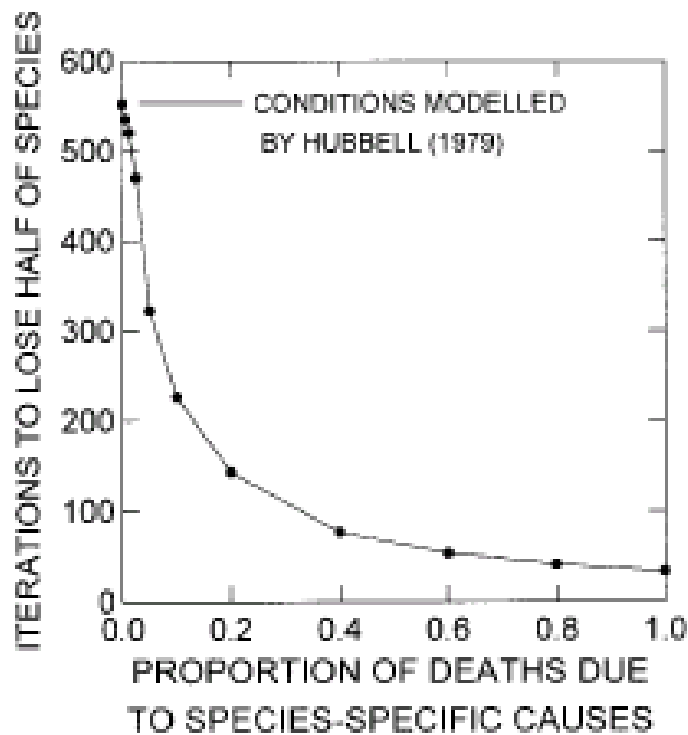


Fig. 4 The transitory maintenance of community diversity by chance (*vertical axis*) decays rapidly as the proportion of deaths determined by chance declines and the proportion determined by species-specific causes increases (*horizontal axis*). Hubbell (1979) simulated the indicated conditions. In contrast, Yu et al. (1998) simulated conditions similar to those at the opposite extreme along the horizontal axis. I used annual probabilities of mortality observed for different tree species on BCI (Condit et al. 1995) for species-specific deaths and equal probabilities for every individual regardless of species for chance deaths and for all births. Following Hubbell (1979), the simulated community included 1,600 trees initially divided among 40 equally abundant species with 160 tree deaths per model iteration

Wright 2001. *Oecologia* 130:1

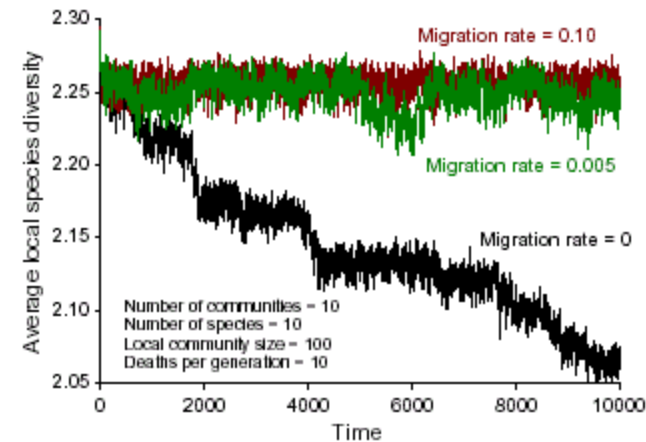


Figure 1. Decline in Shannon–Wiener species diversity in simulated neutral communities composed of ecologically equivalent species. Simulations were run for 10,000 time steps, where 10 death/replacement events constituted a single time step. Initial abundances of species were identical within each of ten local communities. Species diversity was calculated at each time step for each community and then averaged across communities. Higher migration rates slow the decline in species diversity.

**Mass effect
(immigration)
slows down loss of
species from the
community due to
competition**

What prevents a single "superman" tree species, best adapted to local conditions, from out-competing others and prevailing in the rainforest?

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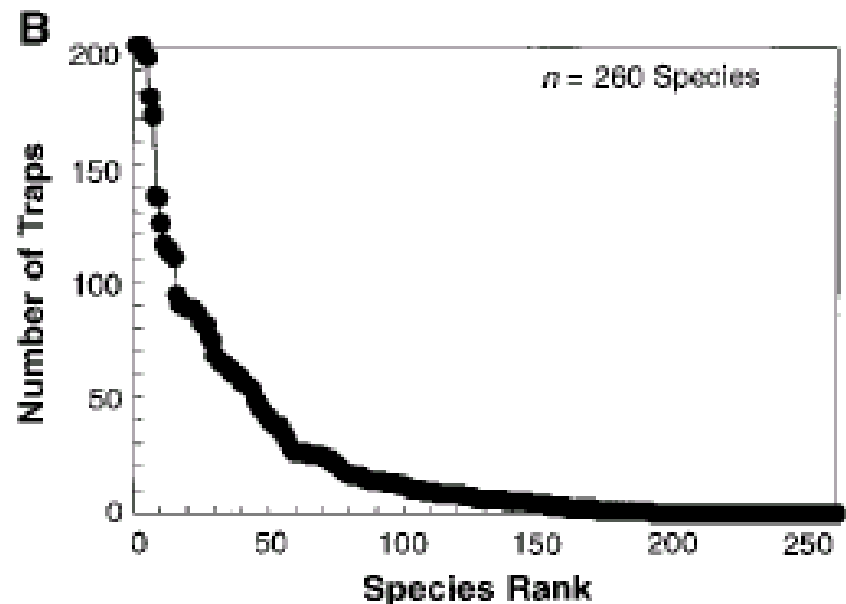
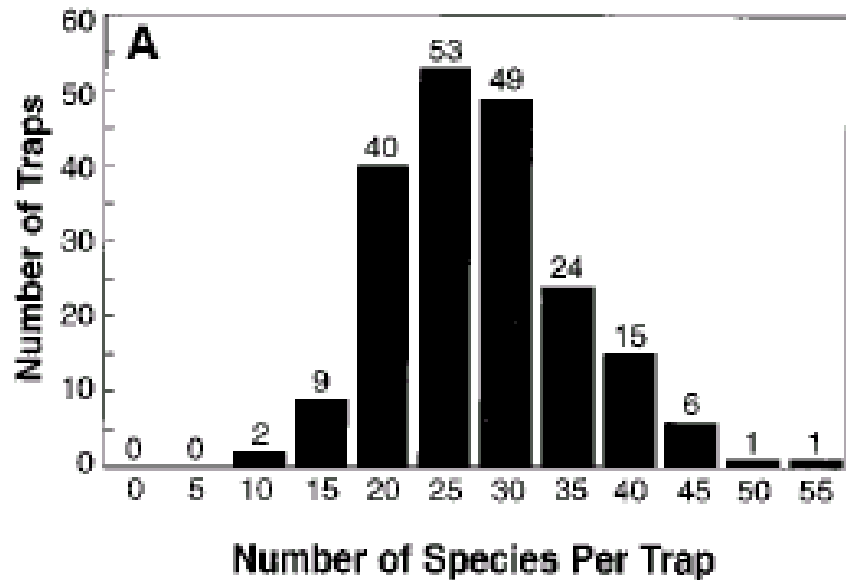
Species do not have opportunity to compete
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Wright, S. J. 2002. Plant diversity in tropical forests: A review of mechanisms of species coexistence. *Oecologia* 130: 1–14.



Seed traps at BCI:

evidence for dispersal limitation,
or most of the tree species does
not get to most of the places most
of the time

Fig. 4. Evidence for dispersal limitation in BCI trees from a 10-year seed trap study using 200 traps in the 50-ha plot. Seeds of a total of 260 species of the 314 species in the plot census were collected at least once. **(A)** Frequency distribution of the number of species captured per trap during the 10-year trapping period (1987–96). The average number of species per trap was 30.8 ± 7.5 SD. **(B)** The total number of traps into which each species dispersed at least one seed during the 10-year trapping period.

We also thank

O. Calderón for counting and identifying $>10^6$ seeds in the seed traps over the years.

20x20m plot in Wanang DBH>1cm

220 stems

85 species

40 singletons

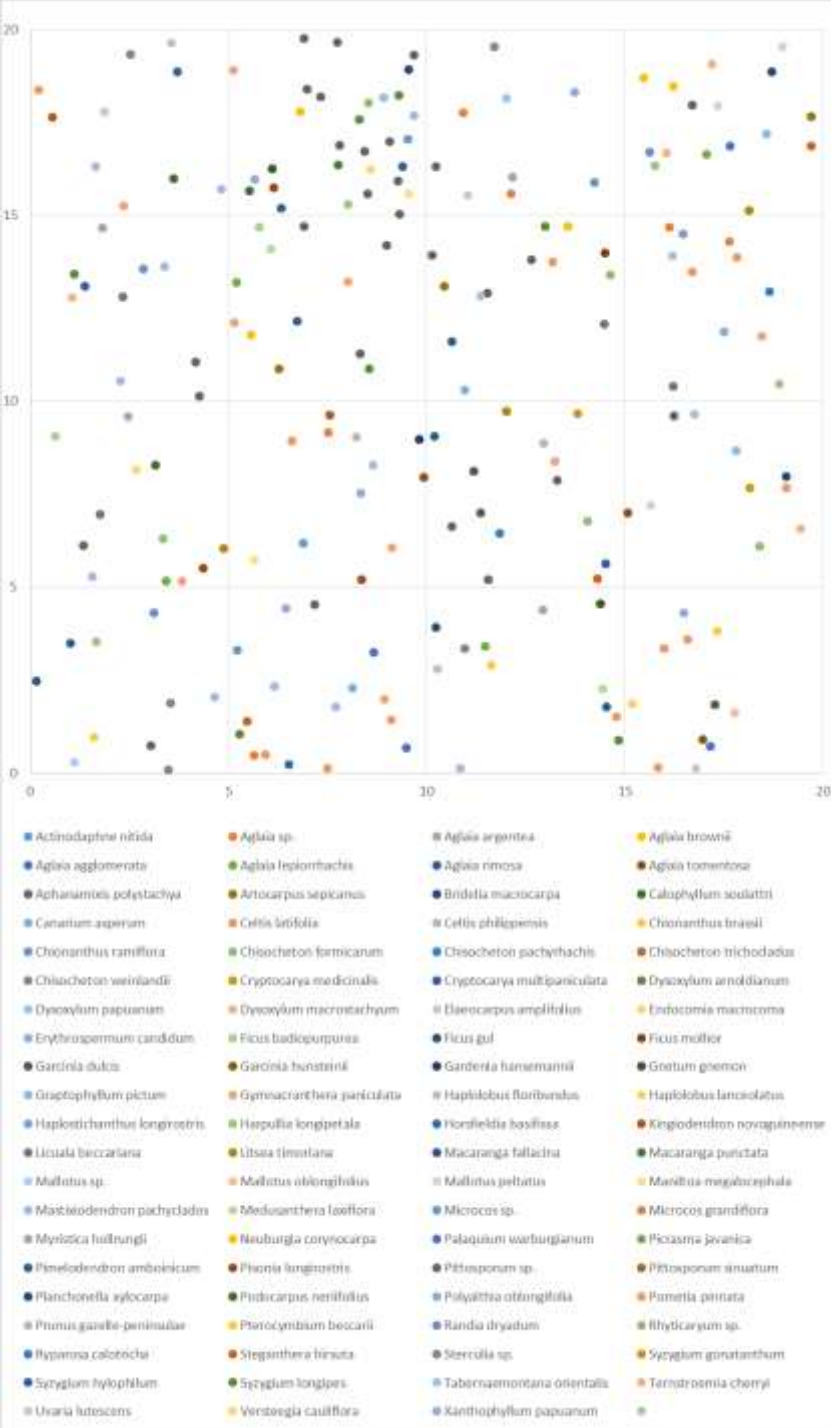
Most common species:

16 stems (7.3%)

Aphanamixis polystachia

85 spp from the species pool of 536
spp (15.9%)

3,570 locally possible pair-wise
interactions from the total of 143,380
(2.5%)

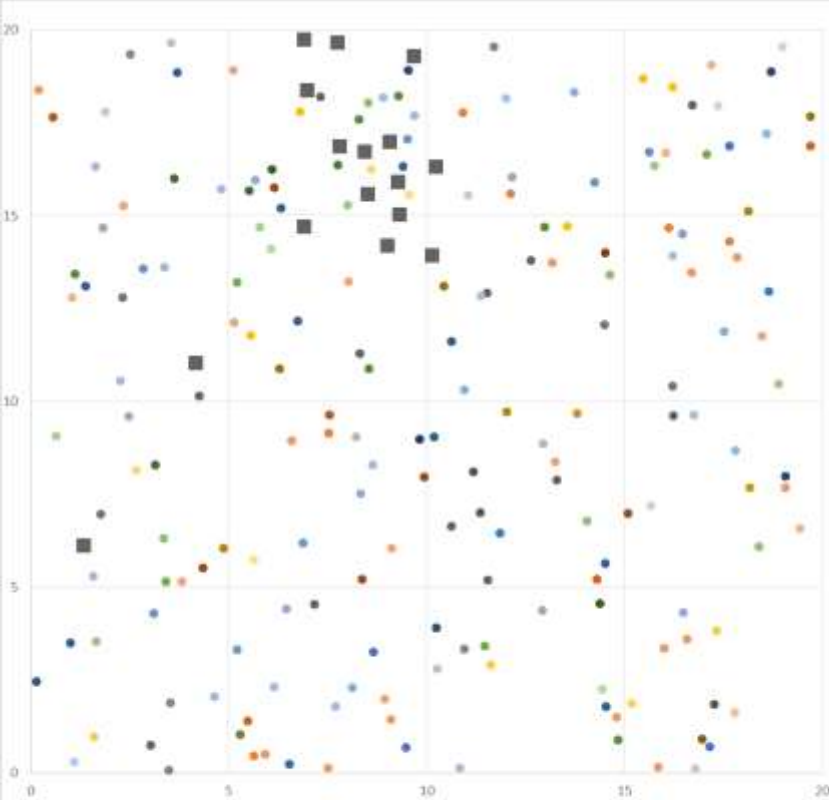


Aphanamixis polystachia

(Meliaceae): the most common tree sp.

Nearest neighbours:

4 conspecific trees and 9 other spp



| | |
|-----------------------------|---|
| Aphanamixis polystachia | 4 |
| Cryptocarya multipaniculata | 2 |
| Manitosa megaloccephala | 2 |
| Pittosporum sp. | 2 |
| Actinodaphne nitida | 1 |
| Artocarpus sepicanus | 1 |
| Harpulia longipetala | 1 |
| Macaranga punctata | 1 |
| Mastixiodendron pachycladus | 1 |
| Planchonella xylocarpa | 1 |



What prevents a single "superman" tree species, best adapted to local conditions, from out-competing others and prevailing in the rainforest?

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Negative density-dependency: four mechanisms

- Intra-specific self-thinning
- Inter-specific competition
- Mortality due to density-responsive pests/pathogens
- Mortality due to distance-responsive pests/pathogens



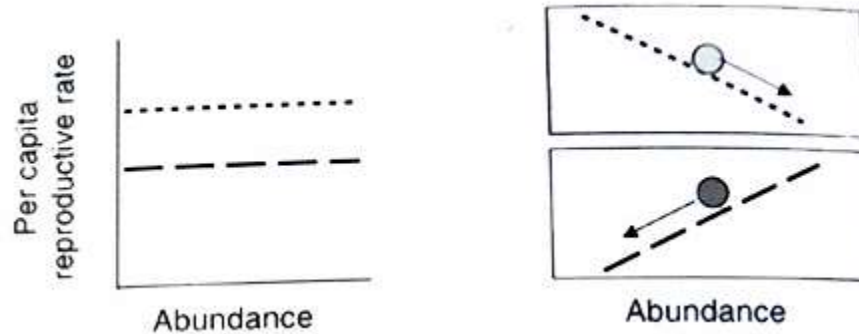
Yoda's power law

mean plant biomass = $c * \text{density}^{-3/2}$

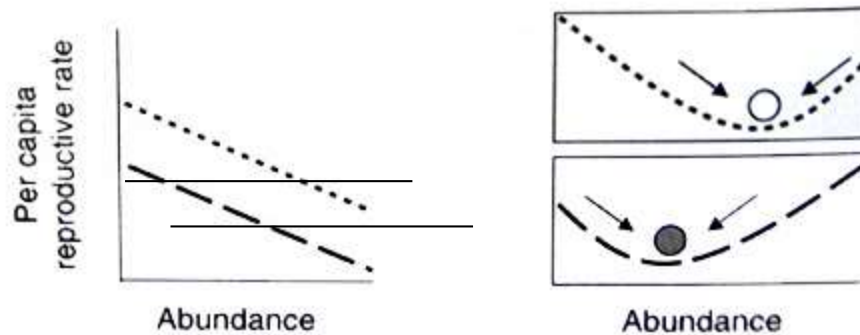
$W^{2/3} * N = \text{constant}$



Density-dependence of reproductive rate [r] and species coexistence



two species differ in their r , which is density independent
species with higher r competitively excludes the other species



two species differ in their r , which is density dependent
two species coexist at densities where they have identical value of r

$$N_{t+1} = r * N_t^{(1 - b)}$$

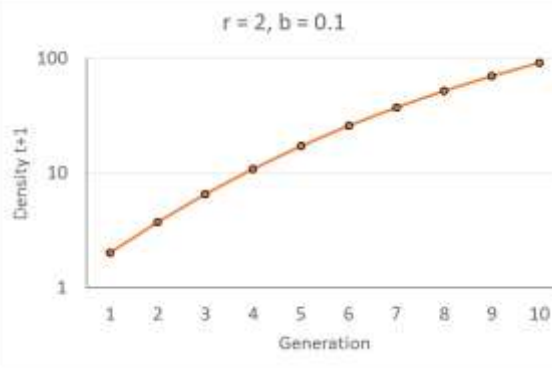
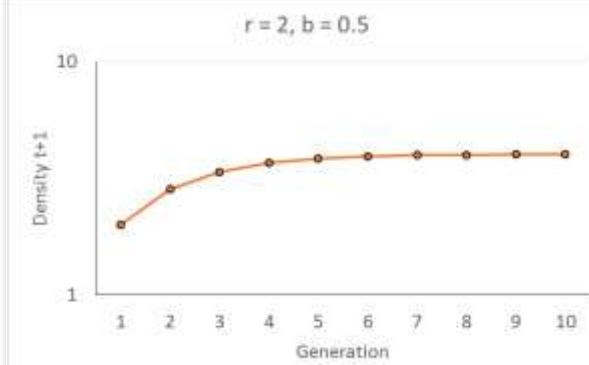
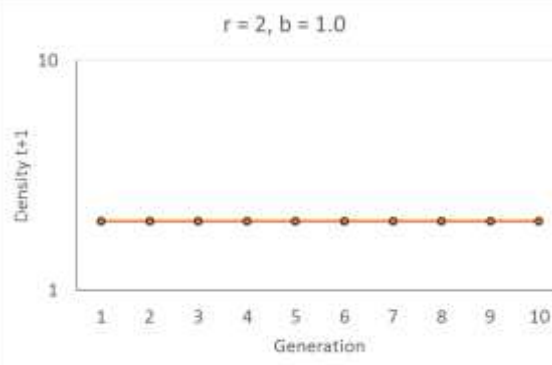
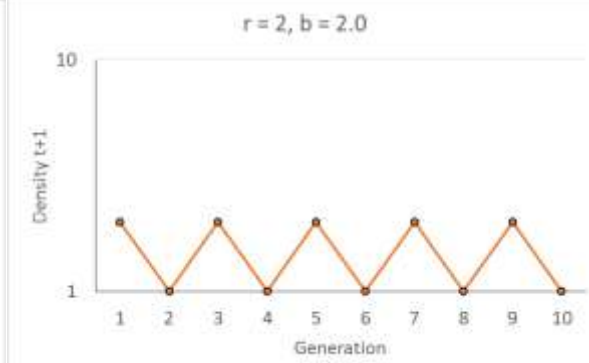
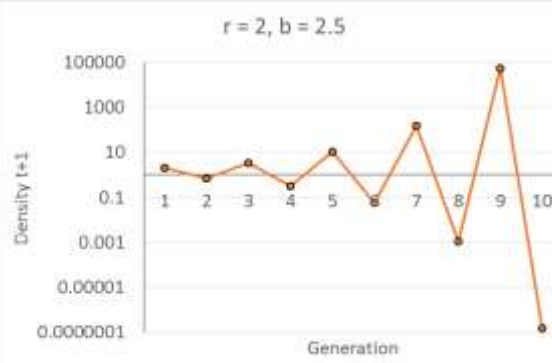
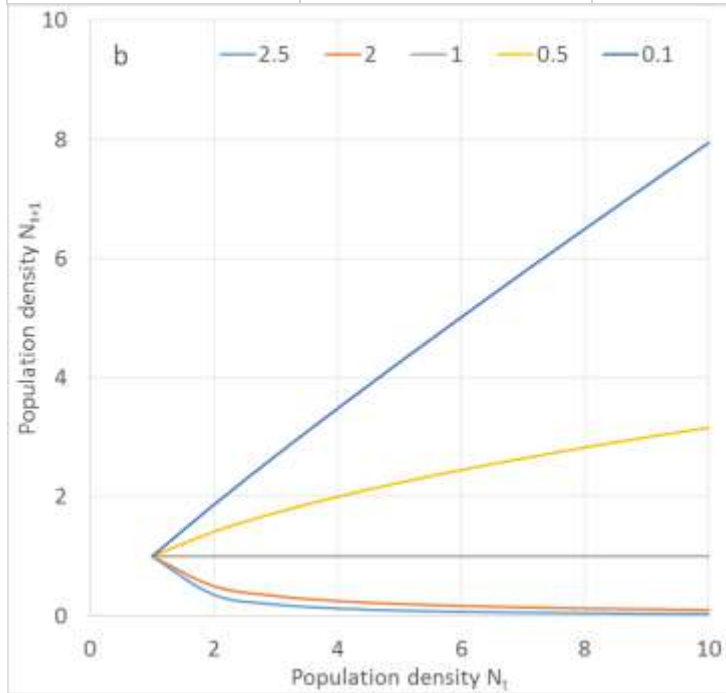
N_{t+1} population density next generation

N_t population density initial [0,1]

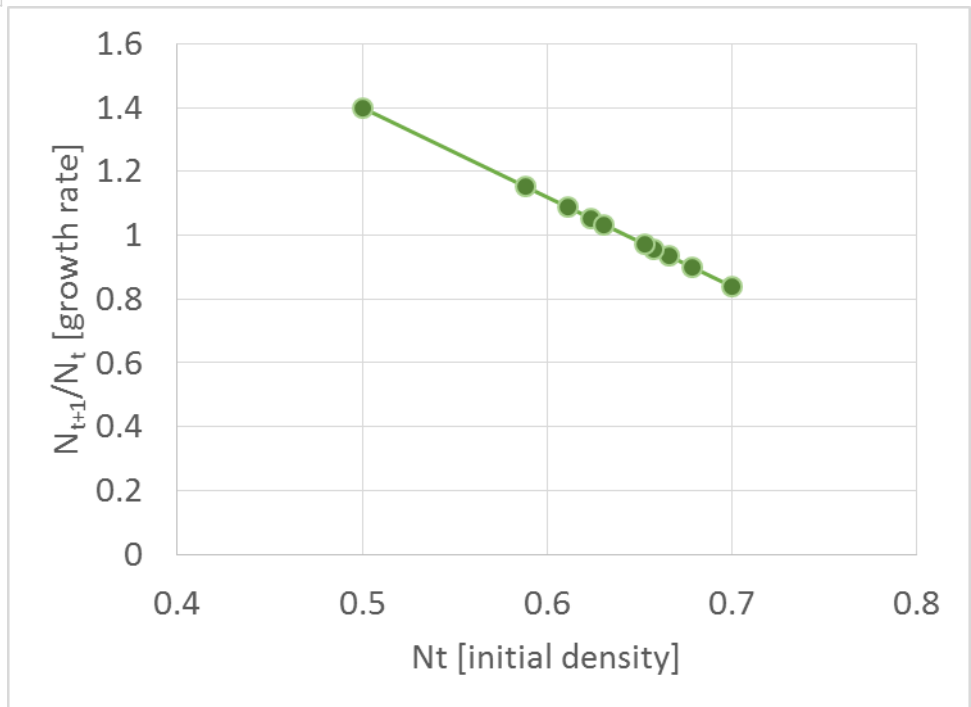
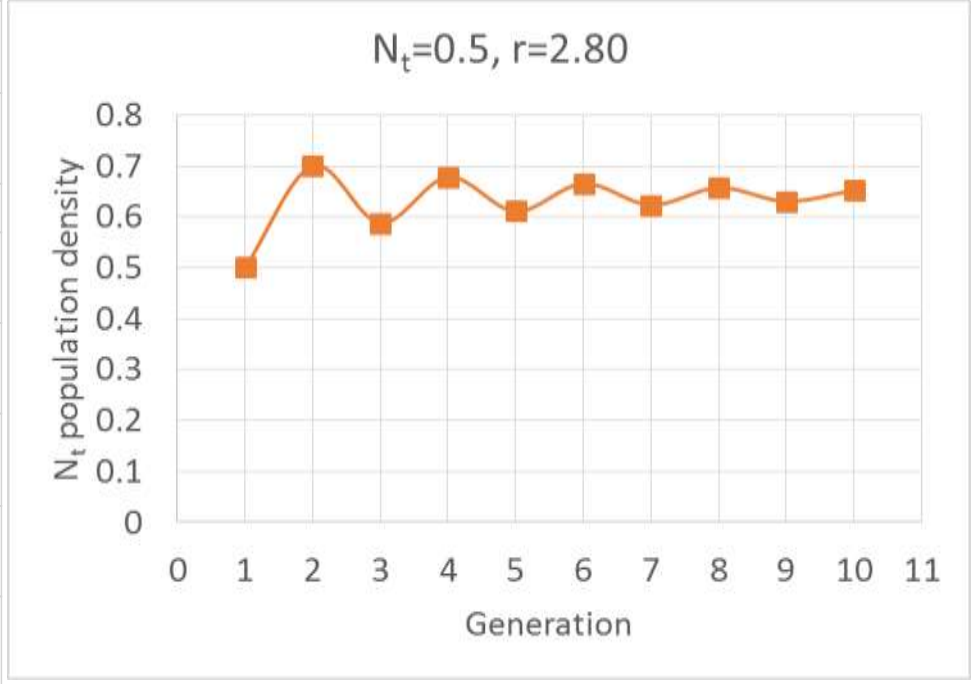
r growth rate

b density dependent limit on growth

$2 > b > 0$ stable population



| | |
|---------------------------------|------------------------------------|
| $N_{t+1} = rN_t - rN_t^2$ | |
| $N_{t+1} = r * N_t * (1 - N_t)$ | |
| N_{t+1} | population density next generation |
| N_t | population density initial [0,1] |
| r | growth rate |
| rN_t | exponential growth |
| $-rN_t^2$ | density dependent limit on growth |



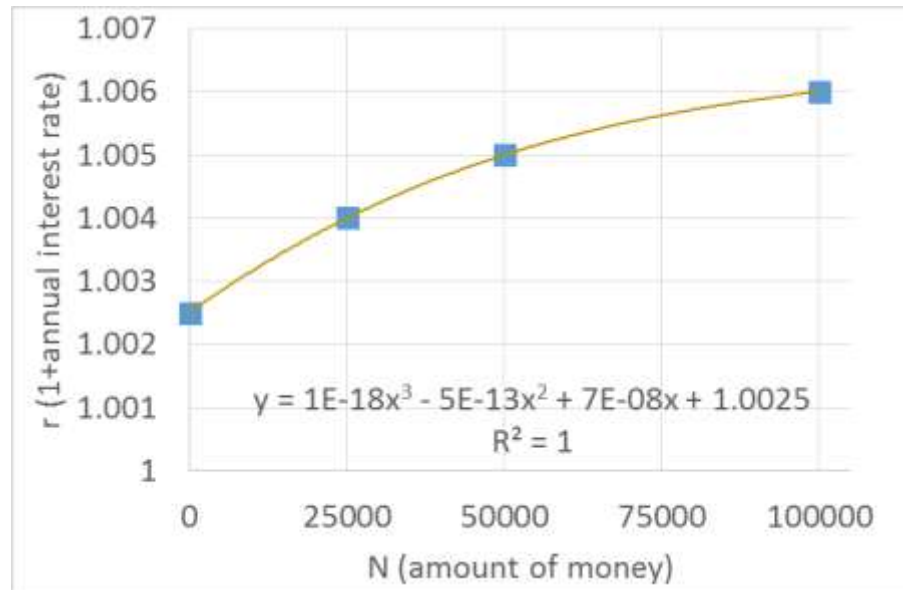


The roots of social inequality problem:
the dynamics of money is
positively density dependent



Current interest rates

| Balance | Gross pa% |
|-------------------|-----------|
| £1 - £24,999 | 0.25 |
| £25,000 - £49,999 | 0.40 |
| £50,000 - £99,999 | 0.50 |
| £100,000 + | 0.60 |



$$N_{t+1} = N_t * (r_1 + 10^{-18} * N^3 - 5 * 10^{-13} * N^2 + 7 * 10^{-8} * N)$$

Distribution of Wealth in the U.S., 2001

Top 1% own 33%
Next 4% own 26%
Next 5% own 12%

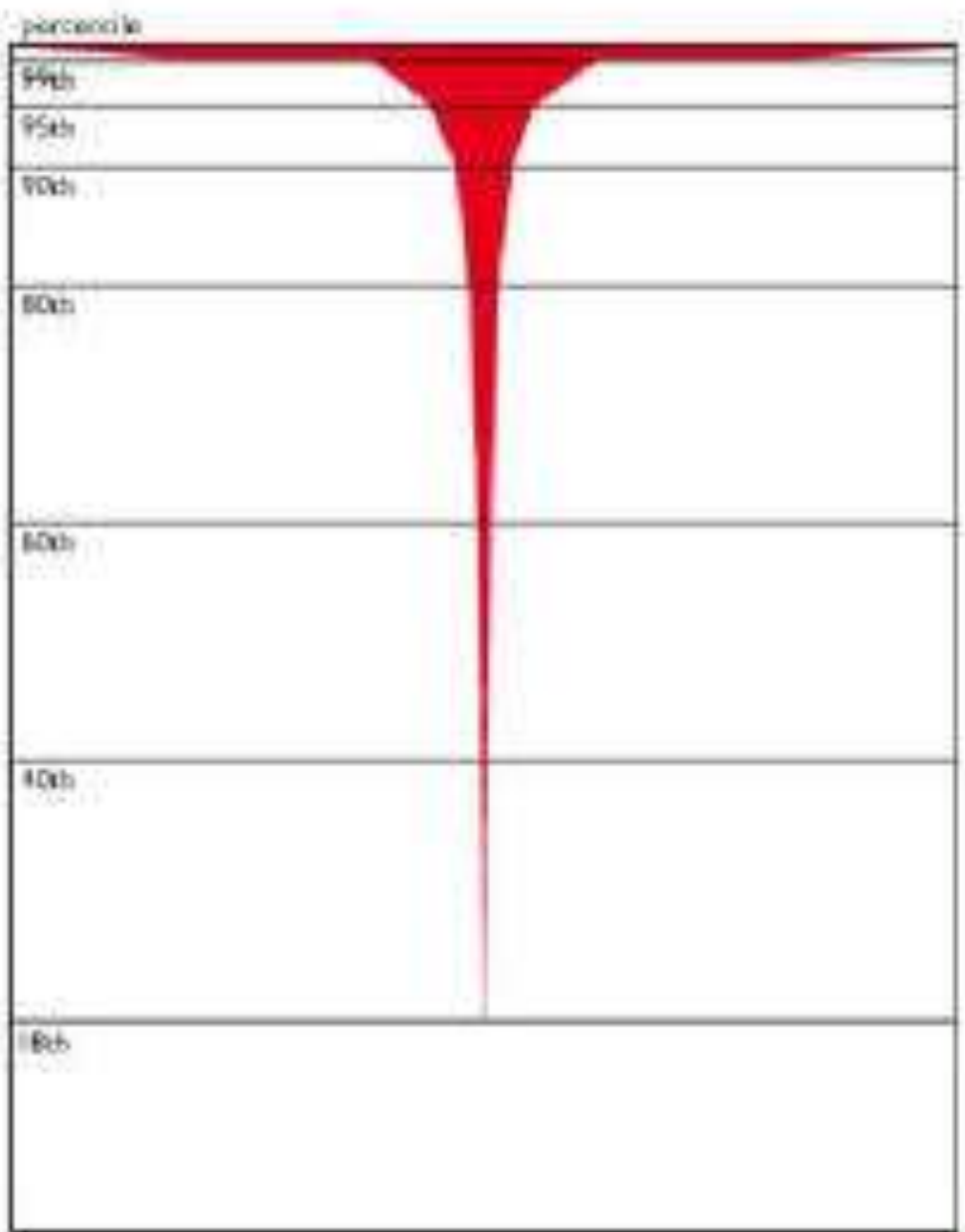
Next 10% own 13%

Next 20% own 11%

Middle 20% own 4%

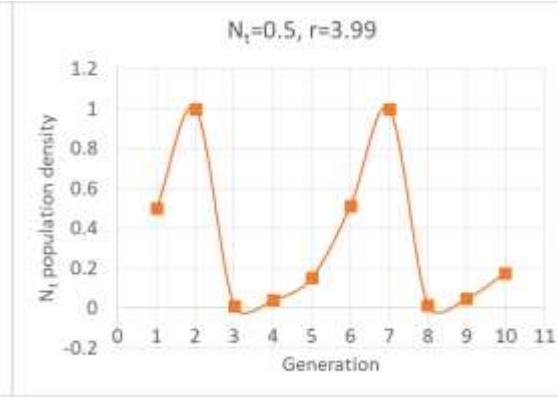
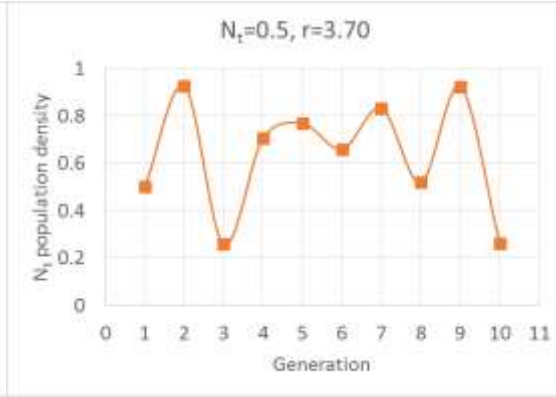
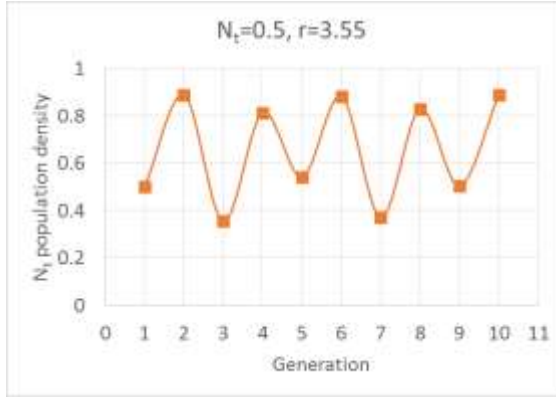
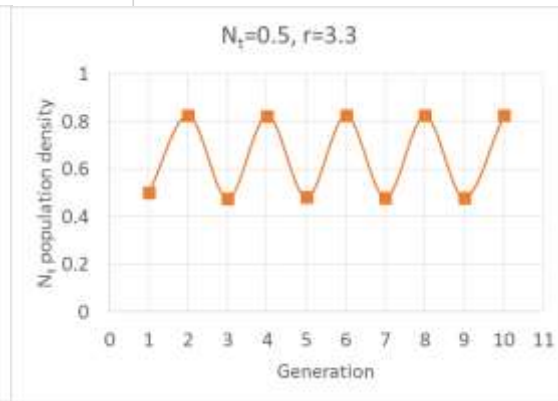
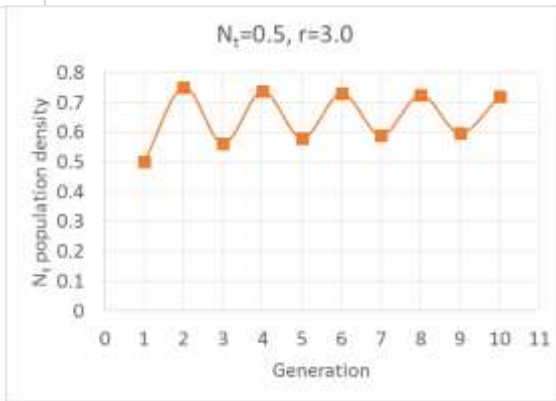
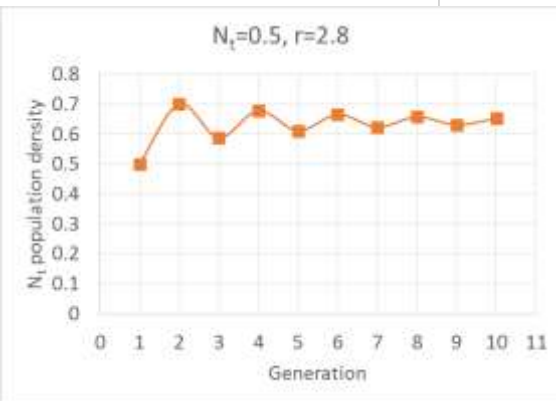
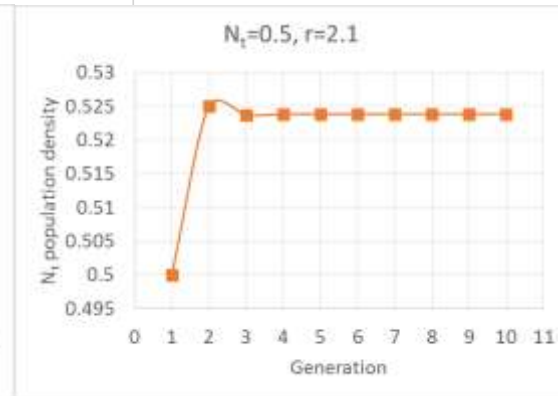
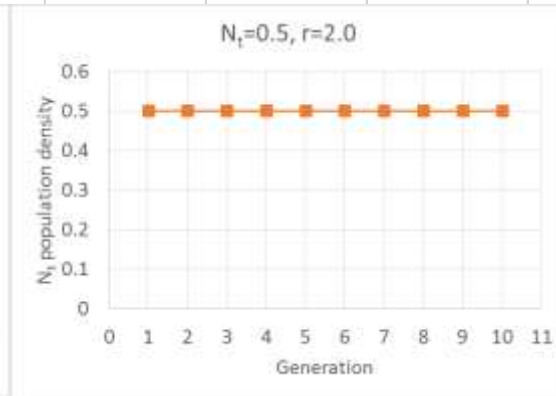
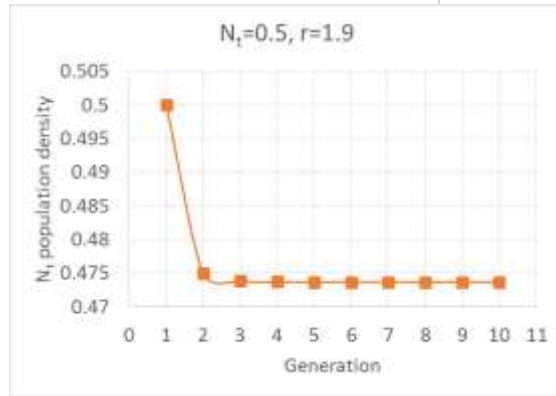
Next 22% own 0.3%

Bottom 18% have zero or negative net worth



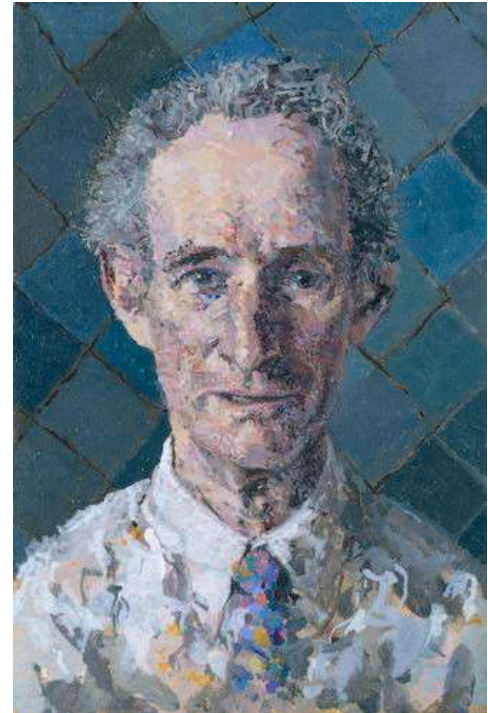
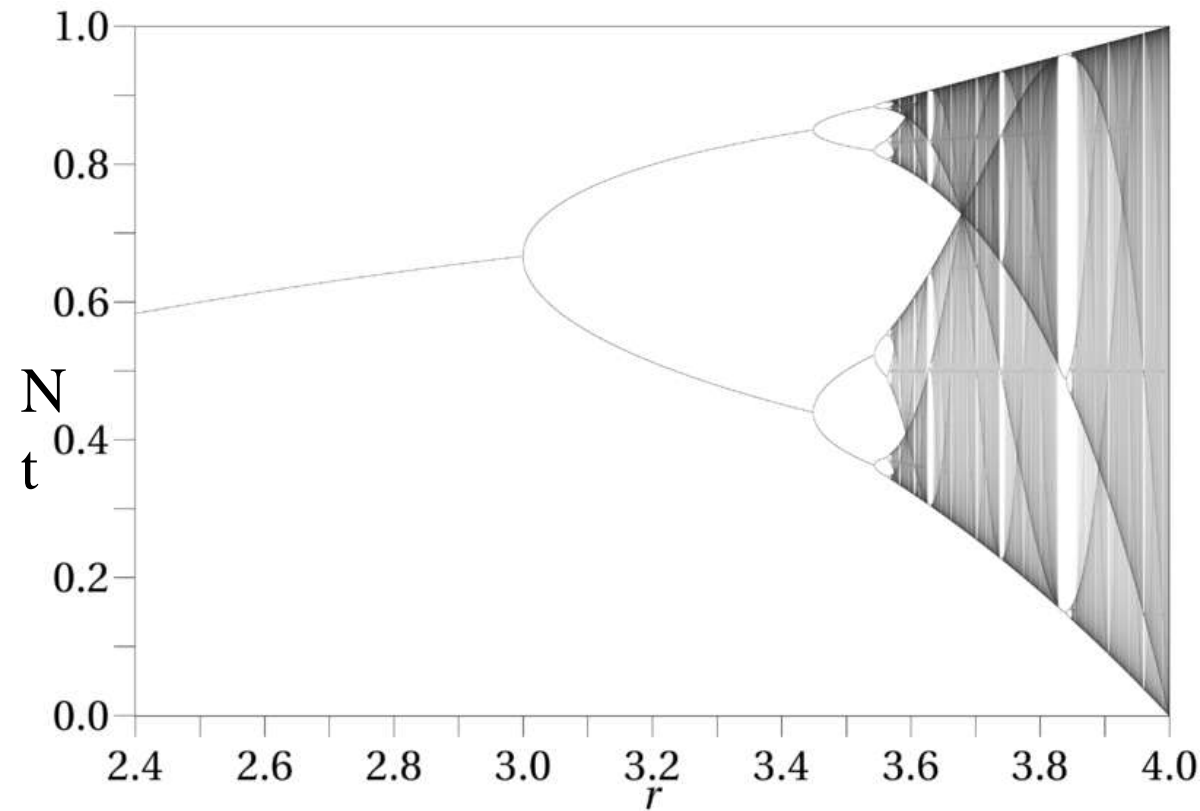
$$N_{t+1} = rN_t - rN_t^2$$

$$N_{t+1} = r * N_t * (1 - N_t)$$

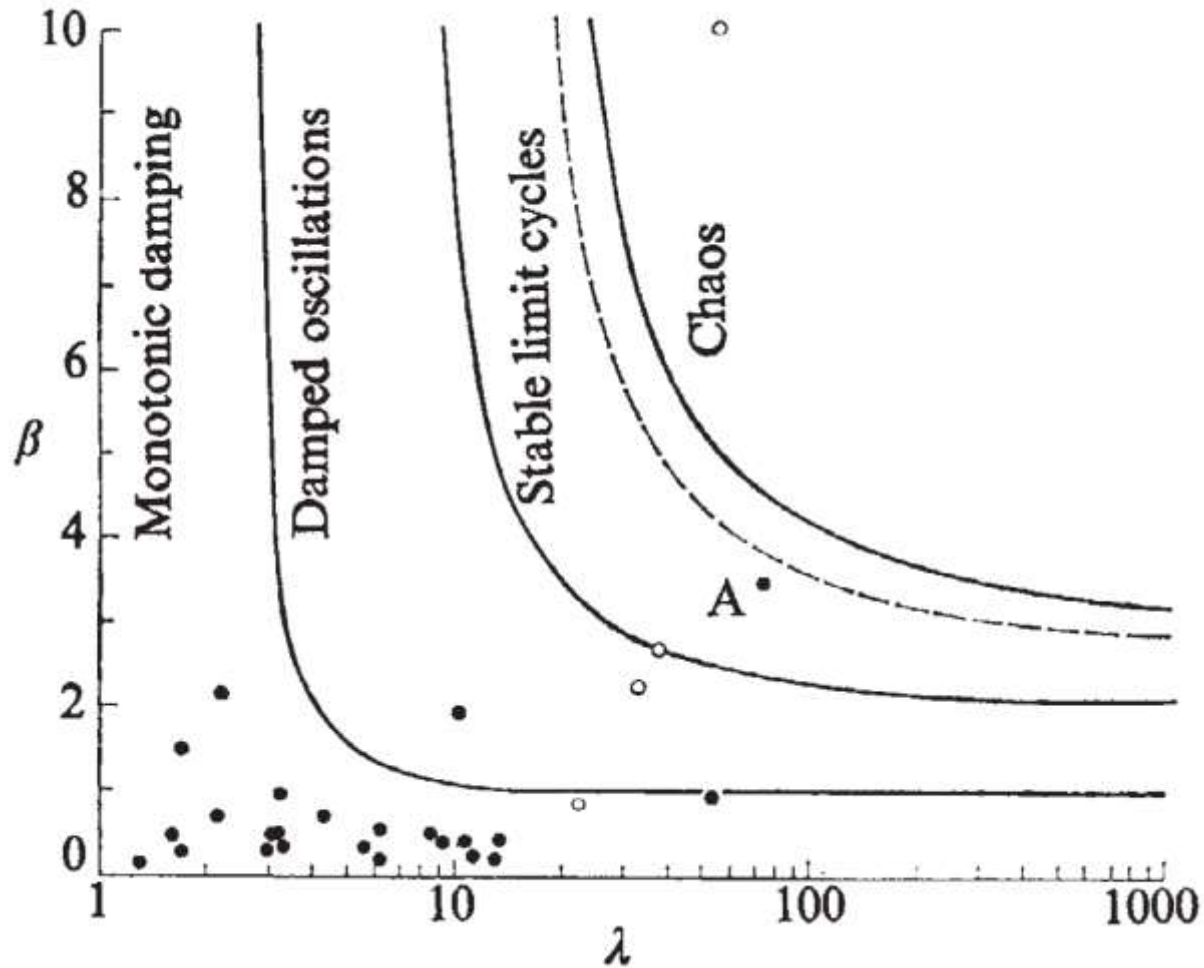


| | | | |
|---------------------------------|------------------------------------|--|--|
| $N_{t+1} = rN_t - rN_t^2$ | | | |
| $N_{t+1} = r * N_t * (1 - N_t)$ | | | |
| N_{t+1} | population density next generation | | |
| N_t | population density initial [0,1] | | |
| r | growth rate | | |
| rN_t | exponential growth | | |
| $-rN_t^2$ | density dependent limit on growth | | |

Simple mathematical models with very complicated dynamics
 RM May Nature 261 (5560), 459-467

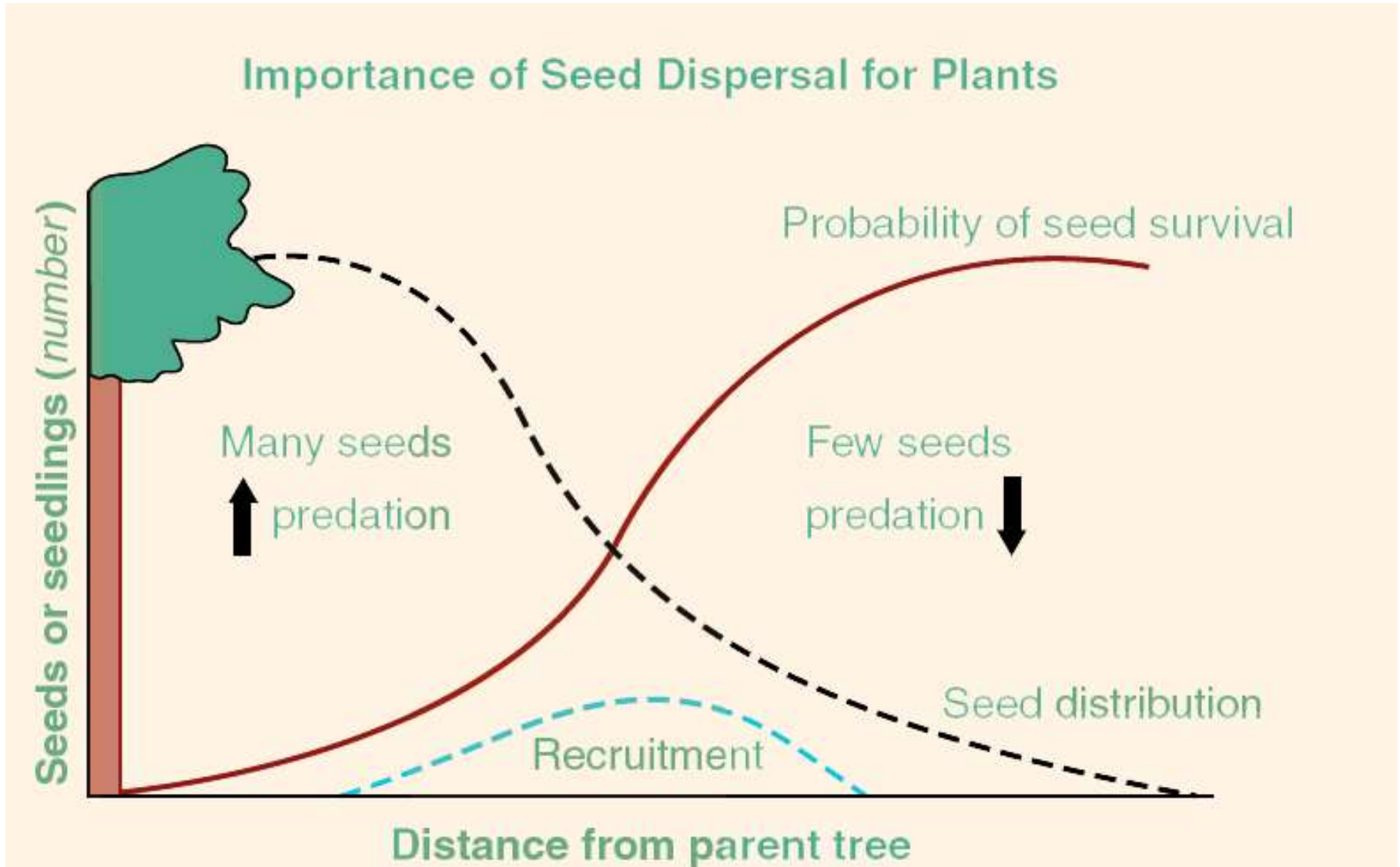


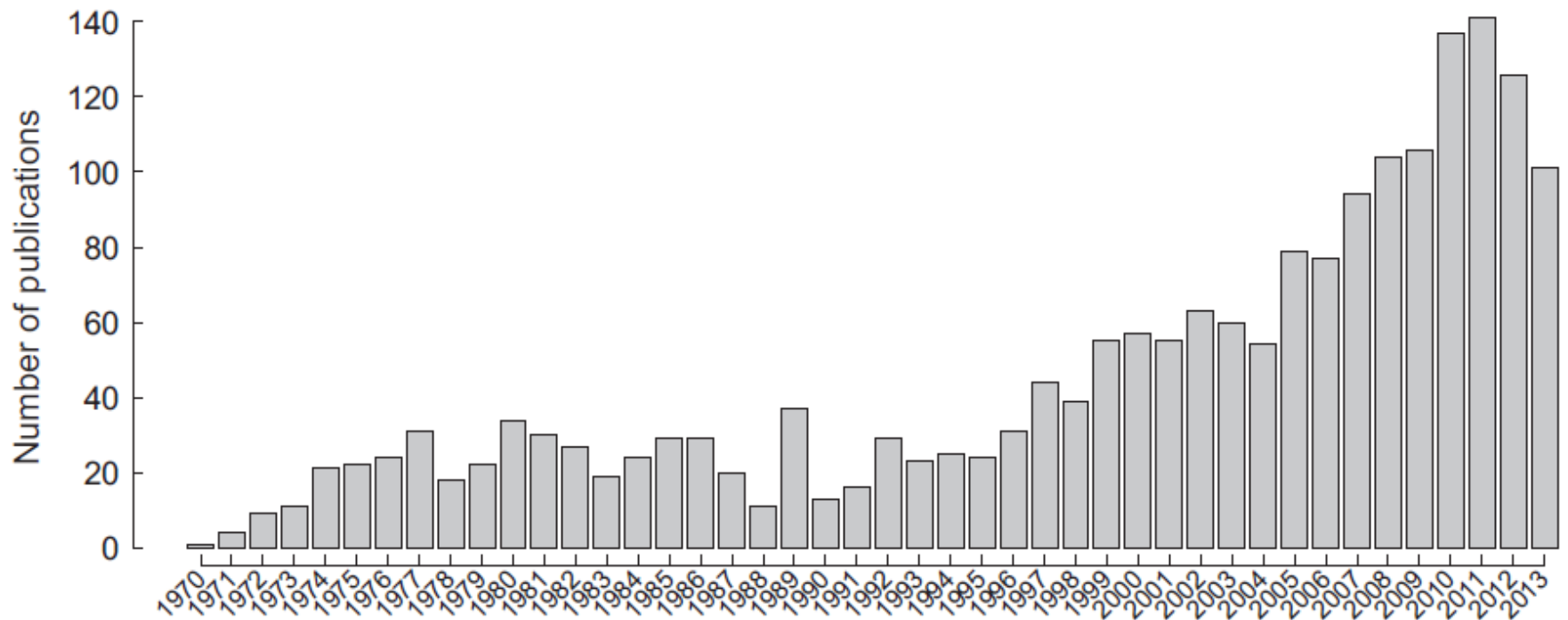
$$X_{t+1} = \lambda X_t (1 + X_t)^{-\beta}$$



Rare species are favoured demographically

The Janzen-Connell hypothesis: seedlings are more likely to be dispersed near their parent, but more likely to survive far from their parent.





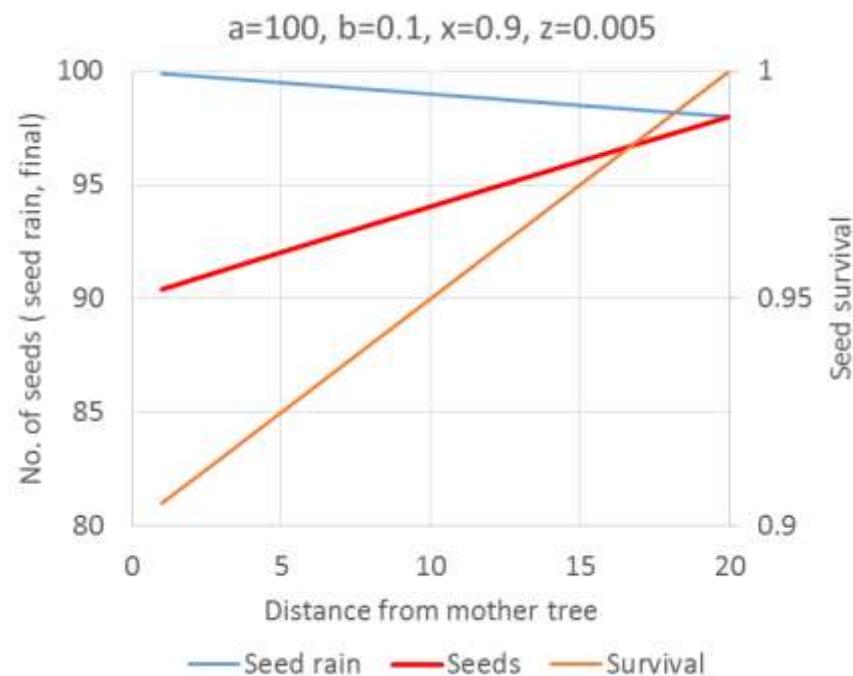
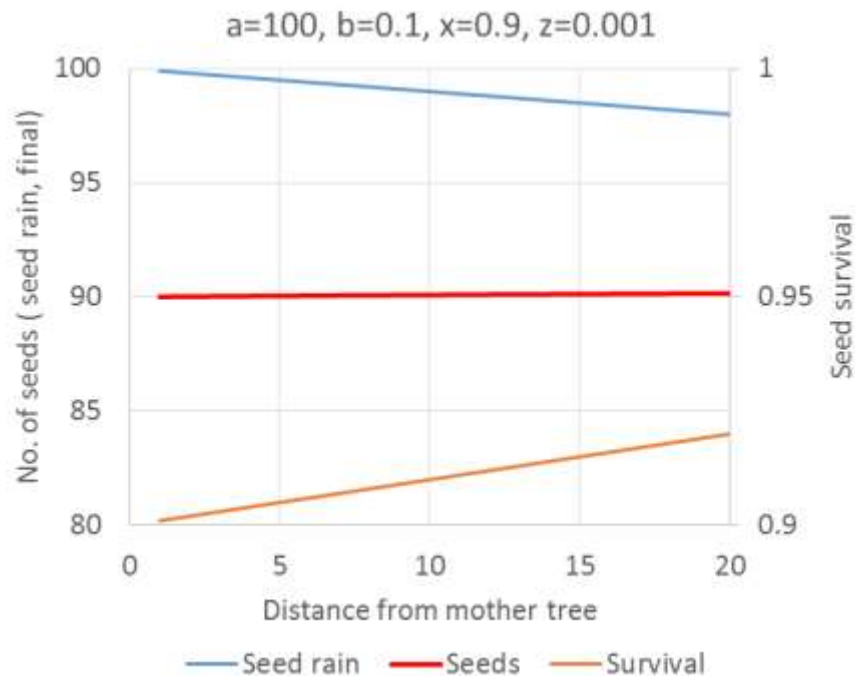
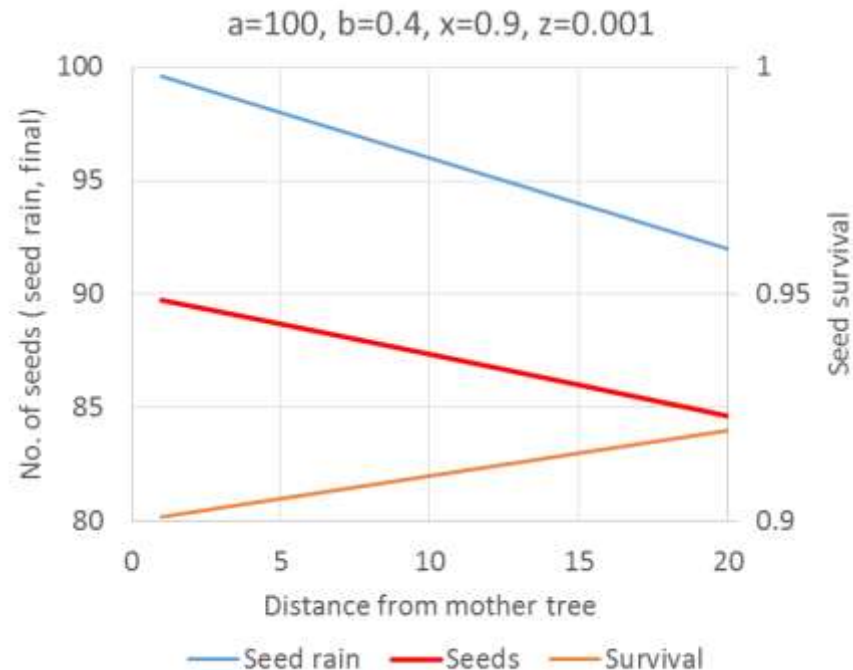
Number of articles citing Janzen (1970) or Connell (1971) between 1970 and 2013 (total = 1976).

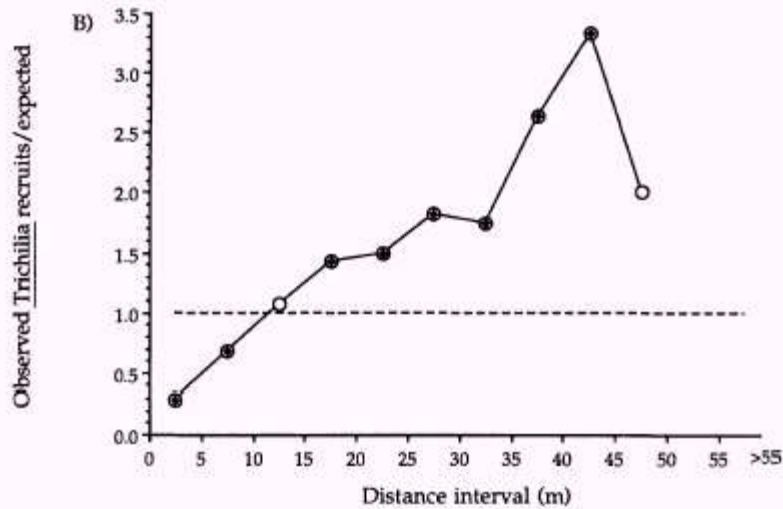
Importance of Seed Dispersal for Plants



$$\text{seed rain} = a - b * \text{distance}$$

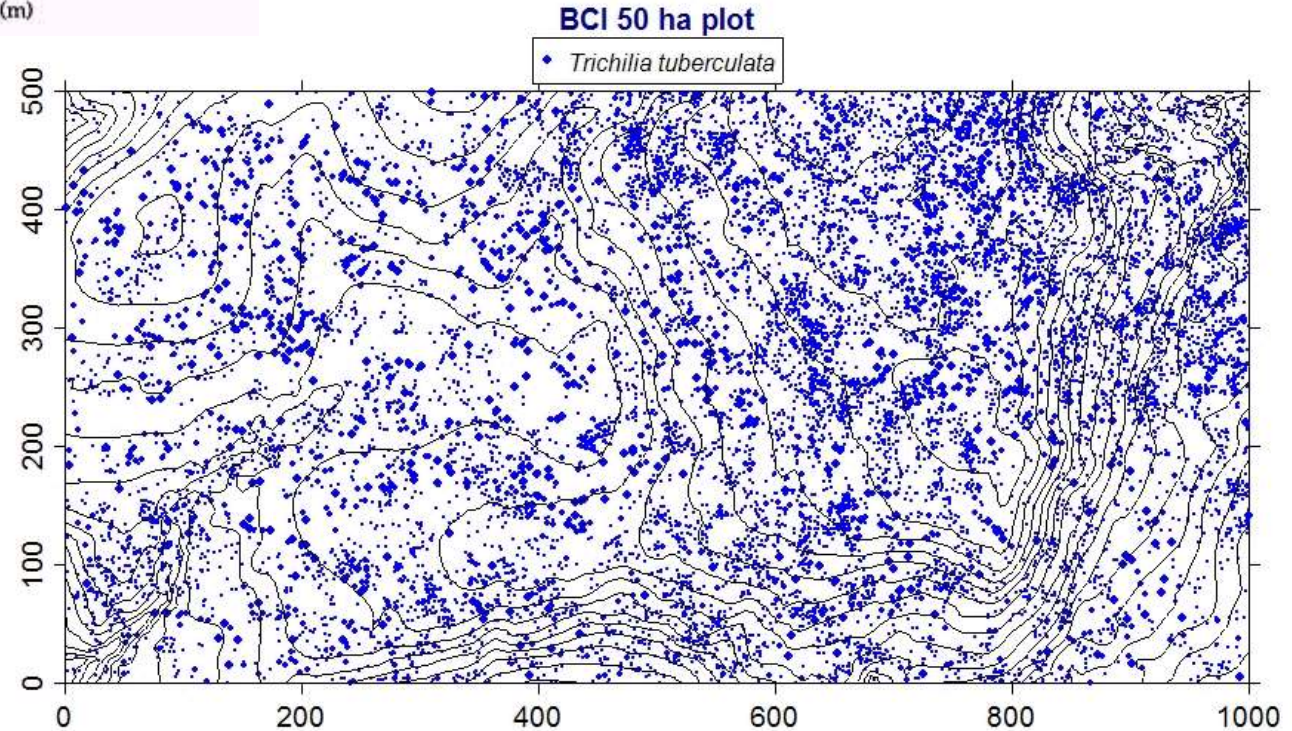
$$\text{survival} = x + z * \text{distance}$$



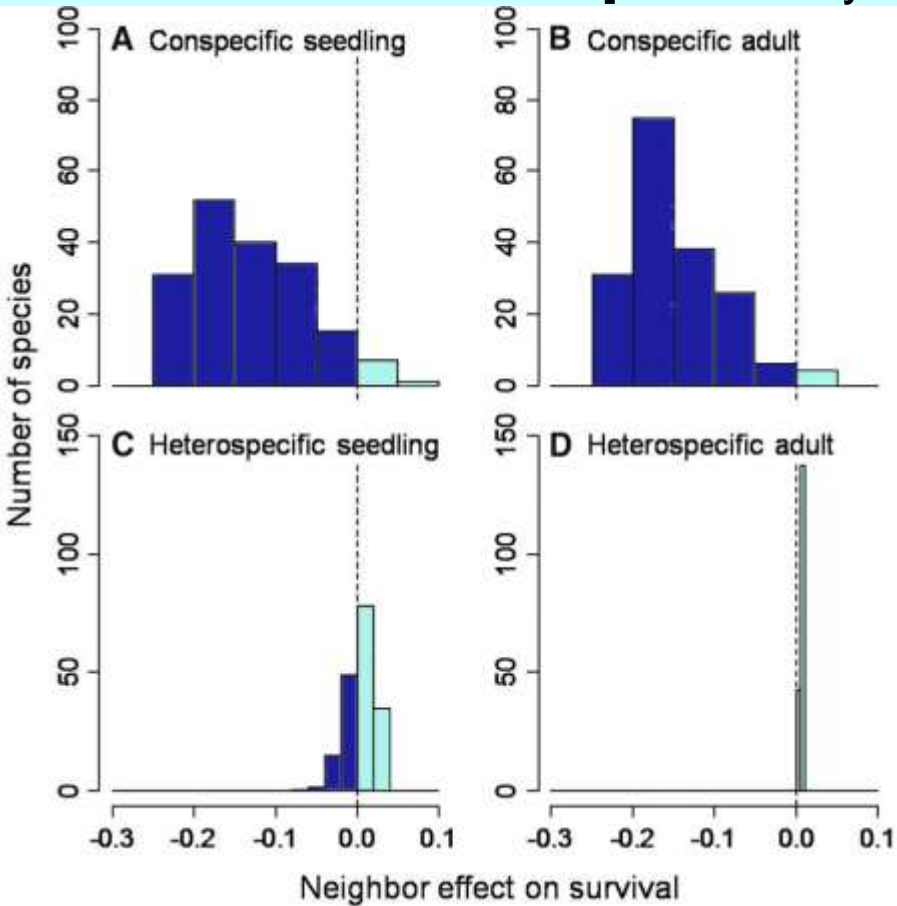


Density dependency:
recruitment increases
with distance from mother tree

Trichilia tuberculata

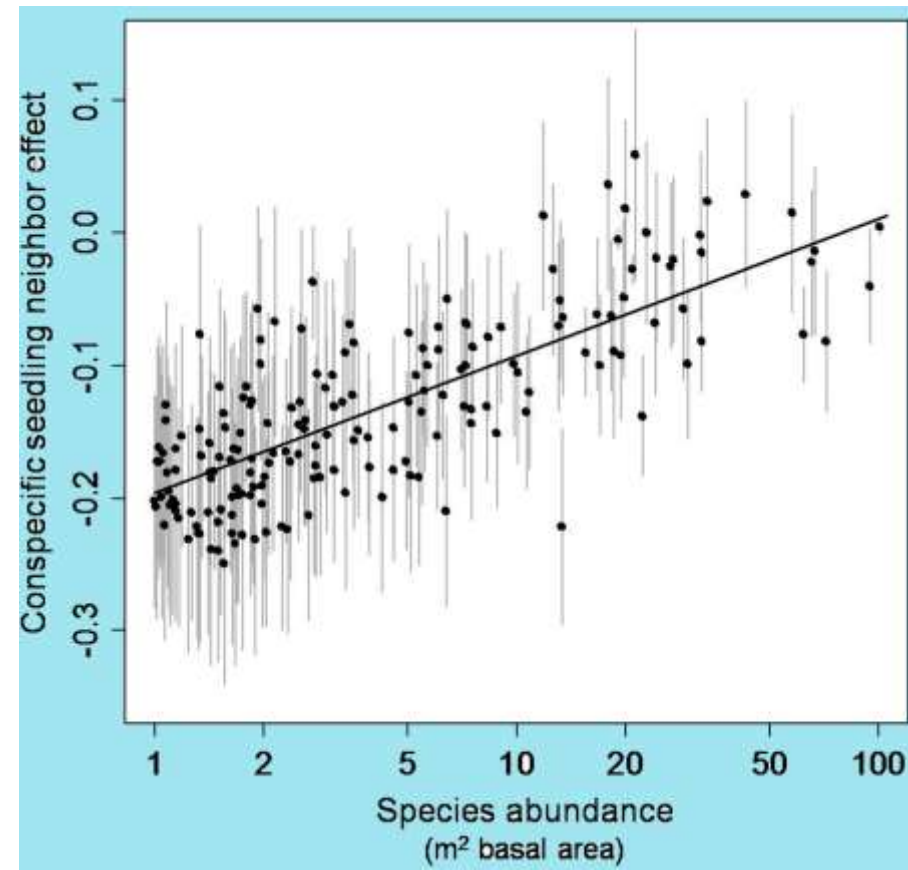


The effect of a conspecific neighbour on seedling survival is **negative** and **varies among species according to their abundance**: rare species suffer more their conspecific neighbours than common species [which may explain their rarity]



Liza S. Comita, et al. Asymmetric Density Dependence Shapes Species Abundances in a Tropical Tree Community. *Science* 329:330, 2010

Neighbour effect on survival by conspecific seedling for tree species on BCI is correlated with their abundance



Effects of conspecific and heterospecific neighbors on seedling survival. [BCI in Panama]. Bars to the left of the dashed zero line indicate species whose survival is reduced by increasing density of neighbors.

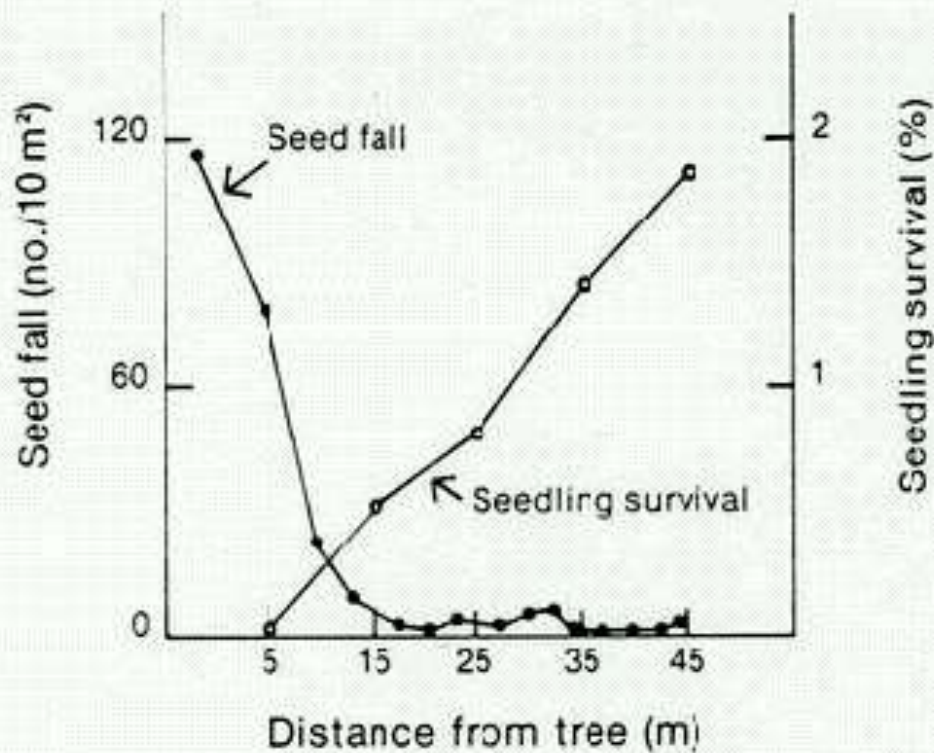
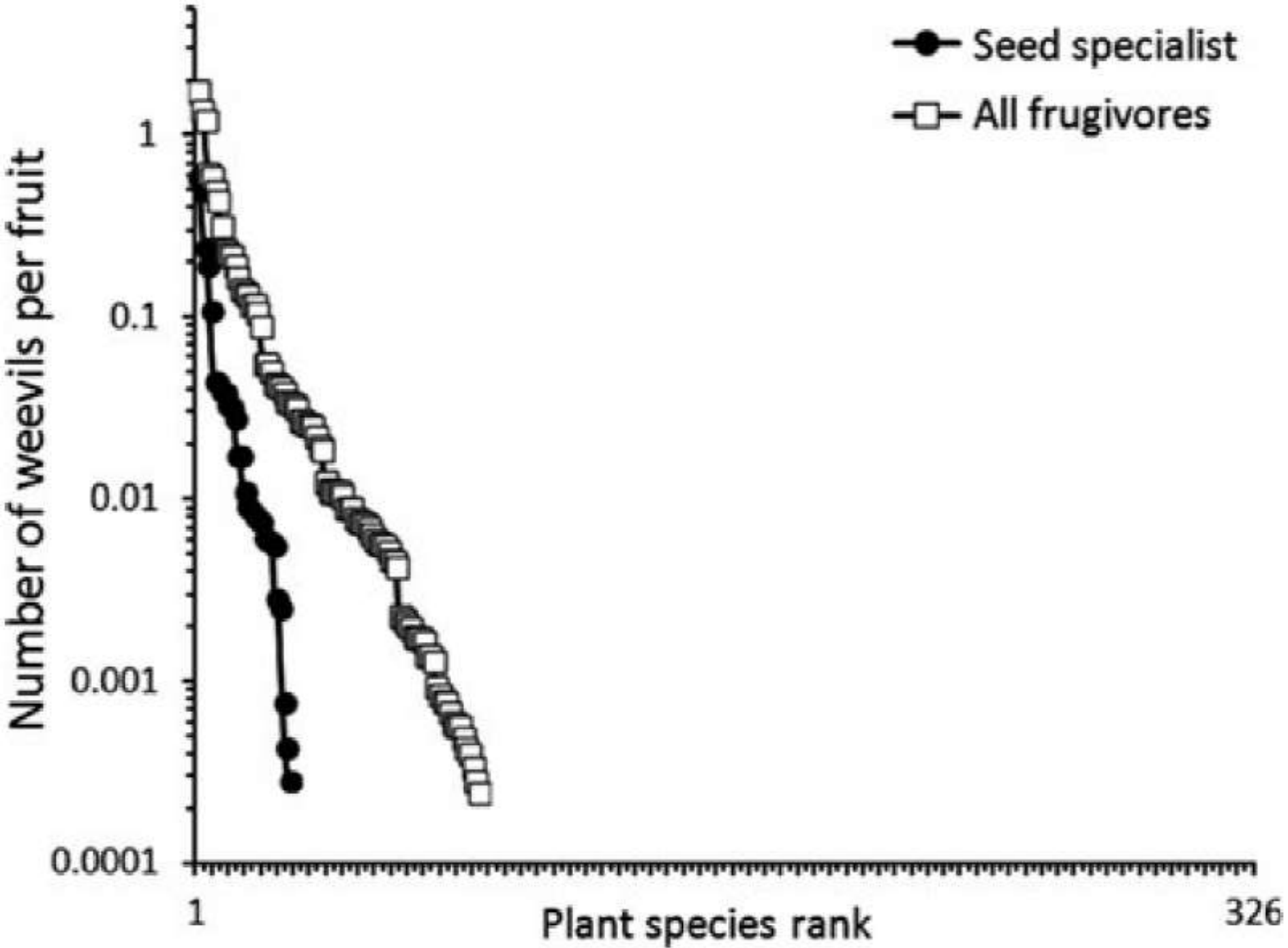


Fig. 7-16 Actual patterns of seed fall and seed and seedling survival three months after fruit drop in the toucan dispersed tree, *Virola surinamensis*, in Panama. Seed and seedling mortality due to *Conotrachelus* weevil infestations are so heavy under the crown that the seed fall and seedling survival curves cross at the crown edge, 10–15 m from the tree base. Not surprisingly, adults of this species are clumped, with an average nearest neighbor distance of 18 m. Overall, there is a 40-fold advantage to seed dispersal only 45 m from fruiting *Virola* trees. Data from Howe et al. (1985).





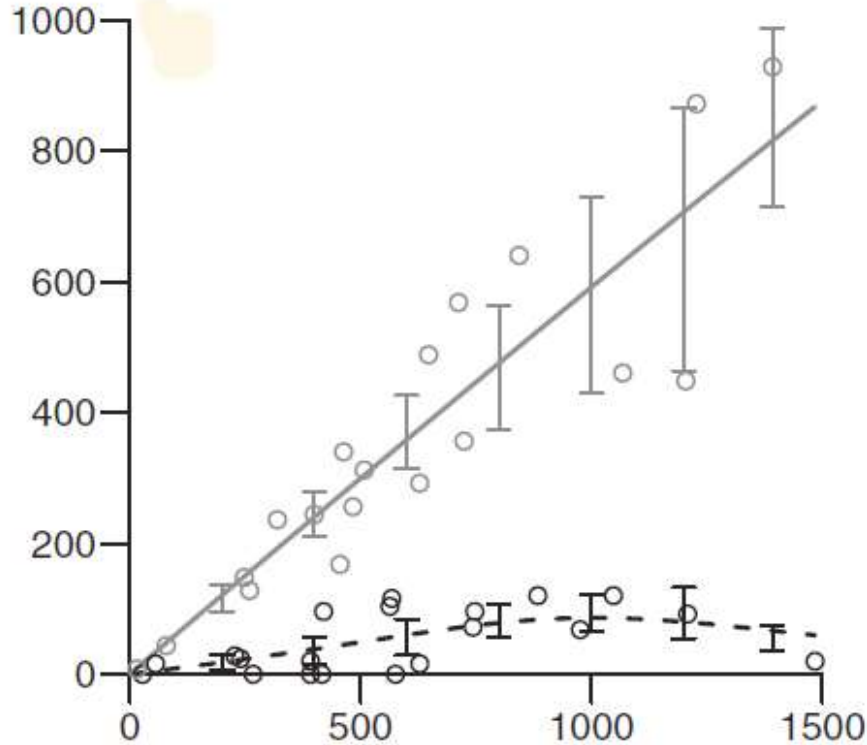
Fungal infection may have Janzen-Connell effect



Negative relationships between dN_t/dN_0 and initial density indicate NDD.

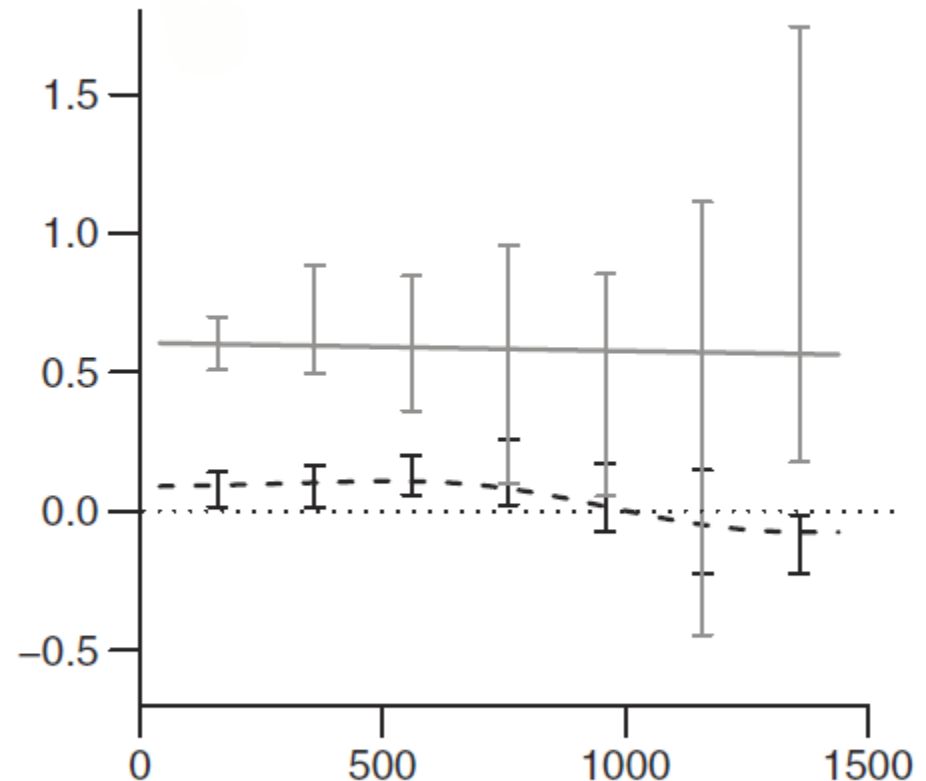
Overcompensation when $dN_t/dN_0 < 0$.

--- Control — Fungicide

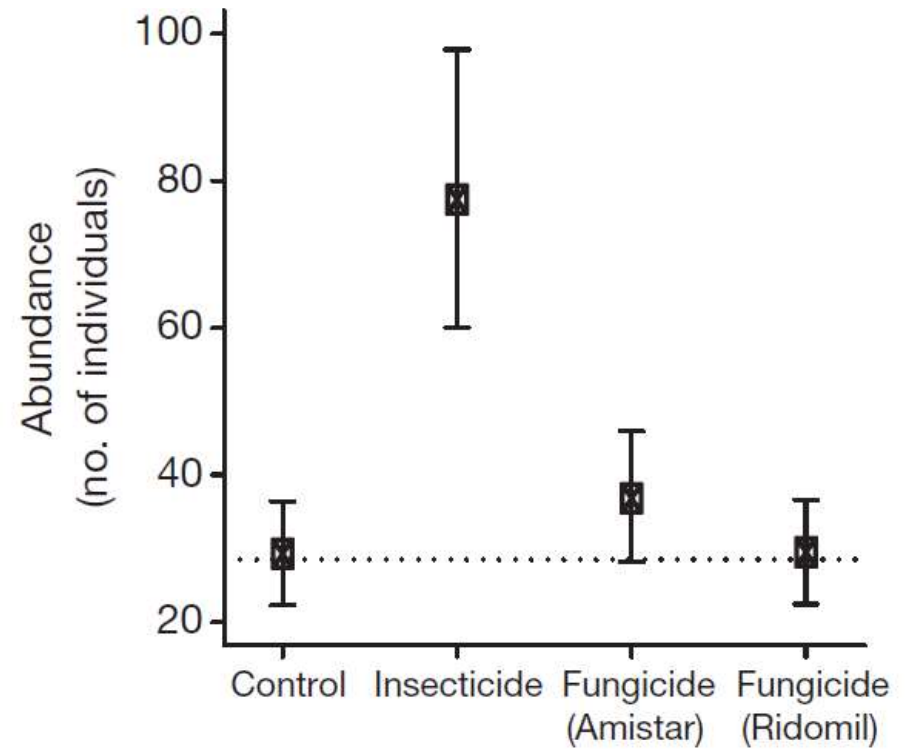
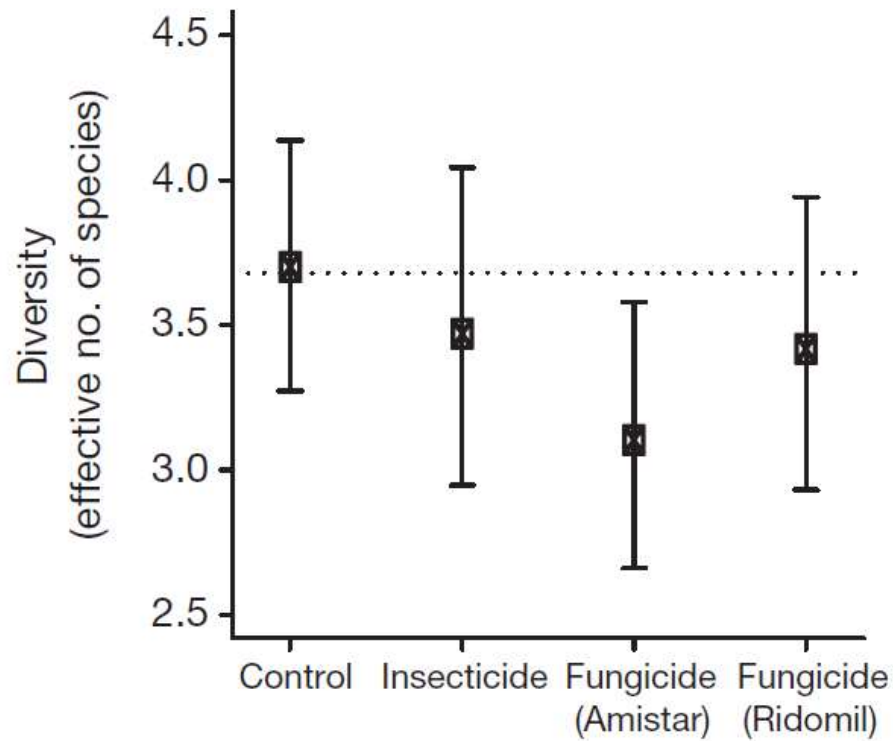


Initial density (seedlings m⁻²)

--- Control — Fungicide

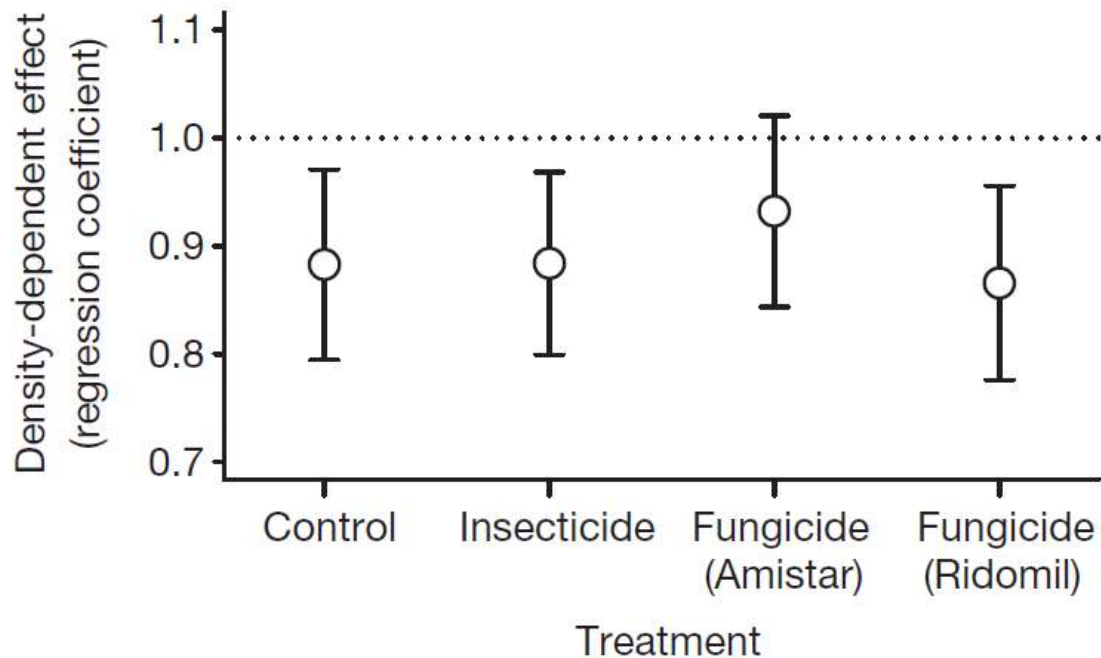
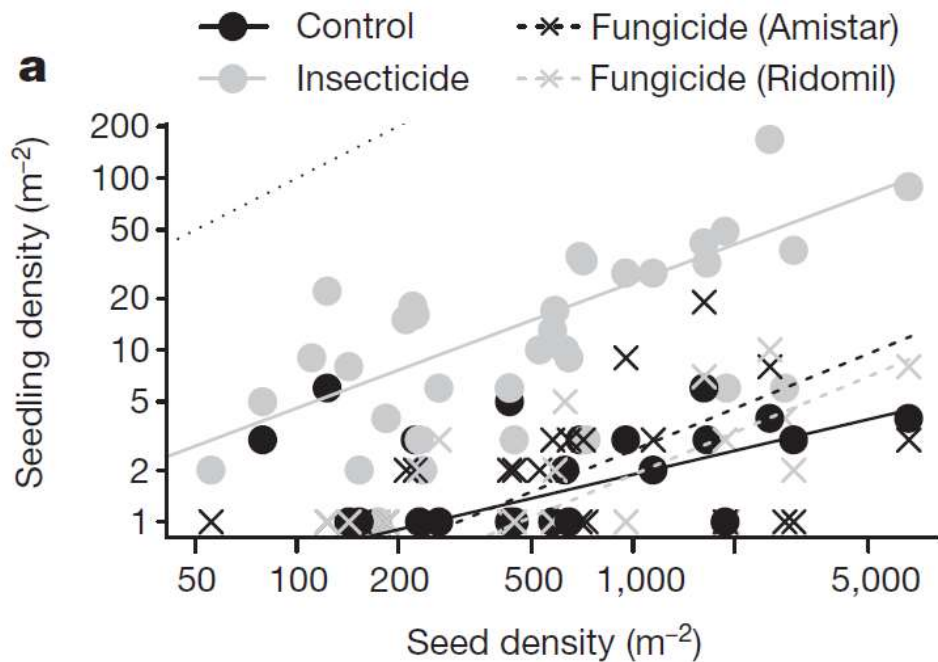


Initial density (seedlings m⁻²)



Pathogens and insect herbivores drive rainforest plant diversity and composition

Robert Bagchi^{1,2}, Rachel E. Gallery^{1,3}, Sofia Gripenberg^{1,4}, Sarah J. Gurr^{5,6}, Lakshmi Narayan¹, Claire E. Addis¹, Robert P. Freckleton⁷ & Owen T. Lewis¹



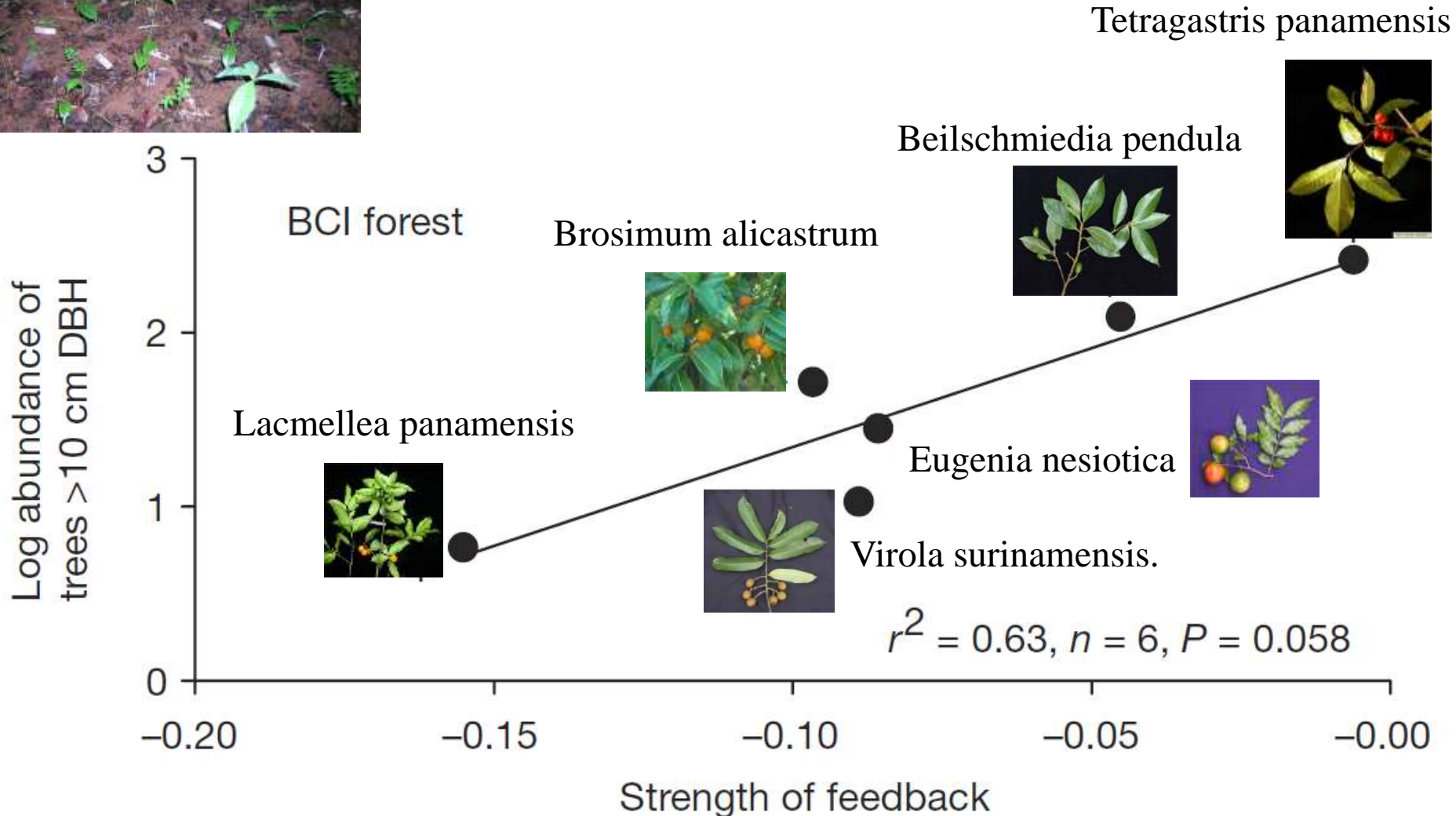
Pathogens and insect herbivores drive rainforest plant diversity and composition

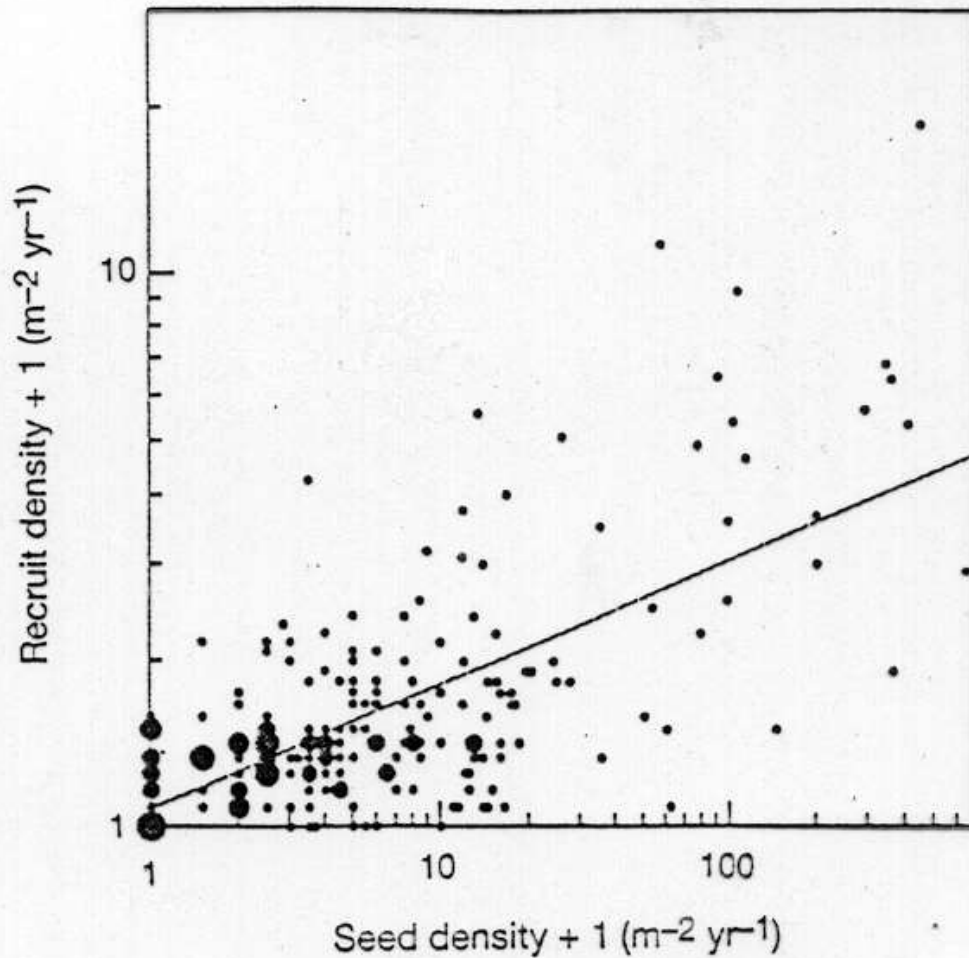
Robert Bagchi^{1,2}, Rachel E. Gallery^{1,3}, Sofia Gripenberg^{1,4}, Sarah J. Gurr^{5,6}, Lakshmi Narayan¹, Claire E. Addis¹, Robert P. Freckleton⁷ & Owen T. Lewis¹

Negative plant–soil feedback predicts tree-species relative abundance in a tropical forest



Seedling biomass after 5 months in experiments in soil from conspecific -heterospecific trees





Seed traps at BCI

slope < 1 \rightarrow
density-dependent
recruitment

Figure 1 The relationship between the seed density and recruit seedling density for *Trichilia tuberculata* (Meliaceae). The slope of the log–log relationship is less than 1, indicating that recruitment is negatively density dependent. Nonetheless, recruit density increases with seed density. Each symbol represents a census station(s) consisting of one 0.5-m² seed trap and three 1-m² seedling plots. Symbol size is proportional to the number of stations (ranging from one to six) with identical counts of seeds and recruits.

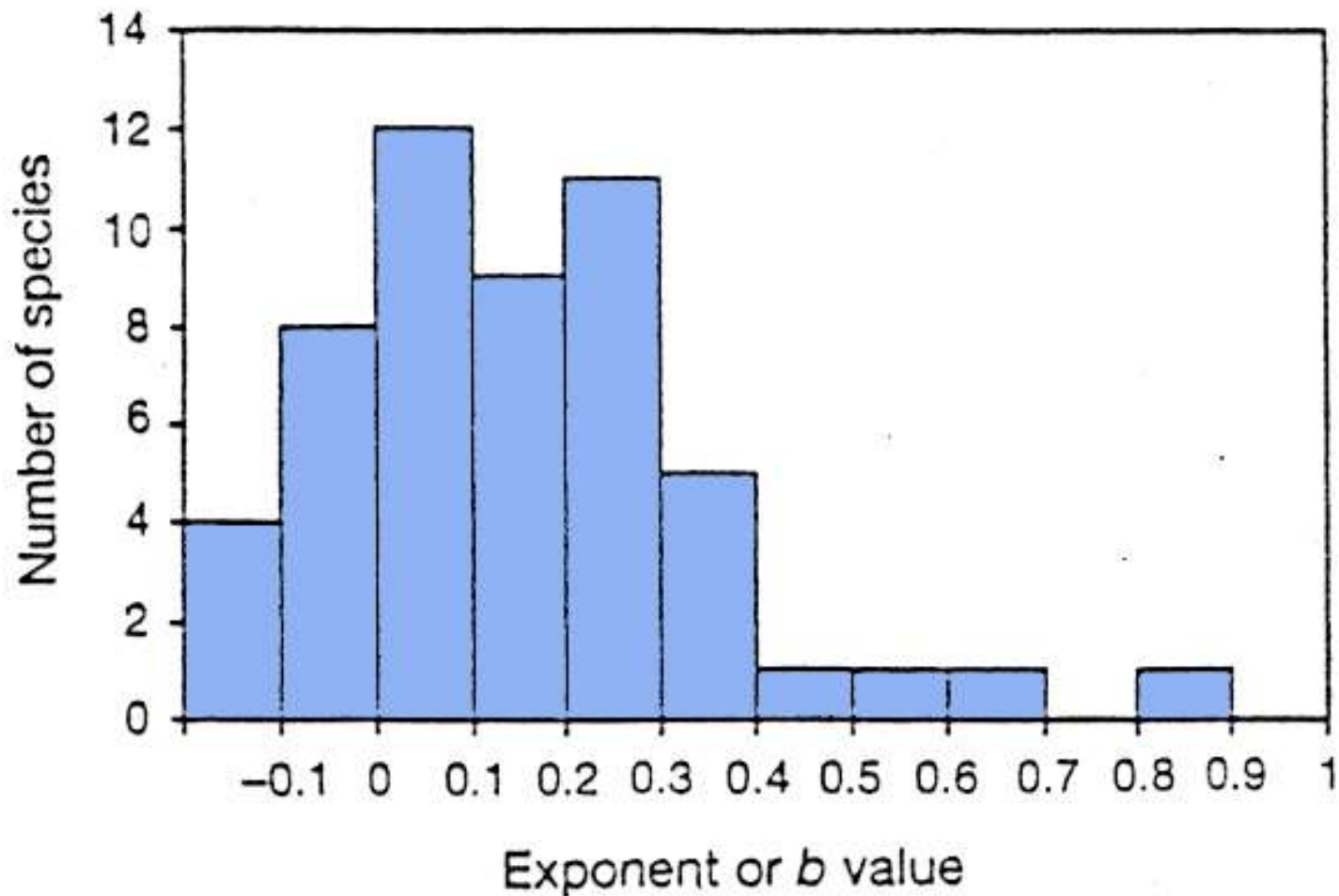


Figure 2 The frequency distribution of the exponent of the relationship between recruit density and seed density for 53 species of shrubs, trees and lianas. All exponents were less than 1 and the median was 0.23, suggesting that strong negative density dependence characterizes seedling recruitment.

Tree mortality increases species diversity

No. of species per 10 randomly selected trees:
 comparison between
 A - those which died and survived
 B - recruits and large survivors
 C - small and large survivors

In all cases the latter group was more diverse, which means the forest diversity increases over time as dying trees are less diverse than those surviving, and this trend continues as survivors grow in size

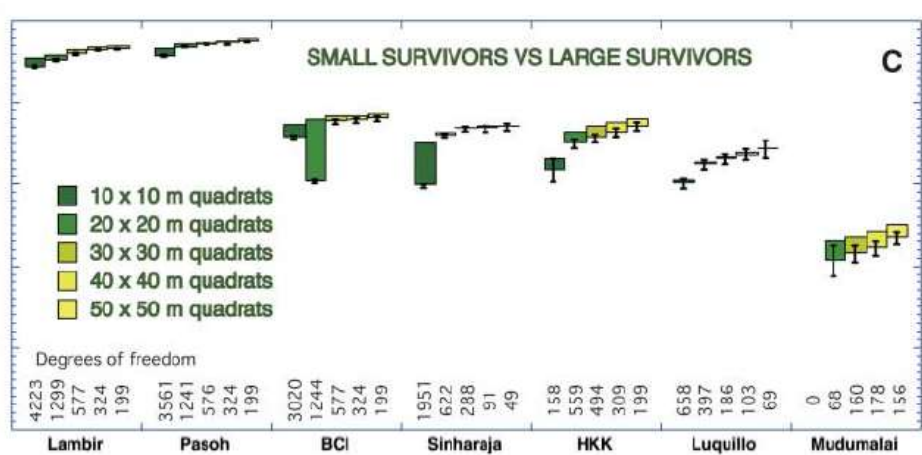
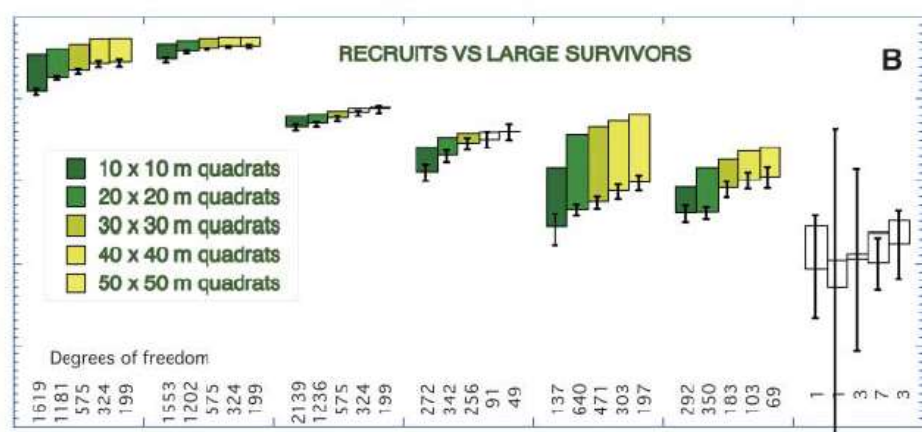
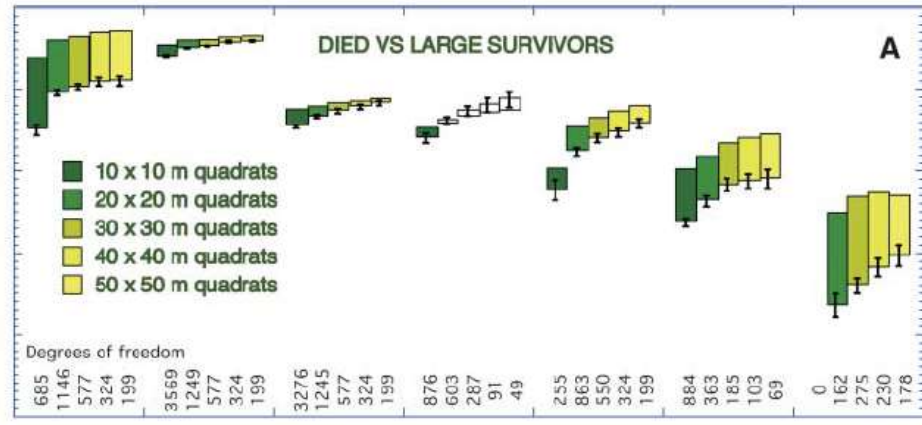


Fig. 2. Comparisons of diversity between different categories of tree, measured as the number of species per 10 randomly selected trees. The FDPs were divided into quadrats of diameter 20, 20, 20, 40, or 50 m, and comparisons were made for all quadrats in which there were at least 10 trees in each category. The differences between the mean diversities are shown as colored bars. The top of each bar indicates the mean diversity of the large survivors, and the bottom of that bar indicates the mean diversity of the category of trees that is being compared to the large survivors. Non-significant differences are shown as white bars. Error bars are the 95% confidence intervals for pairwise t-tests, and the Degrees of freedom are given below each bar. Similar results were obtained for unpaired t-tests at 2 and 5. The diversities vary among FDPs because of differences in species richness. (A) Comparisons between the diversities of trees that died during the census period and those of large survivors. (B) Comparisons between the diversities of trees that were recruited during the census period and those of large survivors. (C) Comparisons between the diversities of small survivors (the smallest quintile of each period) and large survivors (the largest three quintiles of each period).

Nonrandom Processes Maintain Diversity in Tropical Forests

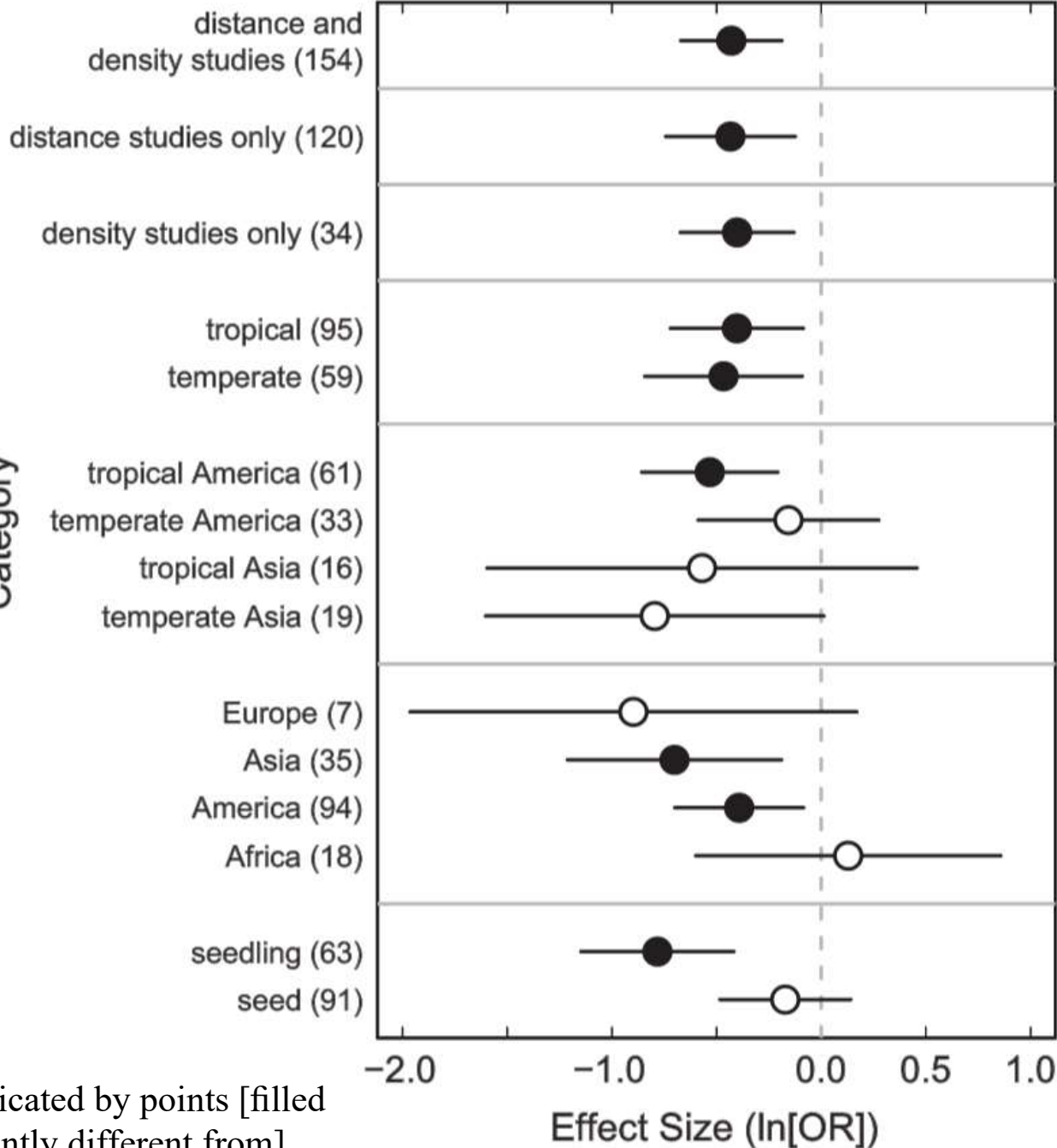
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An ecological community's species diversity leads to a steady state through time as a result of stochastic extinction, competitive exclusion, and variable food energy inputs. The erosion of diversity can be prevented over the short term if neutral or high diversity is a result of preferential recruitment of rare species or disturbance. If rare species survive preferentially, which increases diversity as the age of the individuals increases. Here, we present census data from seven Near and Old World tropical forest dynamics plots that all show the latter pattern. Within local areas, the trees that survived were in a group more diverse than those that were recruited or those that died. The larger land therefore on average older) forests were more diverse within local areas than the smaller survivors. Where species were rare in a local area, they had a higher survival rate than when they were common, resulting in enrichment for rare species and increasing diversity with age and size class in these complex ecosystems.

**Testing predictions of the Janzen–Connell hypothesis:
a meta-analysis of experimental evidence for distance-
and density-dependent seed and seedling survival**Liza S. Comita^{1,2*}, Simon A. Queenborough¹, Stephen J. Murphy¹, Jenelle L. Eck¹, Kaiyang Xu¹, Meghna Krishnadas¹, Noelle Beckman³ and Yan Zhu¹

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Category

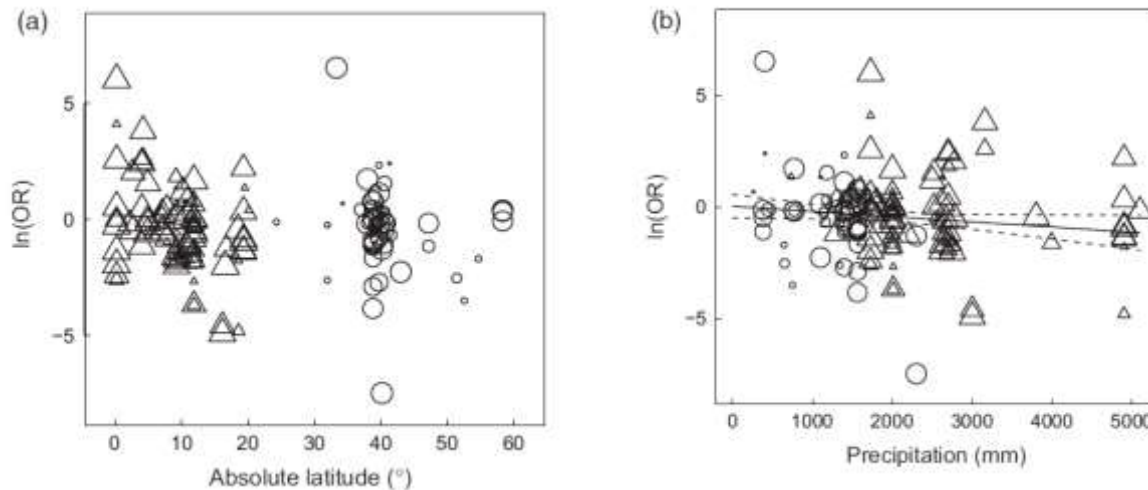


Effect sizes are indicated by points [filled circles are significantly different from 0].

significant support for both the distance- and density-dependent predictions. survival rates reduced near conspecifics compared to far from conspecifics, and in areas with high densities of conspecifics compared to areas with low conspecific densities.

unrelated to latitude, and there was no significant effect of study region. stronger distance and density dependence in wetter sites effects were significantly stronger at the seedling stage compared to the seed stage.

natural enemies are frequently the cause of such patterns, consistent with the Janzen–Connell hypothesis, but additional studies are needed to rule out other mechanisms (e.g. intraspecific competition).



The effect of (a) absolute latitude, (b) precipitation and (c) study duration on the log odds ratio of survival of seeds and seedlings in a meta-analysis of experimental tests of density and distance dependence. The size of each point is proportional to the precision of the estimate. In panels (a) and (b), tropical (circles) and temperate (triangles) studies are indicated.

So, what are the most promising theories to explain high species richness in tropical forests?

- niche differences
- infrequent competition in rare species
- Janzen-Connell effects, particularly for pathogens
- negative density dependence for abundant species
- mass effects and dispersal limitation
- disturbance

But no single effect is strong enough to provide overall explanation



Why Do Some Tropical Forests Have So Many Species of Trees?'

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ABSTRACT

Understanding why there are so many kinds of tropical trees requires learning, not only how tree species coexist, but what factors drive tree speciation and what governs a tree clade's diversification rate. Many report that hybrid sterility evolves very slowly between separated tree populations. If so, tree species rarely originate by splitting of large populations. Instead, they begin with few trees. The few studies available suggest that reproductive isolation between populations usually results from selection driven by lowered fitness of hybrids: speciation is usually a response to a "niche opportunity." Using Hubbell's neutral theory of forest dynamics as a null hypothesis, we show that if new tree species begin as small populations, species that are now common must have spread more quickly than chance allows. Therefore, most tree species have some setting in which they can increase when rare. Trees face trade-offs in suitability for different microhabitats, different-sized clearings, different soils and climates, and resistance to different pests. These trade-offs underlie the mechanisms maintaining α -diversity and species turnover. Disturbance and microhabitat specialization appear insufficient to maintain α -diversity of tropical trees, although they may maintain tree diversity north of Mexico or in northern Europe. Many studies show that where trees grow readily, tree diversity is higher and temperature and rainfall are less seasonal. The few data available suggest that pest pressure is higher, maintaining higher tree diversity, where winter is absent. Tree α -diversity is also higher in regions with more tree species, which tend to be larger, free for a longer time from major shifts of climate, or in the tropics, where there are more opportunities for local coexistence.

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S. Joseph Wright

Plant diversity in tropical forests: a review of mechanisms of species coexistence

SPECIAL FEATURE – STANDARD PAPER

META-ANALYSIS IN PLANT ECOLOGY

Testing predictions of the Janzen–Connell hypothesis: a meta-analysis of experimental evidence for distance- and density-dependent seed and seedling survival

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Summary

1. The Janzen–Connell hypothesis proposes that specialist natural enemies, such as herbivores and pathogens, maintain diversity in plant communities by reducing survival rates of conspecific seeds and seedlings located close to reproductive adults or in areas of high conspecific density. Variation in the strength of distance- and density-dependent effects is hypothesized to explain variation in plant species richness along climatic gradients, with effects predicted to be stronger in the tropics than the temperate zone and in wetter habitats compared to drier habitats.

2. We conducted a comprehensive literature search to identify peer-reviewed experimental studies published in the 40+ years since the hypothesis was first proposed. Using data from these studies, we conducted a meta-analysis to assess the current weight of evidence for the distance and density predictions of the Janzen–Connell hypothesis.

3. Overall, we found significant support for both the distance- and density-dependent predictions. For all studies combined, survival rates were significantly reduced near conspecifics compared to far from conspecifics, and in areas with high densities of conspecifics compared to areas with low conspecific densities. There was no indication that these results were due to publication bias.

4. The strength of distance and density effects varied widely among studies. Contrary to expectations, this variation was unrelated to latitude, and there was no significant effect of study region. However, we did find a trend for stronger distance and density dependence in wetter sites compared to sites with lower annual precipitation. In addition, effects were significantly stronger at the seedling stage compared to the seed stage.

5. *Synthesis.* Our study provides support for the idea that distance- and density-dependent mortality occurs in plant communities world-wide. Available evidence suggests that natural enemies are frequently the cause of such patterns, consistent with the Janzen–Connell hypothesis, but additional studies are needed to rule out other mechanisms (e.g. intraspecific competition). With the widespread existence of density and distance dependence clearly established, future research should focus on assessing the degree to which these effects permit species coexistence and contribute to the maintenance of diversity in plant communities.

Abstract Evidence concerning mechanisms hypothesized to explain species coexistence in hyper-diverse communities is reviewed for tropical forest plants. Three hypotheses receive strong support. Niche differences are evident from non-random spatial distributions along micro-topographic gradients and from a survivorship-growth tradeoff during regeneration. Host-specific pests reduce recruitment near reproductive adults (the Janzen–Connell effect), and, negative density dependence occurs over larger spatial scales among the more abundant species and may regulate their populations. A fourth hypothesis, that suppressed understory plants rarely come into competition with one another, has not been considered before and has profound implications for species coexistence. These hypotheses are mutually compatible. Infrequent competition among suppressed understory plants, niche differences, and Janzen–Connell effects may facilitate the coexistence of the many rare plant species found in tropical forests while negative density dependence regulates the few most successful and abundant species.

Recommended additional
reading for interested students